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INVITED REVIEW

Oxygen dynamics of marine sediments

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Abstract

Benthic O₂ availability regulates many important biogeochemical processes and has crucial implications for the biology and ecology of benthic communities. Further, the benthic O₂ exchange rate represents the most widely used proxy for quantifying mineralization and primary production of marine sediments. Consequently, numerous researchers have investigated the benthic O₂ dynamics in a wide range of environments. On the basis of case studies – from abyssal sediments to microbial phototrophic communities – I hereby try to review the current status on what we know about controls that interrelate with the O₂ dynamics of marine sediments. This includes factors like: sedimentation rates, bottom water O₂ concentrations, diffusive boundary layers, fauna activity, light, temperature, and sediment permeability. The investigation of benthic O₂ dynamics represents a challenge in resolving variations on temporal and spatial scales covering several orders of magnitude. Such an effort requires the use of several complementary measuring techniques and modeling approaches. Recent technical developments (improved chamber approaches, O₂ optodes, eddy-correlation, benthic observatories) and advances in diagenetic modeling have facilitated our abilities to resolve and interpret benthic O₂ dynamics. However, all approaches have limitations and caveats that must be carefully evaluated during data interpretation. Much has been learned during the last decades but there are still many unanswered questions that need to be addressed in order to fully understand benthic O₂ dynamics and the role of sediments for marine carbon cycling.

Key words: *Benthic mineralization, benthic photosynthesis, biogeochemistry, in situ, lander technology, microsensors, planar optodes*

Introduction

Oxygen is a prerequisite for higher fauna and flora and is a central molecule for global element cycling. Oxygen is produced during photosynthesis, and consumed directly or indirectly during the degradation of organic material. Thus, O₂ represents an excellent tracer for biological activity. The availability and turnover of O₂ is an absolute key measure when evaluating the biological status of the environment and the cycling of important elements such as carbon, nitrogen, phosphorus, sulfur and metals – locally as well as globally. The distribution and availability of O₂ in the oceans has thus received considerable scientific attention for many years.

Marine ecosystems basically consist of two very different compartments; the water column and the sediment matrix. The global volume of the pelagic compartment with an average water depth of 3800 m by far exceeds the size of the upper decimeters of the

benthic compartment, which host most of the biological activity. However, the volume-specific production and degradation of organic material in surface sediments, hosting high densities of microbes and metazoans, are typically 100–1000 times higher than the corresponding values for the water column. Especially in coastal, light-exposed sediments the microbial activity (photosynthesis and respiration) of the upper millimeters or centimeters of the sediment can be quantitatively more important than the integrated activity of the entire water column. In shelf areas, it is estimated that 15–50% of the pelagic primary production reaches the sediment (e.g. Canfield 1993; Wollast 1998) and here benthic mineralization plays an important role for the recycling of nutrients and carbon. On a global scale, the balance between burial and mineralization of the organic carbon reaching the sediment has important implications for the O₂ and CO₂ concentration of the

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biosphere and, on geological time-scales, profound impact on climatic conditions and life as we know it today (Berner 1980, 1987; Berner & Canfield 1989; Archer & Meier-Reimer 1994). It is thus highly relevant to quantify the benthic turnover rates of organic material and to investigate the controls regulating the involved process rates – O_2 availability is one important key aspect of this highly complex task (e.g. Canfield 1994; Hulthe et al. 1998; Kristensen & Holmer 2001).

In areas where the seafloor is situated below the depth of the photic zone, organic material is mainly supplied from the overlying water column in the form of agglutinated phytoplankton aggregates, fecal pellets and carcasses. Some of the material is digested by metazoans, but generally microbial-mediated carbon degradation is quantitatively more important for the turnover of the settling material (e.g. Glud et al. 2003). Prokaryotes cannot assimilate molecules larger than 500–600 Dalton (Weiss et al. 1991) and thus benthic microbial degradation is initiated by extracellular depolymerization of exoenzymes released to the interstice or attached to the cell surface. Smaller organic molecules can be completely mineralized to CO_2 by a wide range of aerobic heterotrophic bacteria. However, anaerobic heterotrophic bacteria generally use a narrower spectrum of substrates and thus rely on fermentative transformation products of other microbes providing a more uniform pool of low molecular organic substances for their metabolism (Canfield et al. 2005).

Thermodynamically, O_2 represents the most favorable abundant electron acceptor available, but seawater contains little O_2 , and this oxidant is soon depleted and generally only extends a few millimeters or centimeters into the sediment. In many environments, benthic carbon degradation is therefore mainly mediated anaerobically by microbes using nitrate, manganese oxides, iron oxides or sulfate as electron acceptors. In the absence of these oxidants, organic carbon can also be degraded via methanogenesis, but the process does not represent a net oxidation of organic carbon (Fenchel et al. 1998). Typically, sediments reflect a vertical redox zonation where the electron acceptors overall are depleted in the order outlined above, reflecting the potential energetic gain related to the respective redox processes (Froelich et al. 1979). However, due to faunal activity most marine sediments are heterogeneous and vertical profiles of area-integrated process rates often reflect extensive overlapping zonation of the different heterotrophic pathways (Canfield et al. 1993; Kostka et al. 1999). The relative importance of the respective oxidation pathways depend on the sedimentation and bioturbation rate, the sediment and the water

chemistry, but each have been found to dominate in different environments (Thamdrup & Canfield 2000 and references therein). The reduced products from anaerobic degradation are, to a large extent, reoxidized by O_2 . This occurs either directly or *via* a redox-cascade involving a series of complex abiotic and microbial catalysed redox processes (Fenchel & Jørgensen 1977; Jørgensen 2000). Benthic O_2 consumption is thus used (1) for aerobic heterotrophic activity of fauna and bacteria, and (2) for the reoxidation of reduced inorganic products released during the anaerobic heterotrophic degradation. There is no direct way to quantify the relative importance of the two O_2 -consuming processes, but the two extremes are represented by microbial mats of the chemolithotrophic bacteria *Beggiatoa* sp. where >90% of the O_2 is used for sulfide oxidation (Jørgensen 1982) and the abyss where benthic mineralization of the very low input of organic material can be almost completely covered by the available O_2 (Bender & Heggie 1984).

To the extent reduced inorganic solutes from anaerobic mineralization are fully oxidized within sediments, the total benthic O_2 uptake (TOU) represents a proxy of the total benthic carbon mineralization (Canfield et al. 1993) (Figure 1). The approach ignores minor electron sinks like N_2 release from denitrification and burial of sulfide (mainly in the form of pyrite); however, these two processes rarely represent more than 15% of the electron equivalents of the total carbon mineralization (Jørgensen 1982; Canfield et al. 2005). The approach also does not account for the O_2 equivalents used for nitrification which at maximum can be ~20% of the O_2 consumption during fully aerobic oxidation of typical sedimentary organic material with a Redfield C:N ratio of 8–12 (Anderson & Sarmiento 1994). Conveniently, the two minor components not accounted for when using the benthic O_2 uptake as a proxy for the total carbon degradation counteract each other and are often assumed to cancel each other out. They can, however, also be accounted for by including the processes in the stoichiometry used to convert the O_2 exchange rates to the CO_2 exchange (see below). Alternatively, the nitrification, denitrification and pyrite burial can be determined independently for a given setting if a more accurate determination of the total carbon mineralization is required.

A more serious problem of transforming TOU into total carbon mineralization rates is the assumption that the production and the oxidation of reduced solutes from the anaerobic degradation are at steady state. In many environments, reduced equivalents from anaerobic mineralization accumulate on a seasonal (or diel) basis, and the ' O_2 debt' is

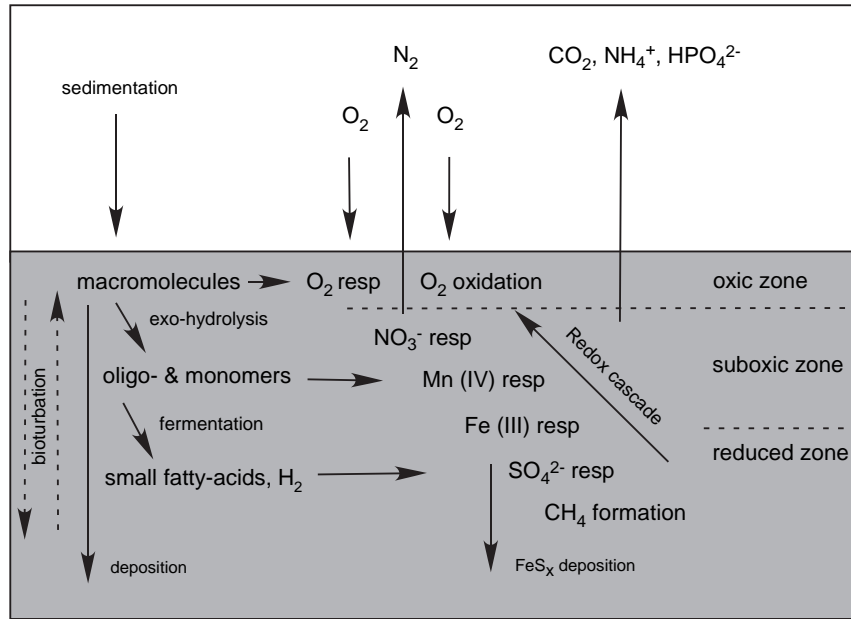


Figure 1. A schematic presentation of some important diagenetic processes in marine sediments (inspired by Fenchel & Jørgensen 1977).

repaid during less productive periods or resuspension events. This has to be kept in mind when a given O_2 exchange rate is recalculated into a total carbon oxidation rate. Thus, it is advantageous to measure the release of dissolved inorganic carbon (DIC) along with the O_2 uptake rate, as DIC represents the ultimate product of organic carbon oxidation (Anderson et al. 1986; Hulth et al. 1997). The ratio between simultaneously determined DIC and O_2 exchange rates provides a tool for evaluating to what extent production and oxidation of reduced species from anaerobic mineralization are at steady state and balanced at the time of investigation, and the ratio can vary extensively on a seasonal basis especially in shallow eutrophicated environments (Therkildsen & Lomstein 1993). In essence the DIC exchange rate provides a more correct estimate of the concurrent mineralization rate and thus reflects a high seasonal variation whereas TOU integrates the mineralization activity over a longer – not always well defined – time span and thereby exhibits lower temporal dynamic (e.g. Therikildsen & Lomstein 1993). Integrated annually the exchange rates converge and the DIC/ O_2 exchange ratio approaches the average oxidation state of the carbon being mineralized. When converting the O_2 exchange rates into carbon oxidation rates the DIC/ O_2 exchange ratio is typically assumed to range between 0.8 and 1.2 (e.g. Smith 1989; Therikildsen & Lomstein 1993; Hammond et al. 1996; Roden & Wetzel 1996). DIC exchange determinations are generally less accurate as they are based on discrete sampling – and in environments of low diagenetic activity the analytic precision may not always match

the small DIC exchange rates. DIC analysis on recovered water samples collected during chamber incubations are time consuming, and the inorganic carbon dynamics can be confounded by dissolution or precipitation of carbonates (e.g. Green et al. 1993). However, at reactive sites the carbonate dissolution–precipitation dynamic can be accounted for by concurrent determination of the Ca^{2+} or alkalinity exchange rate (Anderson et al. 1986; Ståhl et al. 2004c). There are pros and cons of using both the O_2 and the DIC exchange rates for assessing the benthic mineralization rate, and if possible combined determinations are to be preferred. However, TOU determinations remain the most widely used robust approach for assessing benthic carbon mineralization.

Measurements of benthic oxygen uptake rates

Benthic O_2 exchange rates are most commonly determined by enclosure techniques (Figure 2, upper panel). Here the initial O_2 decrease rate of an overlying well-mixed water phase is approximately linear, and the TOU is calculated, accounting for the enclosed area and the water volume. Even though the initial decrease in O_2 concentration is inseparable from a linear decline it is in fact non-linear as the O_2 availability in the sediment declines with the concentration of the overlying water (Bender et al. 1989). This becomes especially apparent in elongated core or chamber incubations where the overlying water is approaching anoxia. Assuming depth-independent O_2 consumption rates and that O_2 in essence was consumed by aerobic

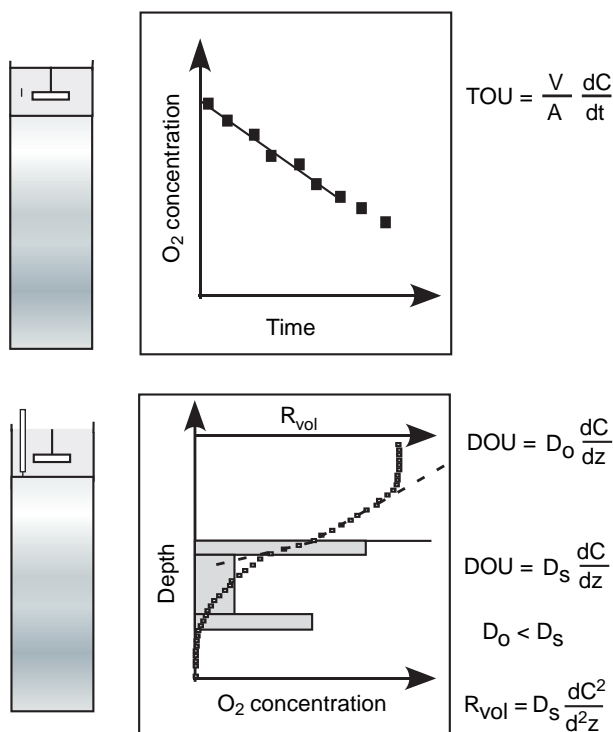


Figure 2. The two most common procedures for quantifying benthic O_2 exchange rates. The upper panel shows an enclosed sediment core and the discrete recordings of an O_2 sensor inserted into the well-mixed overlying water phase. The TOU is calculated from the area of the enclosed sediment (A) and the volume of the enclosed water (V) and the rate by which O_2 is consumed. The lower panel depicts microscale O_2 measurements across the sediment–water interface (horizontal black line) while the O_2 concentration of the overlying water is kept constant. From such profiles the DOU can be calculated from the concentration gradient either within the DBL or just below the sediment surface applying the relevant transport coefficient, the molecular diffusion coefficient (D_o) or the lower, tortuosity-corrected transport coefficient in the sediment (D_s). The volume-specific O_2 consumption rate (R_{vol}) depicted by the grey boxes can be derived from Fick's second law of diffusion, i.e. the curvature of the profiles. Typically such calculations reflect intensified activity at the top and bottom of the O_2 microprofiles due to organic enrichment and high reoxidation rate in the two zones, respectively.

heterotrophic activity, Hall et al. (1989) suggested applying exponential curve fitting to access the initial O_2 decline rate. However, in most sediments a significant fraction of the O_2 consumption is related to reoxidation processes and in such instances more sophisticated biogeochemical modeling is required to fully evaluate the benthic O_2 dynamic (Jørgensen and Boudreau 2001; Berg et al. 2003b). Elongated chamber incubations include conditions within the surface sediment where the O_2 consumption dynamics poorly represent the initial conditions and shifts in fauna activity related to the decline in O_2 concentration may further complicate matters. Thus it is '*a priori*' difficult to predict the most appropriate function describing the elongated O_2 decline of a

sealed sediment enclosure as it is the net result of a complex, transient shift in mobile reactions within the sediment (see below) combined with variations in fauna activity. By far most studies apply a linear approximation to the initial O_2 decline (with R^2 values well above 0.95) and in reality this is fully applicable if the O_2 concentration does not sink below 10–15% of the initial starting point. The enclosure approach quantifies the total O_2 exchange rate across the benthic interfaces including O_2 consumption related to fauna activity (Rasmussen & Jørgensen 1992).

For impermeable sediments the diffusion-mediated benthic O_2 uptake can be calculated from O_2 microprofiles by different approaches. The most common procedure calculates the diffusive O_2 uptake rate (DOU) from a linear approximation to the O_2 concentration gradient resolved within the diffusive boundary layer (DBL – see below) or just below the sediment–water interface applying Fick's first law of diffusion. More recently a non-linear approximation to the DBL gradient was suggested to accommodate apparent higher-order hydrodynamic processes in the DBL (Nishihara & Ackerman 2007). Alternatively, the volume-specific O_2 consumption (R_{vol}) can be determined from the convex-upwards-shaped porewater profile using Fick's second law of diffusion, and the DOU can be calculated accounting for the O_2 penetration depth (or in case R_{vol} turns out to be depth-dependent accounting for the thickness of the respective depth interval) (Figure 2, lower panel). Several procedures facilitating the latter approach exist (Nielsen et al. 1990; Berg et al. 1998), but in contrast to calculations based on the concentration gradient of the DBL, sediment-based calculations all rely on an independently determined solute transport coefficient of the sediment (Figure 2). Therefore, the DBL approach is preferred for simple exchange measurements, while the second approach may provide insight in depth variations of R_{vol} . TOU should in principle always be equal to or larger than the DOU as $TOU = DOU + BMU$, where BMU represent any benthos mediated O_2 uptake, which includes fauna respiration and irrigation.

Clark type microelectrodes are excellent tools for measuring O_2 microprofiles or to follow O_2 changes in enclosed water volumes (Revsbech & Jørgensen 1986). However, optical O_2 determination, taking advantage of the ability of O_2 to act as a dynamic quencher of the luminescence from certain indicator dyes, represent an interesting alternative to polarographic O_2 measurements (Klimant et al. 1995). The so-called micro-optodes simply consist of immobilized O_2 -sensitive luminescent chemistry fixed to the tip of a tapered fiber. Excitation light is guided

towards the sensor tip via the fiber and a fraction of the O₂-sensitive luminescent signal is guided in the opposite direction towards the measuring instrument (Klimant et al. 1995). Optodes offer an interesting alternative due to simpler manufacturing and superior long-term stability especially if combined with life-time base sensing schemes (Klimant et al. 1995, 1997b; Holst et al. 1997; Tengberg et al. 2006). The introduction of micro-optodes to aquatic biology inspired the development of planar optodes, where the O₂ quenchable luminescent chemistry is immobilized on transparent support foils that can be placed along e.g. an aquarium wall. Excitation light is supplied from the outside and by using a digital camera the O₂-sensitive luminescence is imaged and converted into O₂ images (Glud et al. 1996b). Transparent optodes can be applied and this greatly facilitates alignment between patterns in the O₂ distribution and structures of the sediment (Holst & Grünwald 2001; Frederiksen & Glud 2006). Different sensing chemistries and imaging set-ups have been developed to meet the requirements of various specific experimental conditions (e.g. Holst et al. 1998; Precht et al. 2004; König et al. 2005; Oguri et al. 2006; Schröder et al. 2007).

Total O₂ exchange rates and O₂ microprofiles can be measured on recovered sediment cores in laboratory set-ups that maintain *in situ* temperature, *in situ* bottom water O₂ concentration and a well-mixed overlying water phase (Rasmussen & Jørgensen 1992) – this is of utmost importance to obtain trustworthy results. Alternatively, the measurements can be obtained *in situ* by diver-operated modules or for deeper water by so-called autonomous benthic lander systems that carry benthic chambers and profiling units to the seafloor (Smith 1978; Reimers 1987) (Figure 3). In recent years, optodes have been adapted to lander modules (Glud et al. 1999b; Wenzhöfer et al. 2001b; Tengberg et al. 2006) and inverted camera periscopes (Rhoads & Germano 1982) have been used to perform *in situ* planar optode measurements (Glud et al. 2001, 2005). For a review of *in situ* O₂ measurements see Glud et al. (2000a).

Initial comparison between O₂ microprofiles measured shipboard and *in situ* at water depths >1000 m indicated that core recovery introduced artifacts that affected the benthic O₂ distribution (Reimers et al. 1986; Reimers 1987). The differences were mainly explained by disturbance of the sediment structure during core recovery and improper establishment of *in situ* conditions during the onboard measurements. More detailed inter-comparisons clearly revealed that shipboard O₂ microprofile measurements overestimated the DOU and underestimated the O₂ penetration depth at water depth

>600 m (Figure 4). The maximum difference was observed at 5000 m water depth where the DOU measured in the laboratory was 3.5 times higher than the *in situ* value while the O₂ penetration depth in the recovered sediment cores was only 20% of the one measured *in situ* (Glud et al. 1994a; Epping et al. 2002). Furthermore, it was observed that after re-establishing *in situ* conditions in the laboratory (i.e. temperature and bottom water O₂ concentration) the O₂ microprofiles gradually reshaped into the *in situ* microprofiles (Glud et al. 1999a). The observed effect was ascribed to transient heating during core recovery probably combined with lysis of psychrophilic or barophilic organisms, which overall stimulated the microbial activity in the recovered sediment (Glud et al. 1994a, 1999a; Aller et al. 1998; Sauter et al. 2001).

The observed impact generally increased with water depth as it gradually takes longer times for diffusion to re-establish the *in situ* profiles with increasing O₂ penetration depth (Figure 4) – and O₂ penetration increases with the water depth (see below). Attempts to apply general empirical relations correcting the large database of laboratory-based measurements for recovery artifacts have been presented (Sachs et al. forthcoming). However, such approaches are questionable as no strict relationship between *in situ* and laboratory data can be expected due to differences in the local environmental conditions and core handling procedures. Recently, it was suggested that supersaturating N₂ and CO₂ levels in cells could induce cell lysis when cores were recovered from great depth to surface pressure and temperature conditions with lower gas solubility (Hall et al. 2007). For deep-sea environments trustworthy O₂ microprofiles can only be obtained by *in situ* procedures – or by applying extremely long pre-incubation periods which in itself is a problem. Evidence that the same applies to other diagenetically important solutes are accumulating (Hammond et al. 2004; Hall et al. 2007).

In coastal water with a few millimeters O₂ penetration depth, low hydrostatic pressure and generally a smaller temperature difference between top and bottom water, *in situ* and laboratory-based O₂ microprofile measurements align much better (e.g. Lansard et al. 2003). However, extensive comparisons indicate some problems in fully re-establishing *in situ* profiles in the laboratory – especially in diagenetically very active sediments (Glud et al. 2003). The reasons remain unclear but might be related to irreversible disturbance during core recovery or poor re-establishment of the factors regulating the aerobic activity *in situ*. Nevertheless, the observation underlines the importance of performing microprofile measurements *in situ* (or

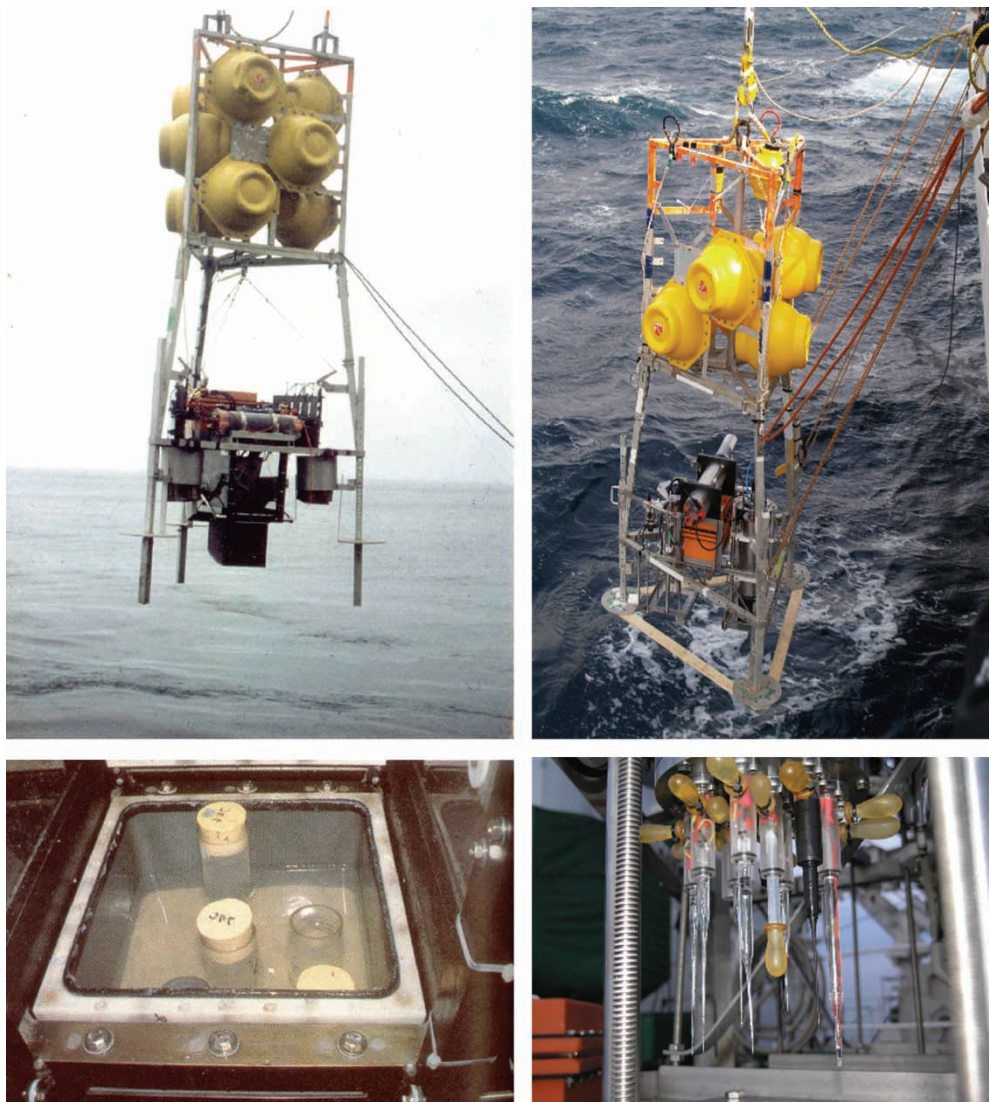


Figure 3. The top, left panel shows the deployment of a benthic lander carrying a chamber for *in situ* incubations, while the top, right panel shows a lander equipped with a microprofiling and planar optode module. The bottom, left panel shows the chamber after recovery from 3000 m water depth. Subsamples of the recovered diatom ooze are being extracted. The bottom, right panel shows an *in situ* microsensor array of microelectrodes for measuring O_2 , pH, H_2S and resistivity (the central pipette ball covers a reference electrode for the pH measurements).

under *in situ* conditions) in order to quantitatively address benthic O_2 dynamics also in coastal waters.

Total O_2 uptake measurements are affected by the same problems as outlined for the microprofile measurements. However, the TOU also includes O_2 consumption related to faunal activity and thus, correct representation of an active and viable fauna is essential for high-quality measurements. Sediment core samplers, typically used for laboratory-based exchange measurements, recover a relatively small sediment area (70–180 cm²), and they might not include undisturbed, active fauna of the correct density. Laboratory-based incubations of intact sediment cores from coastal environments consistently underestimate the fluxes from *in situ* incubations of larger benthic chambers, often by a factor of ~2–3

(Archer & Devol 1992; Glud et al. 1998a, 2003). For correct faunal representation (and TOU rates) it is highly recommended to use relatively large core samplers or benthic chambers – the problem of using small enclosures increases with increasing natural heterogeneity and the average size of macrofauna specimens contributing to the benthic exchange rate. This has been further validated and quantitatively discussed on the basis of computer simulation of sampling artifacts associated with chamber incubations on virtual computer-designed fauna-inhabited sediments (Glud & Blackburn 2002). In each given case the advantage of using larger core samplers and chambers has, however, to be balanced against the simpler logistic and fewer complications of applying smaller tubes. Artifacts associated with changes in

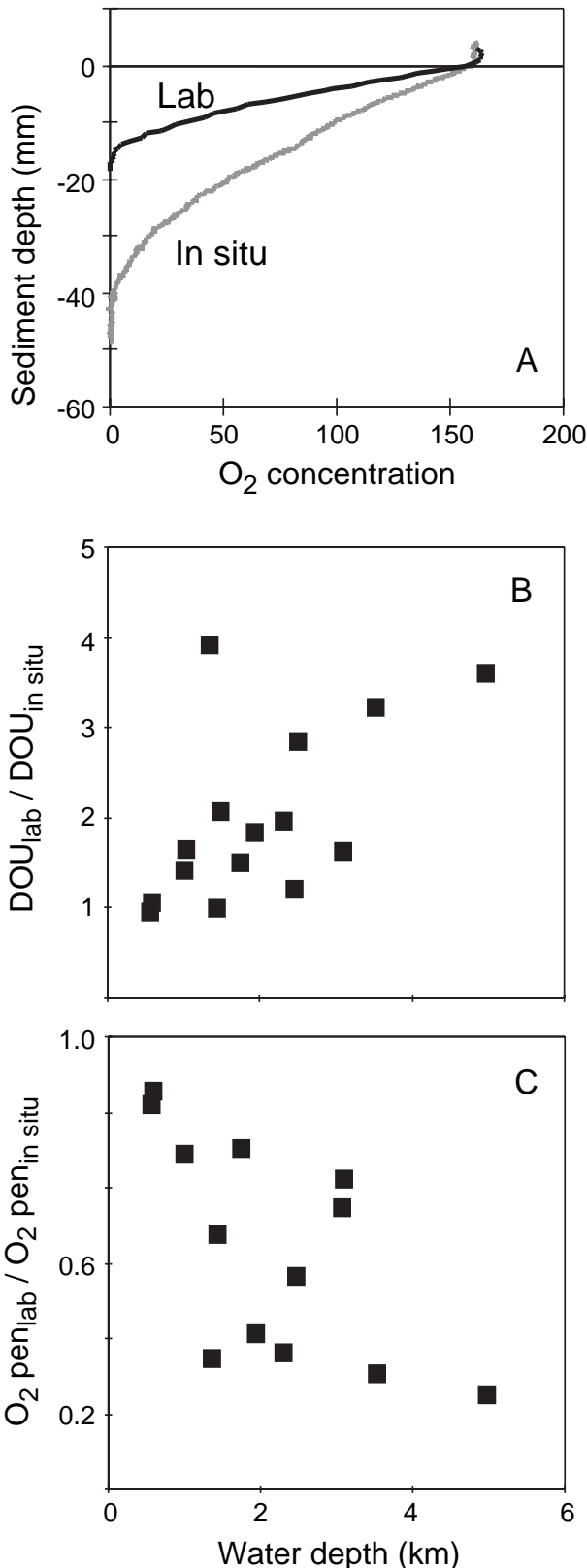


Figure 4. (A) O₂ microprofiles measured *in situ* and in a recovered sediment core from 3500 m depth. (B, C) The ratio between *in situ* and laboratory-obtained data versus station depth as compiled from three different studies (Glud et al. 1994a, 1999a; Wenzhöfer & Glud 2002).

faunal behavior during enclosure will presumably be less when incubations are performed *in situ* rather than in the laboratory where the benthic community has experienced a stressful recovery process.

Recently, 'eddy-correlation' was introduced as a new approach to quantify *in situ* benthic O₂ exchange rates (Berg et al. 2003a; Kuwae et al. 2006). The technique is based on instantaneous flux determinations calculated from simultaneously measured fluctuations in the vertical flow component and the solute concentration in a given point at some distance from the sediment surface. The principle has for many years been used for air-sea or air-soil exchange determination of e.g. CO₂ and CH₄ (e.g. Wyngaard 1990). The approach is truly non-invasive and integrates large sediment areas – up to several m² depending on the sediment roughness and the height above the seafloor where measurements are performed (Berg et al. 2007). The integration of larger sediment areas seems to result in larger exchange rates by including events or benthic communities poorly represented by traditional chamber approaches (Berg et al. 2003a). The measuring principle is superior to any other approach for determining the total benthic O₂ exchange rates and the approach bears great potential especially in highly heterogeneous environments. However, the full potential still needs documentation and the sensitivity of the approach is still unclear.

There have been a number of recent reviews discussing preferred designs and technical solutions for *in situ* benthic exchange measurements (Tengberg et al. 1995, 2004, 2005; Viollier et al. 2003) and it is beyond the scope of the present paper to enter that discussion. However, even though consensus is that *in situ* technology is preferable for quantifying benthic exchange rates and solute dynamics, there are still aspects of sediment disturbance during instrument landing that need to be addressed. Further, the placement of large structures at the seafloor changes the local hydrodynamics and this can potentially affect the benthic biogeochemical conditions (e.g. Reimers et al. 2001; Parker et al. 2002).

Benthic transport and uptake of oxygen

Apart from the shift in the solid-liquid ratio, the major difference between the pelagic and benthic compartments is a change in the processes responsible for solute transport. The open water is dominated by a turbulent flow regime and efficient solute mixing is ensured by eddy transport and advection. However, in cohesive sediments solute transport is mainly mediated by diffusion, even though

fauna-induced irrigation can be important in some areas (see below).

The free-flow velocity and the eddy diffusivity, E , decrease exponentially with the distance to the seafloor, z (Boudreau & Guinasso 1982). This is expressed by $E(z) = \nu A z^P$, where ν is the kinematics viscosity, while A and P represent two constants. The latter typically ranges between 3 and 4 (Shaw & Hanratty 1977) and it follows that eddy diffusivity decreases extremely fast in close vicinity of the seafloor. At the distance where the eddy diffusivity becomes smaller than the kinematics viscosity, the viscous forces start to dominate. This height defines the upper boundary of the 'viscous sublayer' and is typically ~ 5 – 10 mm above the actual seafloor (Boudreau 2001). Even closer to the seafloor the eddy diffusivity falls below the molecular diffusion coefficient (D_o) and hydrodynamically this horizon defines the upper boundary of the DBL and it typically extends one-tenth of the viscous sublayer (Boudreau 2001) (Figure 5). Within the DBL the vertical solute transport is mainly mediated by molecular diffusion.

Diffusion is an extremely fast transport mechanism on small spatial scales; however, as the diffusion time relates to the square of the diffusion distance,

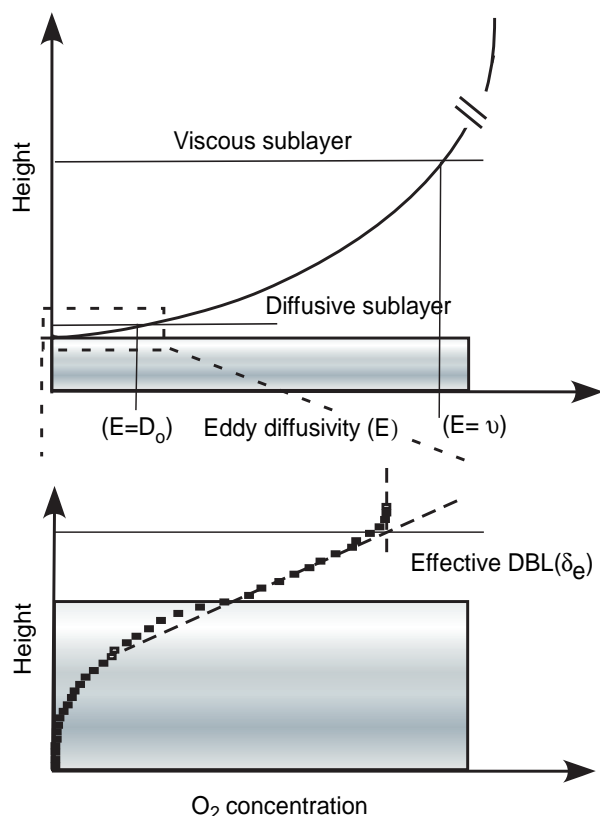


Figure 5. A schematic presentation of the viscous sublayer, the diffusive boundary layer (DBL) and the effective DBL (δ_e) as derived from the O_2 concentration profile (see text for details).

the efficiency progressively declines with the distance (Sten-Knudsen 2002). Biogeochemical process rates at benthic interfaces can thus become limited by mass transport across the 0.5–1.0 mm thin DBL (Schink & Guinasso 1977; Boudreau & Scott 1978; Thibodeaux et al. 1980; Santschi et al. 1983).

The diffusive boundary layer (DBL)

The introduction of O_2 microelectrodes to aquatic biology (Revsbech et al. 1980; Revsbech 1989a) demonstrated the existence of a DBL of relevance for benthic O_2 exchange and allowed detailed investigations of the structure and the dynamics of the DBL (Jørgensen & Revsbech 1985). It was documented that the DBL in general was characterized by a linear O_2 concentration gradient and an 'effective DBL thickness' (δ_e) was defined on the basis of measured O_2 microprofiles (Jørgensen & Des Marais 1990). By this approach, the upper DBL boundary is determined as the intersection between the extrapolated linear concentration gradient of the DBL and the constant O_2 concentration of the overlying well-mixed water phase (Figure 5). For laboratory investigations, the relative position of the sediment surface can be observed directly *via* a stereomicroscope, but can sometimes also be identified by a distinct change in the slope of the concentration profile (Figure 5); this is often used when evaluating *in situ* profiles. The relative position of the sediment surface can in principle also be determined independently by quantifying the change in the reflective index *via* a tapered fiber glued to the O_2 sensor (Klimant et al. 1997a), but the technique is not trivial to apply and has never become widely used. The effective DBL thickness as defined from concentration profiles corresponds well to the thickness defined from hydrodynamic considerations (Jørgensen 2001). Continuous measurements at the respective heights within the DBL have shown how eddies impose the DBL and transplant stochastic O_2 fluctuations down through the DBL (Gundersen & Jørgensen 1990). The fluctuations are damped in the deeper layers of the DBL and vanish in the very upper part of cohesive sediments (~ 50 – 200 μm). The O_2 fluctuations reflect the eddy characteristics, being large and slow at low flow velocities while they are small and fast at high flow velocities (Gundersen & Jørgensen 1990; Jørgensen & Des Marais 1990). In order to obtain the average value for the O_2 concentration within the DBL, it is thus important to monitor sensor signals at the respective heights for longer periods (Røy et al. 2004). Recent investigations argue that the turbulent diffusion (reflected by the fluctuations in O_2 concentrations) adds to the solute transport through the

'effective DBL' and that the true DBL as approximated by power-law scaling is 30% smaller than the microsensor-derived 'effective DBL thickness' (Hondzo et al. 2005).

At the scale of the DBL thickness, the sediment surface is not a flat plane but a landscape with extended topography. The DBL blankets the sediment topography, being thin over protruding structures and thicker over protected interjacent areas, and the DBL surface thus reflects a dampened relief of the sediment topography (Jørgensen & Des Marais 1990). The microtopography has three consequences relevant for the simplified one-dimensional diffusive exchange calculations outlined in Figure 2: (1) the area across which diffusive exchange can take place is larger than the flat plane; (2) vertically measured microprofiles on average protrude the DBL at an angle; and (3) potential horizontal diffusion is not accounted for (Jørgensen & Des Marais 1990). The quantitative importance of these consequences can be evaluated from topographic relief of the sediment (Røy et al. 2002). Such microtopographic 'maps' can be obtained manually (e.g. Gundersen & Jørgensen 1990) – which is an extremely tedious task – or by imposing a horizontal laser-line on the sediment topography and measuring the deflection of the line from a given angle (Røy et al. 2002, 2005). Such measurements have shown that the three-dimensional exchange area for typical coastal sediments is ~5–12% larger than the area of the flat plane (Røy et al. 2002, 2005; Glud et al. 2003). Accounting for both the horizontal diffusion and the larger exchange area, the total diffusive exchange in normal coastal sediments is on average ~10–25% higher than the one-dimensional diffusive flux estimates from vertical microprofiles (Røy et al. 2002, 2005; Glud et al. 2003). For microbial mats, more extreme values of ~50–150% have been presented – these values are not typical, but document the effect in benthic systems with extreme reliefs (Gundersen & Jørgensen 1990; Jørgensen & Des Marais 1990). Apart from the microtopography or the roughness of the seabed, the turbulent energy of the benthic boundary layer (BBL), which can be approximated by the flow velocity of the bottom water, exerts prime control on the DBL thickness. In a series of laboratory experiments, it has been shown how the DBL thickness decreases exponentially with increasing flow velocity (e.g. Gundersen & Jørgensen 1990; Jørgensen & Des Marais 1990; Santschi et al. 1991; Steinberger & Hondzo 1999).

In less energetic, enclosed water bodies (i.e. lakes, embayments or reservoirs) the energy dissipation rate can be out of phase with the current velocity, and in such instances, the BBL turbulence rather than the free flow velocity regulates the DBL

thickness (Lorke et al. 2003). In systems where the diffusion time across the DBL limits the benthic O_2 consumption rate, a reduction of the DBL should increase the O_2 concentration at the sediment surface, enhancing the DOU and increasing the O_2 penetration depth. As an example: a flow-channel study performed on a hypersaline microbial mat showed that by increasing the free flow velocity from 0.3 to 7.7 $cm\ s^{-1}$ the δ_e was reduced from ~560 to 130 μm and the average transport time across the DBL thereby decreased from ~125 to 7 s (Jørgensen & Des Marais 1990). The reduction in the DBL thickness increased the DOU by a factor of 2.4 and thus the DBL thickness clearly impeded the O_2 uptake of the community. This case represents a situation with shallow O_2 penetration depths (~0.1 mm) and fast turnover time of the interstitial O_2 pool (~10–20 s). Such conditions do not apply to normal marine sediments.

Does the DBL control the diffusive mediated benthic O_2 uptake of marine sediments?

The first attempt to directly quantify the *in situ* DBL thickness by microsensors was performed in the O_2 minimum zone off the North American west coast (Archer et al. 1989). Using a video camera mounted on a tripod the position of microelectrode tips relative to the sediment was visually followed during *in situ* microprofiling. From such observations a δ_e range of 0.5–3.5 mm with an average value of 1.3 ± 1.0 mm was estimated for the deep sea (Archer et al. 1989). Some of these measurements were, however, possibly confounded by microtopography. Our present database on published microsensor-derived *in situ* δ_e values is still slim. Values for water depths >1000 m generally range between ~450 and 950 μm , while values measured at the upper shelf or in coastal areas tend to be thinner and generally range between 200 and 700 μm (Figure 6). The trend of increasing *in situ* δ_e values with increasing water depth probably reflects the generally lower current velocities and lower temperatures encountered in deep water – both factors will increase the overall DBL thickness (Jørgensen 2001; Roberts & McMinin 2004).

Archer et al. (1989) assessed the *in situ* DBL impedance by assuming that O_2 was consumed heterotrophically following first-order kinetics with respect to organic carbon and zero-order kinetics with respect to O_2 . It followed that a theoretical removal of a DBL at the average measured thickness would enhance the DOU by ~5–30% at O_2 penetration depths of 10.0–2.5 mm (Archer et al. 1989). Assuming a depth-independent volume-specific O_2 consumption rate, the DOU becomes proportional to the square root of the O_2 concentration at the

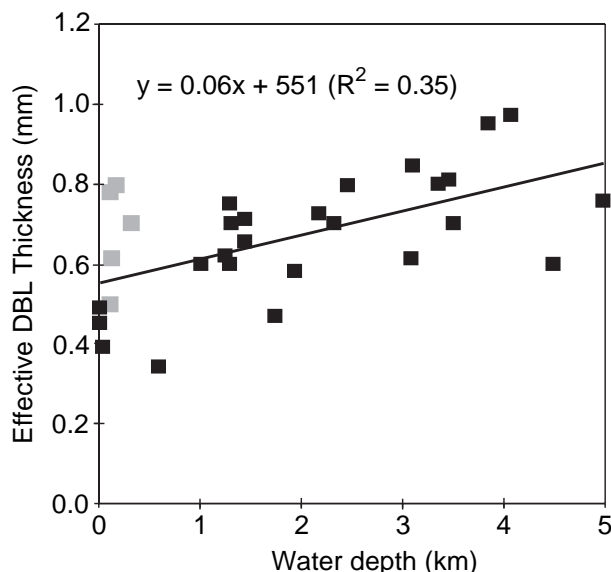


Figure 6. The effective DBL thickness quantified from *in situ* O_2 microprofiles versus the water depth. The grey symbols are measured around Svalbard at temperatures close to 0°C and tend to be above the general trend line (data from Glud et al. 1994a, 1998a, 1999a, 2003; Glud & Gundersen 2002; Lorke et al. 2003; Wenzhöfer et al. 2001a,b).

sediment surface (Bouldin 1968). The increase in DOU upon a theoretical removal of the DBL would thus be proportional to $(C_w/C_o)^{0.5}$ where C_o and C_w are the O_2 concentration at the sediment surface and in the well-mixed water phase, respectively. Evaluating *in situ* O_2 microprofiles by this procedure shows that the estimated DBL impedance becomes almost negligible when the O_2 penetration depth exceeds $\sim 1\text{--}2$ mm (Reimers & Glud 2000). Deep-sea or shelf sediments typically have O_2 penetration depths well above $1\text{--}2$ mm (e.g. Reimers et al. 1986; Reimers 1987) and the O_2 turnover time is expressed in hours (or days). The transport time across the DBL in such environments is negligible and the benthic O_2 consumption is limited by internal diffusion through the pore space rather than the DBL thickness.

The two simplified approaches presented above do not, however, satisfactorily account for the complex kinetics of O_2 consumption in typical marine sediments. The volume-specific O_2 consumption is in most instances sediment depth-dependent and a significant fraction of the benthic O_2 consumption is related to reoxidation of products from anaerobic organic carbon degradation (e.g. NH_4^+ , Fe^{2+} , H_2S). In such instances more complicated and detailed modeling approaches accounting for the mobility of the reaction zones and the kinetics of the different O_2 -consuming processes are required to quantitatively evaluate the DBL impedance (Jørgensen & Boudreau 2001).

A seasonal *in situ* study in central Aarhus Bay, Denmark showed that the O_2 penetration depth annually varied between 0.6 and 4.6 mm (average 2.1 mm), that the turnover time of the internal benthic O_2 pool varied between 2 and 40 min, and that the δ_e varied between ~ 300 and 700 μm (average 451 μm) (Glud et al. 2003). Anaerobic mineralization accounted for the major part of the total carbon mineralization at the site, and thus a significant fraction of the annual O_2 consumption was related to reoxidation processes (Jørgensen 1996b). Obviously, the potential DBL impedance cannot be evaluated as described above. Thus a theoretical investigation was undertaken using a dynamic diagenetic model adapted to the extensive biogeochemical database for the area. The established model included 18 dissolved and solid chemical species, 23 redox-reactions and seven O_2 -consuming processes that linked O_2 consumption to the carbon, nitrogen, manganese, iron and sulfur cycles (Fossing et al. 2002). By applying measured boundary conditions (including the variations of O_2 concentration and temperature of the bottom water), the model reproduced annual variations in measured porewater profiles and benthic exchange rates of the most important biogeochemical solutes (Fossing et al. 2002; Berg et al. 2003b). The annual variations in TOU at a given DBL thickness were mainly driven by variations in sedimentation rates and the O_2 concentration of the bottom water (Glud et al. 2003; see below). By imposing different static values for the DBL thickness and letting all other parameters vary with the seasonal cycle, the theoretical impact of the DBL thickness could be evaluated (Figure 7).

The data showed that during summer, the benthic O_2 uptake gradually increased as the DBL thickness

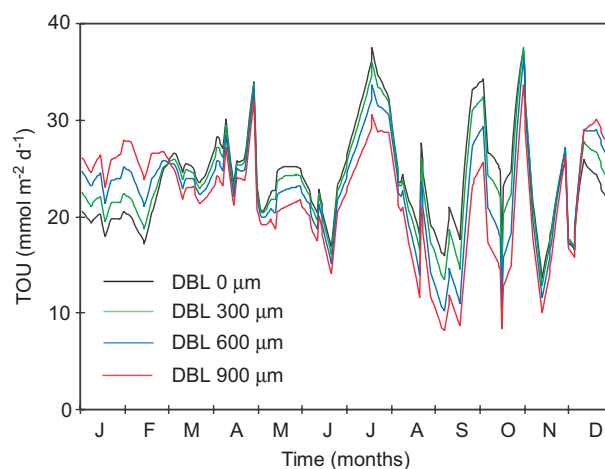


Figure 7. The TOU modeled with the same seasonal scenario but with four different, static DBL thicknesses (redrawn from Glud et al. 2007).

was reduced. However, during winter, the opposite effect was observed. The underlying reason was that during summer, the DBL impeded the O_2 uptake of the carbon-enriched sediment and the reduced solutes and solids from anaerobic degradation accumulated. During winter when the labile organic pool was exhausted the ' O_2 debt' was redeemed, and the demand for O_2 was higher if the accumulation of reduced substances had been intensified during the summer period. The presence of a static DBL thereby dampens the seasonal variations of O_2 uptake (Glud et al. 2007). However, the annual integrated O_2 consumption rate was only marginally affected. Complete elimination of an imposed DBL of 600 μm only enhanced the annual O_2 consumption rate by 5%. The corresponding values at DBL thicknesses of 300 and 900 μm were 2 and 10%, respectively (Glud et al. 2007). Eliminating the DBL led to significant reduction in aerobic mineralization rate, but that was almost balanced by a corresponding increase in the reoxidation processes. The slight reduction in the annual O_2 consumption following DBL imposition was caused by an enhanced benthic release of reduced solutes (NH_4^+ , Mn^{2+} and H_2S) and enhanced pyrite burial (Glud et al. 2007). Estimating the DBL impedance by the simplified original approach as described above predicted a 58% increase in the annual O_2 uptake following the elimination of the 900 μm thick, static DBL. The DBL impedance for the long-term O_2 uptake in typical marine sediments is thus much less than would be anticipated from the simplified approaches not accounting for the kinetics and mobility of the various O_2 -consuming processes. Dynamic modeling showed that theoretical elimination of a DBL increased the DOU and the O_2 penetration depth promptly, but the O_2 uptake soon declined after the immediate stimulation and at steady state the O_2 uptake was only marginally higher (Glud et al. 2007). In other words; the δ_e only affects the long-term DOU of coastal sediments marginally, but sudden changes in the δ_e can affect short-term O_2 distribution and the DOU.

In nature the current velocity (or better the turbulent energy of the BBL) changes constantly and thus the DBL thickness must vary correspondingly. This was shown in a diurnal study of Lake Alpnach where the *in situ* δ_e varied between 160 and 840 μm (average $\sim 390 \mu m$) in phase with changes in the turbulent dissipation energy (Lorke et al. 2003). By applying a proxy for the *in situ* DBL thickness as derived from measured current velocities rather than the static DBL values, the model results revealed that the O_2 concentration within the sediment of Aarhus Bay fluctuated extensively due to an ever-changing δ_e and the porewater profiles thus

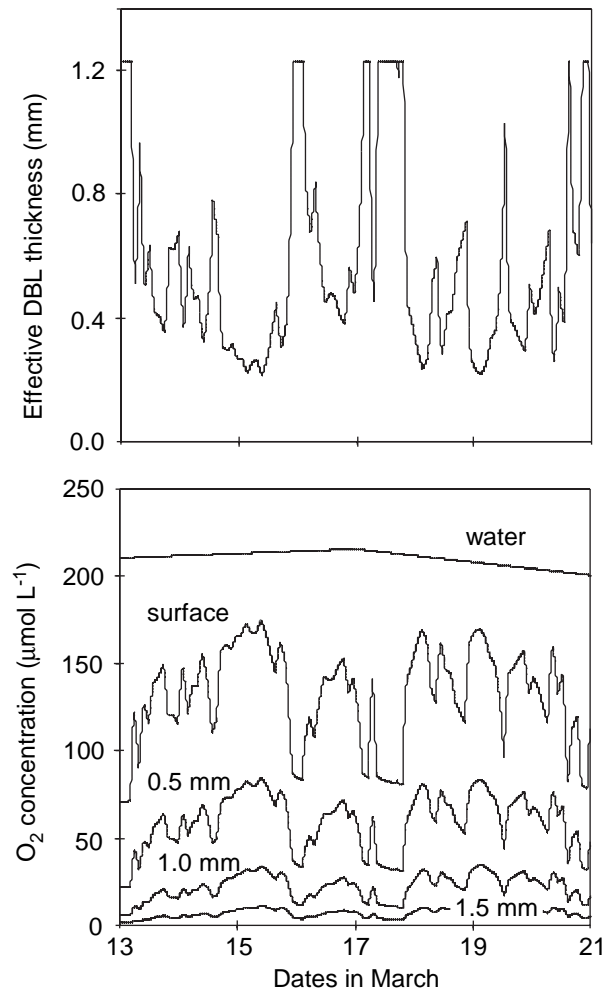


Figure 8. The estimated DBL thickness in central Aarhus Bay during mid-March (upper panel). Periods with constant high values are caused by a truncation of DBL thicknesses above 1230 μm which were considered to be unrealistic. The lower panel shows the O_2 concentration in the bottom water, at the sediment surface, and 0.5, 1.0, 1.5 mm below the sediment surface as derived from the mathematical model – see text (redrawn from Glud et al. 2007).

never reached a steady-state distribution (Figure 8). The DOU fluctuated correspondingly with more than 30% variation within a few hours solely as a consequence of variations in the DBL thickness. However, the average (and annual) O_2 uptake remained unaffected by the variations (Glud et al. 2007). The findings have been confirmed in flume experiments (data not shown), but it remains a non-trivial future task to verify the findings *in situ*. Nevertheless, the theoretical exercise strongly suggests that even though long-term DBL impedance is of minor importance in typical coastal sediments as investigated here, the benthic O_2 distribution is far more dynamic than previously anticipated due to constant changes in the DBL thickness. In essence, short-term variations in the bottom O_2 concentration would have similar impacts on the O_2 dynamics

and O_2 exchange as the variations in DBL thickness discussed above. The biogeochemical and microbial consequences of the short-term O_2 dynamics remain to be investigated.

Microbial mats of chemoautotrophic bacteria often have a very shallow O_2 penetration depth ($<200\ \mu\text{m}$) and an O_2 turnover time of less than a few seconds (Jørgensen & Revsbech 1983). The most widely distributed and best studied mat-forming chemoautotrophic bacteria is the filamentous colorless sulfur bacterium *Beggiatoa* spp. that gains energy by oxidizing H_2S aerobically (Nelson 1992). These bacteria efficiently compete with the chemical oxidation of H_2S with O_2 by enhancing the process rate by a factor of 10^4 – 10^5 (Jørgensen & Revsbech 1983). Most of the O_2 concentration gradient of a well-established mat is within the DBL and the O_2

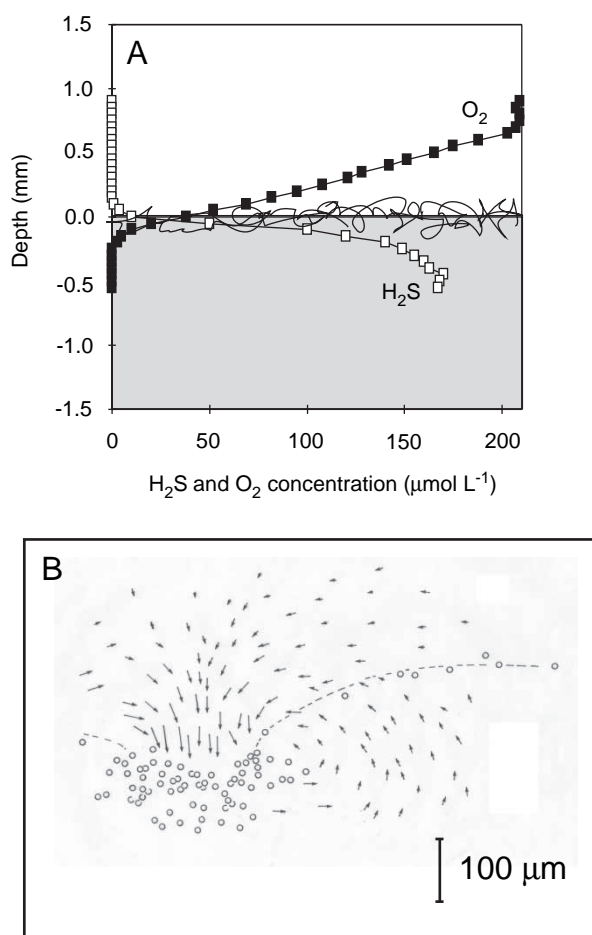


Figure 9. (A) Parallel O_2 and H_2S microprofiles measured across a sediment–water interface with a *Beggiatoa* mat. The two profiles are not measured at the exact same spot, but it is apparent that H_2S is oxidized in the aerobic surface layer. (B) The flow field of water around a cluster of attached, rotating *Thiovulum* cells. The vectors indicate particle displacement within 0.2 s. The dotted line indicates the isoline of 4% air saturation (redrawn from Fenchel & Glud 1998).

distribution and the DOU is truly regulated by the DBL thickness (Figure 9A). The bacteria have a simple but efficient chemosensory behavior that ensures the filaments constantly maintain themselves in the optimal position (i.e. the H_2S – O_2 overlap zone), and the mat structure thus constantly changes with changes in the DBL thickness (Møller et al. 1985). The thin, diffusion-limited *Beggiatoa* spp. mats known from impermeable sediments are contrasted by centimeter-thick mats of giant *Beggiatoa* spp. around hydrothermal vents. Here convective fluid circulation driven by out-simmering hot water overcomes diffusion limitation and ensures convective mixing and supply of H_2S and O_2 to the thick and loose mat structure (Gundersen et al. 1992).

Thiovulum sp. is another well-studied colorless sulfur bacterium gaining energy from aerobic H_2S oxidation (e.g. Garcia-Pichel 1989; Fenchel 1994). *Thiovulum* cells are 5–10 μm in diameter and they are among the fastest free-swimming bacteria (200 – $1000\ \mu\text{m s}^{-1}$). *Thiovulum* follows a helicoid swimming path, and the cells aggregate in zones where H_2S and O_2 co-exist. Under suitable conditions, the cells secrete a thin mucus thread that attaches to solid surfaces, but the anchored cells keep rotating counter-clockwise (seen from the posterior end) in the developing veil (Jørgensen & Revsbech 1983; Fenchel 1994). It was observed that attached cells clustered in groups of 20–100 individuals and that the continued rotation of the bacteria induced small advective flow-cells around the clusters (Figure 9B). Through their self-organization, the attached rotating cells enhanced the counter-current diffusive supply of H_2S – O_2 by a factor of ~ 40 . Oxygen-depleted and H_2S -enriched water emerged in the holes of the bacteria-veil and was mixed with water above the veil that was sucked down towards the center of the cell clusters (Fenchel & Glud 1998). The cells thereby overcame the diffusion-limited supply of their vital metabolites. Recently it has been observed that a large, vibrioid bacteria uses the same mechanism for enhancing the O_2 supply to veil forming communities (Thar & Kühl 2002). The cases presented above are only two examples of adaptations of large, surface-associated chemoautotrophic bacteria for overcoming diffusion limitation in their metabolite supply – other spectacular examples includes bacteria like *Thioploca* spp. and *Thiomargarita namibiensis* (Fossing et al. 1995; Schulz et al. 1999; Schulz & Jørgensen 2001). Filter feeding of sessile ciliates hosting bacteria symbionts represents another fascinating example of how chemoautotrophic metabolism can be stimulated by microscale advection (Vopel et al. 2002).

Experimental considerations for DBL and benthic exchange studies in cohesive sediments

Most microsensor studies have been conducted with the sensor being inserted from above. However, the presence of a microsensor changes the structure of the DBL (Glud et al. 1994b). The reason is still not fully understood but is probably related to flow acceleration around the electrode shaft compressing the DBL below the sensor tip and expanding the DBL on the leeside of the sensor. Values of δ_e derived from micropfiles thus underestimate the actual value by ~25–45% (Glud et al. 1994b). This may explain why alabaster plate dissolution techniques tend to give 25–60% higher δ_e values than microsensor-derived measurements (Santschi et al. 1991; Jørgensen 2001) and the effect should be considered when evaluating the current database on *in situ* DBL thicknesses.

In regions with O_2 penetration depths above a few millimeters, DBL disturbance during sensor insertion hardly affects the interstitial O_2 distribution. However, in coastal environments with shallow O_2 penetration, O_2 micropfiles will often be measured as the O_2 distribution transiently recovers from the DBL perturbation. The quantitative importance of this depends very much on the local conditions and the applied profiling procedure, but the potential impact ought to be better evaluated. In microbial mats with shallow O_2 penetration, a new steady-state O_2 micropfile following a given DBL perturbation will quickly re-establish. But here, the DOU at the measuring spot will be substantially increased due to the microsensor-induced DBL compression. As an example, DOU determinations in a *Beggiatoa* mat were 60% higher when calculations were based on micropfiles measured from above as compared to profiles made from below (Glud et al. 1994b). The DBL compression effect should not be taken lightly when performing microsensor work in extremely active microbial communities or when microsensor-derived solute profiles are aligned to the distribution of solutes, solids or organisms in e.g. biofilms. Oxygen images of a biofilm base growing directly on an O_2 planar optode revealed that the base of all cell clusters was anoxic, while microelectrodes measurements performed from above indicated a complete oxic biofilm. The oxygenation of the biofilm was induced by DBL compression as the microelectrode approached the biofilm (Glud et al. 1998b). Undisturbed DBL measurements can be performed by inserting the microelectrode from below. This can be done in the laboratory but just how this should be accomplished *in situ* is hard to imagine.

Considerable efforts have been dedicated to validate various chamber designs for benthic exchange measurements (e.g. Buchholz-ten Brink et al. 1989; Hüttel & Gust 1992a; Glud et al. 1995a; Smith et al. 1997; Tengberg et al. 2004, 2005). In consensus, the chamber hydrodynamics is of minor importance working in areas where the DBL impedance can be ignored (i.e. O_2 penetration above a few millimeters) (e.g. Reimers et al. 2001; Tengberg et al. 2004). It is, however, essential that the stirring of the chambers is sufficient to ensure a well-mixed water column. For very diagenetic-active cohesive sediments, the DBL impedance can potentially be of importance for the exchange rate (see above) and thus it is advantageous to know the δ_e imposed by the chamber design for later evaluation. Further it is advantageous to have a relatively invariable δ_e covering the chamber-bottom. This can be optimized by selecting different geometric designs or additional pump-systems (Tengberg et al. 2004, 2005). The imposed δ_e can be determined in the laboratory by measuring O_2 micropfiles during incubations at different stirring speeds and water column heights (Glud et al. 1995a). If burrow flushing is significant, it can be advantageous to use chambers with relative small horizontal pressure gradients in order not to induce or enhance passive burrow flushing (Webb & Eyre 2004a).

Interstitial solute transport of cohesive sediments

The diffusive exchange across the DBL can be directly inferred from the concentration gradient and the molecular diffusive coefficient (D_0). However, as molecules in the sediment cannot move along a direct path but have to circumvent the sediment particles, the sediment molecular transport coefficient (D_s) must be corrected for the tortuosity.

The D_s relates to the tortuosity (θ) as $D_s = D_0/\theta^2$ (Berner 1980). It is, however, complicated to quantify the tortuosity in a matrix of particles of different size and geometry as encountered in natural sediment. Thus, there have been established empirical relations between the sediment diffusive transport coefficients (D_s) and the sediment porosity (ϕ): $D_s = D_0\phi^{m-1}$ and $D_s = D_0(1 + n(1 - \phi))^{-1}$, where m and n represent sediment characteristic values of 2–5 and 2–3, respectively, largely depending on the sand content (Ullman & Aller 1982; Iversen & Jørgensen 1993). The constants are not well-defined, and in reality, it is often difficult to objectively select the right value for an analysis and recently a more advanced geometric model expressing the tortuosity dependence on porosity in mud has been proposed (Boudreau & Meysman 2006).

Oxygen microsensors enable us to obtain concentration profiles with a high spatial resolution ($<50\ \mu\text{m}$); this is well below what is possible for porosity determinations. At best, porosity measured by core slicing can be determined by a 3–5 mm depth resolution. Porosity (or tortuosity) can also be approximated by the electric resistance of the sediment (Klinkenberg 1951), but the spatial resolution of such measurements are generally not much better than traditional core slicing (e.g. Andrews & Bennett 1981). Porosity profiles reflect a steep gradient in the upper sediments layers – but the gradients are poorly spatially resolved in the oxic zone. Thus, detailed quantitative analysis of interstitial O_2 microprofiles are essentially limited by the lack of high-resolution profiles of the transport coefficient.

Microprofiles typically have a characteristic break in the O_2 concentration gradient right at the sediment surface (Figure 2, lower panel). This reflects the impeded transport just below the interface. The ratio between the concentration gradient measured just above and below the interfaces has been used to quantify the D_s just below the surface assuming mass conservation of O_2 across the interface (e.g. Epping et al. 1999). The approach requires very detailed microprofiles and does not resolve the depth profile of D_s . High-resolution microprofiles of D_s have been obtained in chemically oxidized and biologically inactivated benthic communities using O_2 as an inert tracer (Revsbech 1989b). The data revealed a strong gradient in D_s within the upper 2.5 mm of a river sediment and that the microbial mat communities had significantly lower D_s than would be derived from the porosity relations above – probably as a consequence of high intracellular water content and water-saturated exopolymers which lead to overestimation of fluid-filled pore spaces (Revsbech 1989b). A more gentle approach using N_2O as an inert tracer in active benthic samples where N_2O production and consumption had been inhibited confirmed these findings (Glud et al. 1995b). The upper part of benthic communities expressed a significant small-scale depth variation in D_s roughly decreasing by $5\% \text{ mm}^{-1}$, but also showed significant horizontal variation. The absolute values for D_s declined after biological inactivation, indicating that meiofauna activity enhanced the D_s values in such communities (Glud et al. 1995b).

An elegant principle for quantifying D_s at high spatial resolution in active benthic communities was introduced by the so-called ‘apparent diffusivity’ microsensor (Revsbech et al. 1998). The sensor contains a gas reservoir (e.g. H_2) enclosed by a glass casing and a permeable silicone membrane at the sensor tip. The concentration of the tracer gas at the sensor tip is continuously monitored and is depen-

dent of the ambient molecular transport rate (Revsbech et al. 1998). The sensor has been applied in different environments and documented extensive small-scale vertical and horizontal variability in D_s (e.g. Revsbech et al. 1998; Elberling & Damgaard 2001; Stief et al. 2004). Wide application of the sensor has, however, been limited by very slow response time (5–20 min), lack of stability, and microbial transformation (or production) of the tracer gas. Recently, a similar sensor design was used to measure flow profiles at extremely slow flow (Brand et al. 2007).

Fine-scale measurements of D_s in benthic communities have also been inferred from pulse field gradient ^1H nuclear magnetic resonance spectroscopy using H_2O as a tracer (Wieland et al. 2001). The approach can non-invasively obtain vertical (or horizontal) images of the apparent diffusivity within benthic communities. Such measurements have confirmed the extensive variation in D_s of microbial mats – a variability that correlated to the distribution of exopolymers and microorganisms (Wieland et al. 2001). The approach is fascinating, but requires very specialized equipment and skills and it will never become a general applicable tool for D_s determination. The approach still remains to be tested in sediments.

A common observation when making detailed O_2 microprofile analyses in very active benthic communities is an inconsistent mass-balance between the DOU derived from the DBL gradient (using D_0) and the DOU calculated on the basis of the concentration profile measured in the community using the D_s – ϕ relations above (Wieland et al. 2001; Rabouille et al. 2003). This could partly be caused by high content of water-saturated exopolymers or the intercellular water. However, inert tracer experiments have also documented that dispersion induced by meiofauna in surface sediments can enhance the interstitial solute transport by a factor of 1.5–2.0 (Aller & Aller 1992; Rysgaard et al. 2000; Berg et al. 2001). One meiofauna group that has a large potential for enhancing the D_s is filter-feeding ciliates which typically have filtration rates in the order of $5\text{--}50\ \mu\text{l day}^{-1}$ (Fenchel 1986). Quantification of filter activity of *Euplotes* spp. and *Uronema marinum* in densities encountered in microbial mats and carbon-enriched sediments indeed documented that this activity enhanced the solute transport of small molecules by a factor 1.1–10 (Glud & Fenchel 1999). This can be especially important in the surface of shallow-water sediments and along burrow walls that often host high densities of ciliated protozoa (e.g. Fenchel 1996b), but the density in typical marine sediments is probably too low to

affect the D_s value significantly (Glud & Fenchel 1999).

Advective solute transport in cohesive sediments

There have been numerous investigations documenting the importance of fauna for benthic diagenesis and solute exchange (e.g. Aller & Yingst 1985; Kristensen 1988; Aller 2001). Deposit- and suspension-feeding fauna play an important role by digesting settling organic material, making it more accessible for subsequent microbial degradation. In the short term, faunal activity creates a heterogeneous interface with tracks, fecal mounds, burrows and funnels that may represent diagenetic hotspots, but in the long term, the fauna serves as an efficient particle mixer that homogenizes the upper sediment layers. Fauna-induced sediment mixing plays an important role for the net downward transport of labile organic material and metal oxides and the net upward transport of reduced metal-sulfur complexes (e.g. Aller 1990; Thamdrup 2000). The associated redox oscillations experienced by organic substrates play an important role in regulating the degradation efficiency especially of more refractive organic material (Aller 1994; Kristensen & Holmer 2001; Meile & van Cappellen 2005). Recent two-dimensional time-lapse movies have presented impressive visualization of bioturbation by following luminophore displacements, thereby quantifying particle mixing rates in the laboratory and *in situ* (Gilbert et al. 2003; Solan et al. 2004).

The sandworm *Arenicola marina* represents a frequently used model organism for studying the effects of bioturbation and bioirrigation (e.g. Meysman et al. 2005; Timmermann et al. 2007). For the shallow Wadden Sea, it has been estimated that on an annual basis this species completely mixes the upper 15 cm of the sediment and irrigates its burrows with $\sim 80 \text{ m}^3$ of water per m^2 (Riisgård & Banta 1998). Even though this represents an extreme case, it illustrates that faunal irrigation can significantly enhance benthic solute exchange. In the shallow North Sea, the activity of *Callianassa subterranean* and *Lanice conchilega* in natural densities enhanced the TOU by 85% (Forster & Graf 1995). Similar values for fauna-enhanced O_2 uptake rates have been reported from laboratory experiments on the polychaete *Hediste diversicolor* (e.g. Banta et al. 1999), the brittlestar *Amphiura filiformis* (Vopel et al. 2003) and from *in situ* chamber measurements on sediments completely dominated by *Trypaea australiensis* (Webb & Eyre 2004b). The relative importance for fauna-induced irrigation of benthic exchange rates obviously depends on the species involved, the size of the specimens, and the local

sediment chemistry. Most investigations on faunal irrigation have for good reasons been performed in areas dominated by a single or few specimens or in well-regulated, small laboratory set-ups, and the findings have then been extrapolated back to *in situ* conditions. To evaluate the relative importance of faunal activity for *in situ* O_2 uptake in mixed natural communities, the most simple and robust procedure is to subtract DOU from the TOU (see above). As outlined above, it is, however, important to incubate as large a sediment area as possible in order not to underestimate the fauna activity.

The present database on simultaneously obtained *in situ* measurements of DOU and TOU is small, but a compilation of the available *in situ* data does reflect a general pattern (Figure 10). The TOU is markedly higher than the DOU in areas with high benthic O_2 uptake rates. Applying the simple trend line of Figure 10A, it follows that at TOU rates $>12 \text{ mmol m}^{-2} \text{ day}^{-1}$ the fauna-related O_2 uptake accounts for $>50\%$ of the benthic O_2 uptake. However, in areas with low diagenetic activity and little fauna, the values converge, and are not significantly different at abyssal water depths – this also provides some confidence in the robustness of two different measuring approaches. The general observation that the relative importance of the non-diffusive O_2 uptake (TOU – DOU) increases with benthic diagenetic activity has previously been reported from smaller data compilations (Jahnke 2001; Meile & van Cappellen 2003). Measurements along the isobar at 1300 m water depth off the West African continent showed that the TOU and DOU aligned well until entering the upwelling zone off Namibia – here the two data-sets diverged (Wenzhöfer & Glud 2002). As the sedimentation rate of organic carbon both depends on the production in the photic zone and the water depth (see below), it is no surprise that the DOU/TOU ratio of the various depth ranges show some scatter as reflected by the error bars in Figure 10C. The database also covers measurements in a wide range of environments presumably dominated by macrofauna of different biology and size – which also will induce scatter in the simple relationship between the DOU/TOU ratio and the water depth. Nevertheless, the general trend is apparent, and the relative importance of fauna-related irrigation for benthic O_2 uptake tends to increase as one moves from the open waters towards the coast (Figure 10C). The diffusive-mediated O_2 uptake completely dominates at water depths below 3500 m, but the fauna-mediated O_2 uptake becomes quantitatively more important at water depths shallower than 1000 m (Figure 10). On average, the DOU only accounts for 43% of TOU in

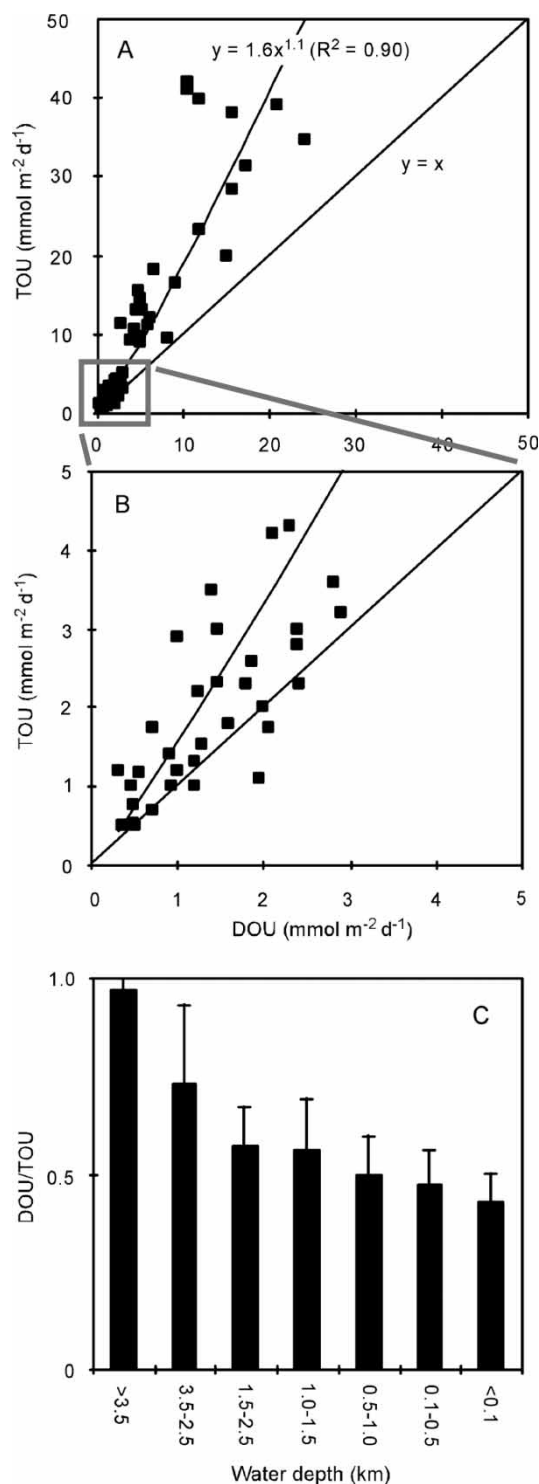


Figure 10. (A, B) *In situ* TOU plotted against parallel DOU measurements for a wide range of non-photic sediments ($n=65$). (C) Ratio between DOU and TOU versus the depth ranges (data from Archer & Devol 1992; Reimers et al. 1992; Glud et al. 1994a, 1998a, 1999a, 2000b, 2003; Forster et al. 1999; Glud & Gundersen 2002; Wenzhöfer et al. 2001a, b, 2002; Wenzhöfer & Glud 2002; Witte et al. 2003a).

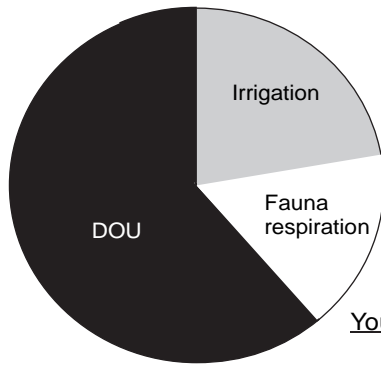
non-photic, cohesive sediments at water depths < 100 m.

The simplest way to quantify faunal abundance is to measure the biomass. However, the weight poorly reflects the biology or the irrigation activity, and mass-specific respiration is size-dependent. Thus, the macrofaunal biomass of natural communities does not correlate well to the ratio or the difference between TOU and DOU even in seasonal studies performed at the same location (e.g. Rasmussen & Jørgensen 1992; Moodley et al. 1998; Glud et al. 2003). The macrofaunal biomass is therefore not a good proxy for the fauna-related O₂ uptake of natural benthic communities.

The general observation that the relative importance of fauna-related O₂ uptake decreases with increasing water depth is probably a combined effect of (1) a relatively steeper decline in macrofaunal biomass than microbial biomass with increasing water depth (Pfannkuche & Soltwedel 1998; Rex et al. 2006); (2) a shift in biology from very active irrigating specimens to less active deposit-feeding specimens when moving towards carbon-depleted sediments (Flach et al. 1998; Pfannkuche 2005); and (3) less quantitative importance of a given irrigation event for the O₂ uptake in highly oxidized (or oxic) sediments. The relative importance of these causes still remains to be fully evaluated.

Fauna-related O₂ uptake is caused by two different processes: (1) fauna respiration and (2) activities that expose otherwise anoxic sediments to O₂ (i.e. irrigation, excavation, turbation). There is no simple way to quantify the relative importance of these two processes. Typically, fauna respiration has been inferred from laboratory measurements on specimens enclosed in inert compartments without sediment (e.g. Banta et al. 1999) or by extrapolating relations between mass-specific respiration rates and biomass (e.g. Gerlach et al. 1985; Mahaut et al. 1995; Heip et al. 2001). In many instances, it is questionable to what extent such approaches reflect the *in situ* respiration rate of the investigated specimens – but alternative procedures are difficult to realize. By quantifying the benthic fauna distributions either by camera observations or by sieving recovered sediment and using various algorithms for extrapolating laboratory-determined respiration rates, it has been estimated that macrofaunal respiration (or megafaunal and macrofaunal respiration together) may represent as much as 25–50% of the TOU (e.g. Piepenburg et al. 1995; Ambrose et al. 2001; Heip et al. 2001). However, in typical marine sediments fauna-mediated O₂ uptake markedly exceeds the respiratory demand of the macrofauna (e.g. Forster & Graf 1995; Hansen & Kristensen 1997; Glud et al. 2000b, 2003; Vopel

Aarhus Bay (st. 6)



Young Sound (st. A)

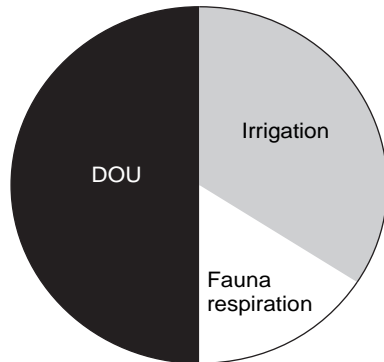


Figure 11. Processes responsible for the O_2 consumption in two coastal sediments. Both data-sets reflect that faunal activity is responsible for a significant fraction of the TOU and that the irrigation is a quantitatively more important process than respiration (data extracted from Glud et al. 2000b, 2003 and updated fauna information from Sejr & Christensen 2007).

et al. 2003) and aerobic microbial activity and chemical oxidation stimulated by irrigation and turbation is often the quantitatively most important component of the fauna-mediated O_2 uptake (Figure 11).

Sediment-dwelling polychaetes represent a quantitatively important infauna group and in natural densities they can significantly enhance the area of the oxic-anoxic interface and the oxic sediment volume (Fenchel 1996a; Kristensen 2000). The burrow lining can be enriched in labile organic material (Kristensen 2000), and the radial geometry ensures a short diffusion distance between the oxic lumen and the highly reduced ambient sediment, facilitating local oxidation of solutes released during the anaerobic mineralization (i.e. Fe^{2+} , Mn^{2+} , H_2S , NH_4^+). Detailed O_2 microsensor studies have shown that the burrow lining can act as hot-spots with intensified O_2 consumption markedly above the activity along the primary interface (e.g. Wenzhöfer & Glud 2004; Jørgensen et al. 2005).

In shallow-water environments where light may reach the sediment surface, diel rhythms in faunal behavior may complicate matters. *In situ* time-lapse

photography has demonstrated that the sediment-dwelling suspension-feeding brittle star *Amphiura filiformis* exhibits a diel rhythm with intensified feeding activity at night time (Rosenberg & Lundberg 2004). This obviously must have implications for the benthic O_2 exchange rate. A detailed *in situ* study performed at a shallow-water, net-heterotrophic sediment completely dominated by juvenile specimens of *Hediste diversicolor* also reflected a distinct rhythm in the TOU (Wenzhöfer & Glud 2004). At dusk, the TOU increased by a factor of ~ 5 to a maximum rate of $\sim 110\text{--}140\text{ mmol m}^{-2}\text{ day}^{-1}$ while the DOU remained at a constant level of $10\text{--}12\text{ mmol m}^{-2}\text{ day}^{-1}$. During the night, the TOU gradually declined to the pre-dusk value. Exclusion of primary production during the night and intensified respiration of leaked photosynthates (see below) during early evening could not alone explain the observation (Fenchel & Glud 2000). Rather, the diurnal variation in TOU was ascribed to light-induced shifts in faunal activity and changes in the bottom water O_2 concentrations. This was confirmed by time-lapse O_2 imaging revealing elevated burrow excavation and ventilation activity and intensified deposit feeding at the surface during night time; this to the extent that the sediment surface was completely remodeled the next day. The TOU during night time corresponded to 70% of the diurnal net O_2 uptake and only 25% of this was estimated to be related to animal respiration while the remainder was caused by a stimulated sediment uptake (Wenzhöfer & Glud 2004). Recently, diel variation in the activity patterns of *Hediste diversicolor* were confirmed in a mesocosm experiment (Tang & Kristensen 2007) and preliminary investigation on *Nereis virens* suggests that subpopulations to various extents exhibit diel cyclic and endogenous rhythmic behavior in response to tide and light (Kim Last, personal communication). To what extent similar events or dynamics are occurring in other areas remains to be demonstrated.

Overall, it can be concluded that fauna-mediated O_2 uptake can be quantitatively significant for the benthic O_2 uptake. This is especially true for the coastal area that represents an important zone for global carbon mineralization (see below). The relative importance of fauna-mediated O_2 uptake (quantified as the difference between TOU and DOU) has also been used to estimate the relative importance of diffusive versus irrigation-related exchange rates for other diagenetically important solutes (Meile & van Cappellen 2005).

Tubes or burrows left by polychaete infauna are often surprisingly stable and can persist for months to years after being vacated (e.g. Aller & Aller 1986). Relict tubes can act as traps for labile organic

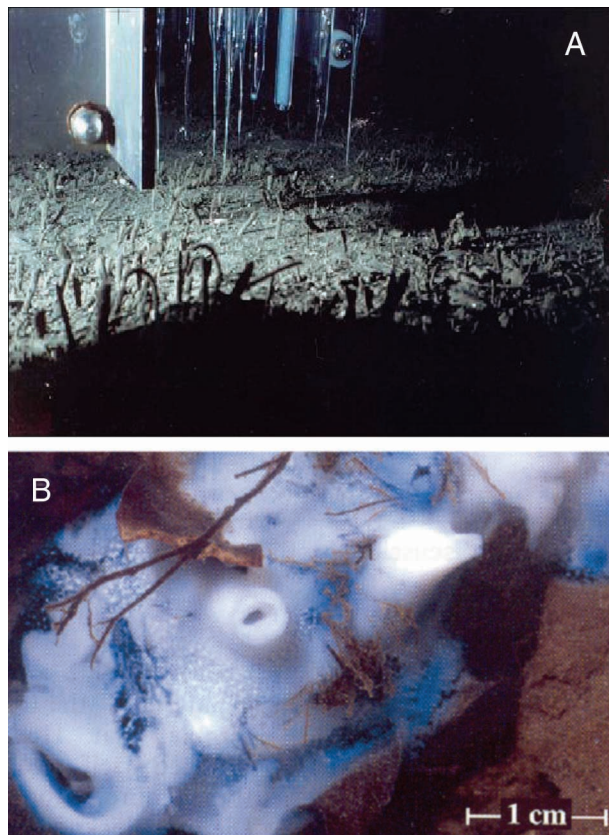


Figure 12. (A) Photograph taken at 155 m water depth in Hornsund, Svalbard. The sediment contains numerous permanent-like tubes extending out of sediment and investigations of recovered sediment documented that many were relict (Glud et al. 1998a; Jørgensen et al. 2005) (photograph by O. Holby and R.N. Glud). (B) 'Chimneys' extending out from a complex microbial mat structure found the high Arctic (redrawn from Glud et al. 2004).

material and thereby represent hot-spots with intensified diagenetic activity (Aller & Aller 1986; Zhu et al. 2006b). To the extent they remain open they enhance the sediment permeability (Weaver & Schultheiss 1983). Pressure gradients between the burrow openings caused by the vertical flow gradient can induce flushing through such relict burrows (Vogel & Bretz 1971; Ray & Aller 1985; Libelo et al. 1994). Despite a large potential for such an exchange mechanism in many sediments (Figure 12A), the effect has rarely been evaluated in natural systems. The sediments of central Skagerrak host a large density of relict, rigid tubes from the polychaete *Spiochaetopterus bergensis* (1300–8000 m⁻²) (Rosenberg et al. 1996; Forster et al. 1999). Detailed microsensor investigations revealed that many of the tubes were exposed to passive flushing in the order of 5–11 µl min⁻¹ at moderate, realistic flow velocities of 2–6 cm s⁻¹ (Munksby et al. 2002). The flushing induced dynamic, suboxic/anoxic plumes downstream of the tubes and created a very

dynamic and heterogeneous O₂ distribution along the sediment–water interface. The flushing enhanced the O₂ consumption rate by a factor of ~2 (Munksby et al. 2002). The activity must have profound impact on the local biogeochemistry where the heterotrophic activity is dominated by microbial manganese respiration (Canfield et al. 1993). The complex flushing may act as a shunt in the upward net-transport of Mn²⁺ that would otherwise be driven by diffusion (Thamdrup et al. 1994) and burrow excavation may efficiently bury the precipitating manganese oxides below the oxic zone, maintaining high microbial manganese respiration.

Another example of passive tube flushing was recently encountered in a conspicuous H₂S-oxidizing microbial mat from the high Arctic (Glud et al. 2004). The centimeter-thick, flocculent mat covered loosely packed, decaying macroalgae and exhibited a complex three-dimensional structure dominated by cracks and chimney-like protrudings (Figure 12B). Microsensor measurements documented the advection of solutes within the mat and that anoxic sulfide-rich water emerged through the protruding tubes presumably being replenished through larger basal openings in the mat coverage (Glud et al. 2004). The complex, flexible structure ensured a flow-induced advective transport of H₂S, O₂ and dissolved organic matter throughout the entire mat structure, which grossly enhanced the benthic exchange rate. Similar mats have now been observed in a range of Arctic fjord systems in Greenland, often covering several 100 m² of the seafloor.

Advection in permeable sediments

Globally, cohesive, impermeable, sediments dominate the seafloor. However, it has been estimated that 70% of the coastal shelf is covered by relict sands (Emery 1969) and that a little more than 50% of the relict sand areas are permeable (i.e. $k > 5 \times 10^{-12}$ m⁻²) (Hüttel & Gust 1992b). As sand generally holds little organic material, sands have been considered to only have minor importance for local and regional carbon cycling. However, the little organic material that is held by sand is highly labile not being diluted by accumulating less degradable material facilitating a high carbon turnover rate. Another obstruction to sand studies has been the simple fact that it is very difficult to sample intact undisturbed sand cores, and until recently, sand has received relatively little attention as compared to cohesive sediments. From a biogeochemical point of view, the prime difference between cohesive and permeable sediments is that advection adds to the transport of solutes and solids (Thibodeaux & Boyle 1987). The driving force for the advective transport

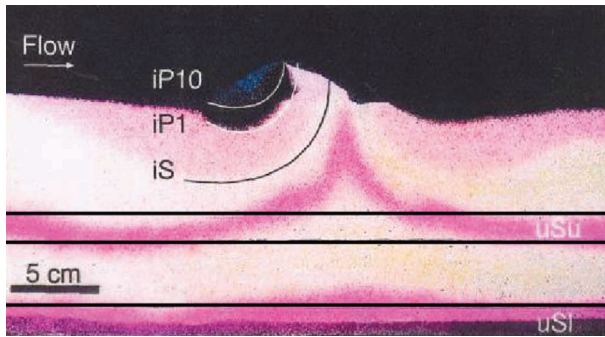


Figure 13. The image shows the distribution of red solutes and 1 and 10 μm large black particles after a laminar flow of 10 cm s^{-1} had been imposed on a sediment hump in sand for 16.5 h. Initially all color in the sediment was dissolved only in the two horizontal layers bordered by the black lines and particles were only added to the overlying water. The image clearly shows that particles and solutes were forced through the pore spaces at the front of the hump and that partial pressure differences induce a local upwelling of porewater at the top and behind the hump (redrawn from Huettel et al. 1996).

can either be fauna activity, differences in water density, wave action or interactions between flow gradients and topographic structures that induce pressure differences forcing water through the sediment matrix (Figure 13).

Advective porewater transport means that the permeable seabed functions somewhat like biocatalytic filters as known from purification treatment in recirculating water systems. Advection markedly enhances benthic solute exchange rates as compared to a situation with only diffusive-driven exchange (Huettel & Webster 2001). In natural environments with constantly changing wave action, flow characteristics and bottom topography, the benthic O_2 distribution becomes extremely dynamic (Ziebis et al. 1996; Precht et al. 2004; de Beer et al. 2005; Cook et al. 2007a). Oxygen can be advected deep into the sediment, and the O_2 penetration depth in such areas may exceed several centimeters (de Beer et al. 2005). However, in other regions, outwelling induced by partial pressure differences may force anoxic porewater right to the sediment surface (Ziebis et al. 1996). As sand ripples migrate along the seabed, down- and upwelling regions may progressively move along with the topography (Precht et al. 2004). Thereby, a stationary volume of sediment below the moving ripples experience repetitive cycles of oxia and anoxia and upon changing flow conditions, pockets of oxygen may stay behind in the otherwise anoxic sediment (Cook et al. 2007a). If the ripples move too fast, the upwelling regions will never establish, and a several centimeters broad oxic zone may develop (Precht et al. 2004). Advection can stimulate the degradation of organic material buried deep in the sediment

(Franke et al. 2006). The volume-specific O_2 consumption (R_{vol}) in permeable sediments has been estimated from slurry incubations and by following the O_2 decline in the pore fluid after the advective O_2 supply has been eliminated (e.g. Precht et al. 2004; de Beer et al. 2005; Polerecky et al. 2005; Franke et al. 2006; Cook et al. 2007a). The few available values generally range between 0.5 and $12\text{ }\mu\text{mol cm}^{-3}\text{ day}^{-1}$ with two outliers in the range of $20\text{--}35\text{ }\mu\text{mol cm}^{-3}\text{ day}^{-1}$. The values compare with values from coastal cohesive sediments even though they tend to be in the lower end of the range (see below). However, given the wider oxic zone of permeable sediments, the values indicate a potentially high TOU of sand. Only a few chamber or core measurements have been performed in sandy sediments, and they generally reflect lower TOU than measurements performed in muddy sediments (Cook et al. 2007a, and references therein). However, the vast majority of these studies have not considered imposing advection during the incubations and, in most instances, percolation has probably been markedly less as compared to the *in situ* conditions. The compiled database for sandy sediments may thus severely underestimate the *in situ* TOU (Cook et al. 2007a).

It is not trivial to perform chamber incubations while maintaining an advective component driving a controlled percolation of the sediment. It can be done by applying circular, centrally disc-stirred hydrodynamically well-calibrated chambers (Huettel & Gust 1992a; Khalili et al. 1997). The stirring imposes a known pressure gradient and by knowing the sediment permeability, a controlled percolation rate can be induced (Glud et al. 1996a; Janssen et al. 2005). However, care should be taken when performing such incubations as it may take relatively long preincubations before quasi-steady-state O_2 distributions have been established. Simple hydrodynamic-biogeochemical modeling routines can be valuable tools for predicting the correct timing (Cook et al. 2007b). Alternative procedures for *in situ* incubation of sandy sediments include flume-like chambers that induce a unidirectional flow across the natural topography of the seabed (Webb & Eyre 2004a,b).

The TOU and DIC release rates as measured in sandy sediments generally increase with the imposed percolation rate, and the rates can be very high as compared to incubations of cohesive sediments (Figure 14). As discussed above, a steady state between anaerobic heterotrophic activity and the oxic reoxidation of the released inorganic metabolites is assumed when using the O_2 uptake rate as a proxy for the total turnover of organic material. This assumption is rarely fulfilled in sandy sediment

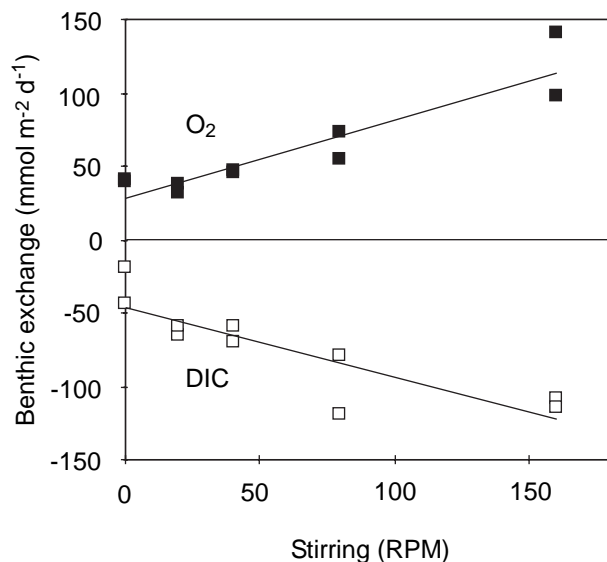


Figure 14. Benthic exchange rates of O₂ and DIC plotted against the stirring rate in two parallel circular chambers. The sandy sediment had a permeability of $2 \times 10^{-11} \text{ m}^2$ and was pre-incubated at the respective stirring rates for 16–20 h before the actual incubations were initiated. The included trend lines should not indicate a linear response. 0 RPM means that the overlying water was gently mixed without establishing a static partial pressure (redrawn from Cook et al. 2007a).

exposed to a constantly changing advective pore-water transport. Reduced solutes (and solids) will accumulate in periods with low percolation rates and out-flushing and reoxidation of metabolites will be initiated when percolation is induced. Even in a situation where the O₂/DIC exchange ratio is close to 1 after a pre-incubation period (as in Figure 14), the exchange rates could reflect a central out-flushing of accumulated DIC and concurrent reoxidation of reduced solutes and solids rather than actual carbon mineralization rates. Therefore, it is difficult to relate total benthic exchange rates of O₂, DIC or nutrients to the actual organic carbon mineralization rate in permeable sediments. Most of the high R_{vol} or TOU rates obtained in sandy sediments probably reflect intense reoxidation of accumulated reduced inorganic components. A reasonable estimate of benthic carbon mineralization rate in sandy sediments is far more difficult than for cohesive sediments and it requires a suite of complementary percolation and slurry techniques preferentially combined with modeling approaches (Cook et al. 2007b).

Permeable sandy sediments gradually become clogged as material is advected into the interstice and as the microbial community continues to grow. In order to maintain their biocatalytic filter capacity, sandy sediments must be re-set. This happens

during storm events or by strong tidal currents, where huge amounts of sands are resuspended, and moved. However, in less vigorous environments, the activity of infauna like *Arenicola marina* might also be important for keeping sands permeable (Volkenborn & Reise 2006).

Subtidal sandy sediments often host relatively few macroscopic animals, but one of the more conspicuous – apart from *Arenicola marina* – is the lesser sandeel (*Ammodytes tobianus*). It is peculiar among fish in that it buries itself in the sediment during the night, when frightened, and in the winter months. The behavior presumably lowers the predation level, and may also represent an energy-saving strategy. Planar O₂ optode measurements showed how the fish during most of the time advected oxygen toward the mouth by gill ventilation, but occasionally channeled oxygenated water down along the body by wrinkling movements. Mass-balance calculations revealed that the fish lowered its O₂ requirement by 25–30% when buried (Behrens et al. 2007). With estimated average winter densities of 60 m^{-2} (up to 200 m^{-2}) in the central North Sea (Høines & Bergstad 2001), the buried fish would enhance the benthic O₂ uptake by a factor of up to two, of which approximately half is related to the respiration of the fish (Behrens et al. 2007).

There are still only a very limited number of investigations on benthic O₂ dynamics, carbon mineralization and primary production (see below) in permeable sediment and it remains an important challenge to quantitatively resolve the importance of sandy sediments for marine carbon cycling.

Benthic oxygen uptake rates in cohesive sediment from the coastal zone to the deep sea

As outlined above, benthic O₂ uptake represents a relatively robust proxy for total benthic mineralization rate in cohesive sediments. Therefore, the database on *in situ* benthic O₂ exchange rates has been steadily increasing since the first bell jar incubations (Pamatmat & Fenton 1968) and the pioneering lander work (Smith et al. 1976; Reimers 1987). Even though certain areas of the deep-sea and polar regions are under-sampled and we have very little information from complex and dynamic sediments like permeable sand, rocky shores and fluidized muds (Aller 2004), the database has reached a size allowing generalizations and approximations of regional and global benthic O₂ uptake rates (Jahnke 1996; Christensen 2000; Wenzhöfer & Glud 2002; Andersson et al. 2004; del Giorgio & Williams 2005).

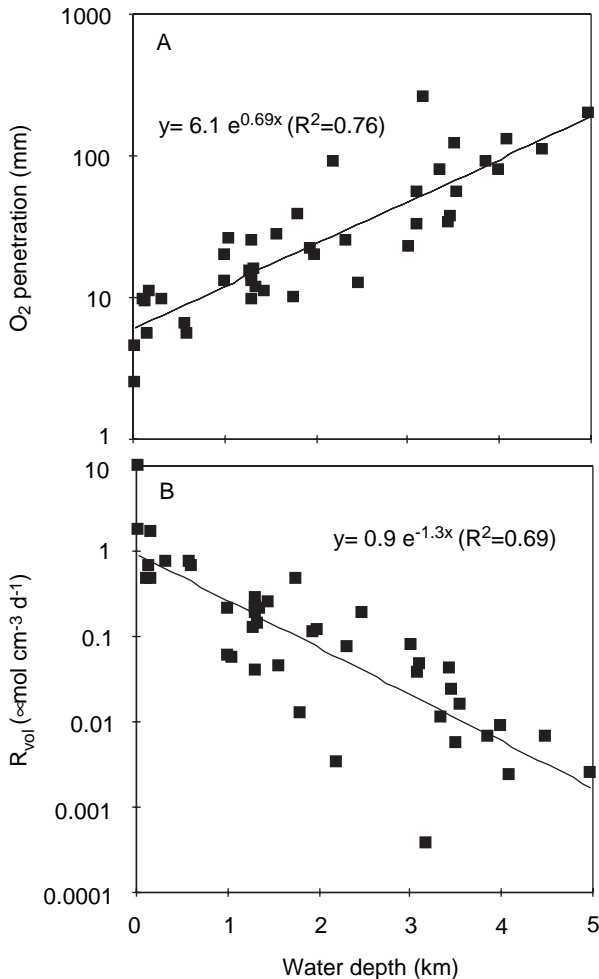


Figure 15. The *in situ* O_2 penetration depth (A) and the average *in situ* volume-specific O_2 consumption rates (B) as a function of water depth compiled from a broad range of non-photic environments. R_{vol} is calculated by dividing the DOU with the O_2 penetration depth (data extracted or calculated from Glud et al. 1994a, 1998a, 1999a, 2003; Wenzhöfer et al. 2001a,b; Wenzhöfer & Glud 2002).

Large-scale gradients

The *in situ* O_2 penetration depth increases exponentially with water depths from a few millimeters in coastal depositional areas to >10 cm at abyssal sediments (Figure 15). Recent onboard measurements indicate that O_2 may even penetrate sediments to several meters in the central oceanic gyres (T.G. Ferdelman et al., personal communication). Rather than changes in O_2 availability, this general pattern reflects a gradual decrease in the average volume-specific O_2 consumption of the sediment, which decreases by more than four orders of magnitude from an average of $\sim 10 \mu\text{mol cm}^{-3} \text{ day}^{-1}$ at 16 m water depth in central Aarhus Bay to $0.003\text{--}0.0004 \mu\text{mol cm}^{-3} \text{ day}^{-1}$ at 3–5 km water depth in the Atlantic and Pacific oceans (Figure 15). The attenuation in aerobic activity mirrors the

general trend of decreasing organic matter sedimentation rate with increasing water depth (e.g. Berger et al. 1988), which also is reflected in the decreasing biomass of benthic fauna and active microbes (e.g. Queric et al. 2004).

Likewise *in situ* rates of TOU and DOU decline with increasing water depth. The compilation depicted in Figure 16 is based on 52 studies covering a wide geographic range and contains 142 DOU and 232 TOU observations. Photosynthetically active sediments have been excluded, and a few data points in strongly O_2 -depleted environments (i.e. intense upwelling zones or semi-enclosed basins) which otherwise would confound the general depth relation have been omitted. The data-sets reflect a large scatter without any latitude trend, but are reasonably well fitted by two simple power functions (Figure 16). The fit is not corrected for skewness or the fact that some data points, to varying extent, reflect averages of several observations. Overall, benthic O_2 uptake rates decrease by more than three orders of magnitude from coastal fauna-rich environments to

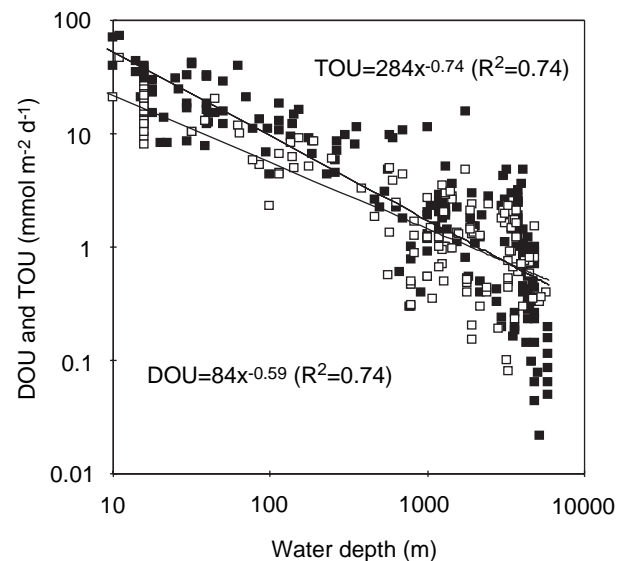


Figure 16. The total O_2 uptake (closed symbols) and diffusive O_2 uptake (open symbol) plotted as a function of the water depth (the compiled data-set only includes *in situ* data as extracted from Pamatmat & Banse 1969; Pamatmat 1971; Smith 1978, 1987; Smith et al. 1978, 1994, 1997; Hinga et al. 1979; Rowe & Gardner 1979; Berelson et al. 1987, 1990, 1996; Reimers 1987; Jahnke & Christensen 1989; Jahnke 1990; Jahnke et al. 1990, 1994; Pomeroy et al. 1991; Archer & Devol 1992; Reimers et al. 1992; Pfannkuche 1993; Boucher et al. 1994; Glud et al. 1994a, 1998a, 1999b, 2000b, 2003, 2005; Hales et al. 1994; Miller-Way et al. 1994; Sayles et al. 1994; Gundersen et al. 1995; Hammond et al. 1996; Nielsen & Glud 1996; Tahey et al. 1996; Duineveld et al. 1997a,b; Epping & Helder 1997; Hales & Emerson 1997; Lohse et al. 1998; Moodley et al. 1998; Forster et al. 1999; Witte & Pfannkuche 2000; Sauter et al. 2001; Wenzhöfer et al. 2001a,b, 2002; Glud & Gundersen 2002; Wenzhöfer & Glud 2002; Berg et al. 2003a; Rabouille et al. 2003; Ståhl et al. 2004a,b; Witte et al. 2003a).

the central gyres of the open oceans. In correspondence to Figure 16, the DOU and TOU seem to converge with increasing water depth, but as expected from a general compilation (rather than only using studies with parallel DOU and TOU measurements) the attenuation coefficients of the two data-sets are not significantly different.

Combining the relationship for TOU versus water depth with the bathymetry of the global ocean (as extracted from the ETOPO5 global topography data-set), the global benthic O_2 consumption amounts to $152 \text{ Tmol } O_2 \text{ year}^{-1}$. Assuming a respiratory quotient (i.e. the DIC/ O_2 exchange ratio) of 1.2, this corresponds to an annual global carbon mineralization rate of 1.52 Gton C (the corresponding values for the DOU relation amounts to 1.26 Gton C). Continental margins (0–200 m) contribute 39%, continental slopes (200–2000 m) 17%, while the deep sea (>2000 m) accounts for 44% of the global benthic carbon mineralization. Using a more conventional definition of the deep sea (i.e. >1000 m), the mineralization rate of this compartment amounts to $74.7 \text{ Tmol } O_2 \text{ year}^{-1}$, equivalent to roughly 50% of the global benthic carbon mineralization. Previous estimates based on other approaches range from 54.3 to $79.6 \text{ Tmol } O_2 \text{ year}^{-1}$ for the same depth interval (Jahnke 1996; Christensen 2000). These values bracket the simplified approach above. The simple extrapolation above predicts that sediments in the depth range of 0–1000 m are responsible for half the benthic mineralization in the ocean, even though they only account for 7% of the ocean area (Figure 17).

The value for coastal areas is poorly constrained as sandy environments are grossly under-sampled and sites expressing net benthic photosynthesis are excluded. Shallow water sediments with periodically

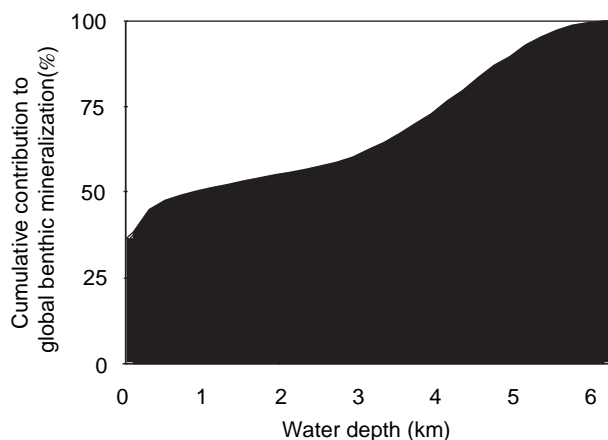


Figure 17. The cumulative contribution of sediments in various depth ranges to the global benthic mineralization. The figure is calculated on the basis of the relations in Figure 16 and the bathymetry of the oceans.

benthic primary production probably represent areas with very high turnover rates of organic material (e.g. del Giorgio & Williams 2005). But generally the quantitative importance of benthic primary production and its recycling is poorly confined (see below).

The average annual primary production of the water columns overlying sediments in the following depth ranges: 0–200, 200–2000, >2000 m, have been estimated to 230 , 150 and $94 \text{ g C m}^{-2} \text{ year}^{-1}$, respectively (Wollast 1998). It follows from the simple calculations above that benthic mineralization amounts to 15.2, 5.4 and 2.3% of the primary production in these respective compartments. The remainder is primarily respired in the water column and to lesser extent retained in the sediment record.

In a recent compilation (including shipboard incubations at water depths <1500 m) the TOU–depth relation was approximated by an equation hosting two exponential relations to account for depth variation in the vertical transport of particulate organic material: $\text{TOU} = K((1-p)e^{-az} + pe^{-bz})$, where z represents the water depth, and K , P , a , b , are various fitting coefficients (Andersson et al. 2004). The approach improves on the significance of a single exponential fit (to $R^2 = 0.68$, $n = 490$), but the fitting coefficients were not well constrained even though their inclusion is well rationalized when evaluating changes in degradation kinetics of sinking aggregates (Andersson et al. 2004). Using this approach the global benthic O_2 consumption (water depths >200 m) was estimated to be $157 \text{ Tmol } O_2 \text{ year}^{-1}$, while values for the deep sea (defined as water depths >1000 m) became as high as $129 \text{ Tmol } O_2 \text{ year}^{-1}$ – well above any previous (and the present) estimates (Andersson et al. 2004).

Water depth is the one single parameter that best correlates to measured values of benthic O_2 uptake – this has been concluded from a number of investigations (e.g. Wenzhöfer & Glud 2002; Andersson et al. 2004). However, as indicated by the scatter of Figure 16, benthic O_2 consumption differs substantially between sediments resting at the same water depth. Causes include variations in the O_2 concentration of the bottom water, the temperature and the faunal community structure. Furthermore, vertical carbon transport velocities are highly variable and differ between settings (e.g. Berelson 2002) and lateral transport of organic material down-slope of margin sediments has been observed (e.g. Hecker 1990) and probably explains intensified degradation activity along some ocean margins (Jahnke et al. 1990). However, the most obvious single parameter that varies between sediments along isobaths is the surface primary production. Plotting benthic O_2

exchange rates for the South Atlantic Ocean against the water depth reveals clustering where data from upwelling areas follow an upward-shifted attenuation curve as compared to data from less productive areas (Figure 18).

To account for spatial variability in primary production, double exponential fitting equations for TOU and DOU in the South Atlantic Ocean were developed: $\text{TOU}_C = \text{PP}^{1.407} z^{-0.492}$ and $\text{DOU}_C = \text{PP}^{0.736} z^{-0.331}$, where the subscript C indicates a conversion from O_2 to carbon equivalents, PP represents the primary production ($\text{g C m}^{-2} \text{ year}^{-1}$), and z the water depth (m) (Wenzhöfer & Glud 2002). The improved fitting equations were used to construct maps of benthic O_2 uptake rates using digitized maps of the two variables: the water depth (from the ETOPO5 database) and the annual primary production estimated from satellite mapping (Behrenfeld & Falkowski 1997). The maps reflect the observed regional differences, and the importance of primary production for the spatial variations in benthic O_2 exchange rates (Figure 19). Likewise, maps for the fauna-mediated O_2 uptake and the O_2 penetration depth were constructed (Wenzhöfer & Glud 2002).

Extrapolated to the entire South Atlantic Ocean (water depth >1000 m), the TOU_C amounted to $0.17 \text{ Gton C year}^{-1}$ while the value for DOU_C was $0.13 \text{ Gton C year}^{-1}$ corresponding to 2.1 and 1.7% of the estimated annual primary production for the area (Wenzhöfer & Glud 2002). Extrapolated to the entire Atlantic Ocean, the values were 161 and 128% higher than estimates made by Jahnke

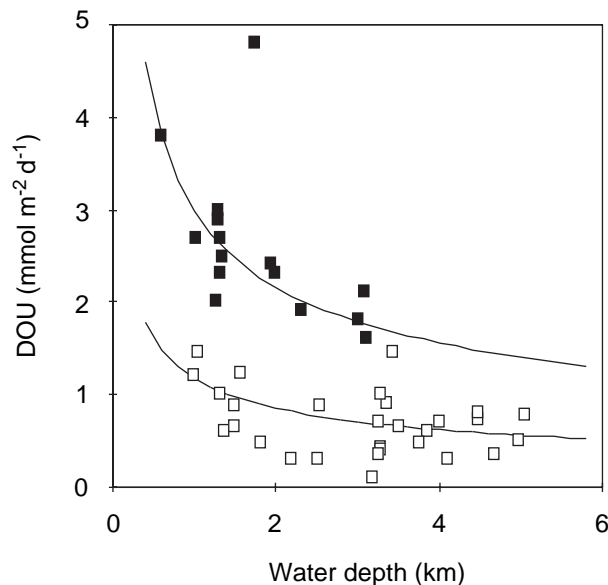


Figure 18. *In situ* DOU measured in areas of high (closed symbols) and low productivity in the Atlantic Ocean. Curve fit described in the text (redrawn from Wenzhöfer & Glud 2002).

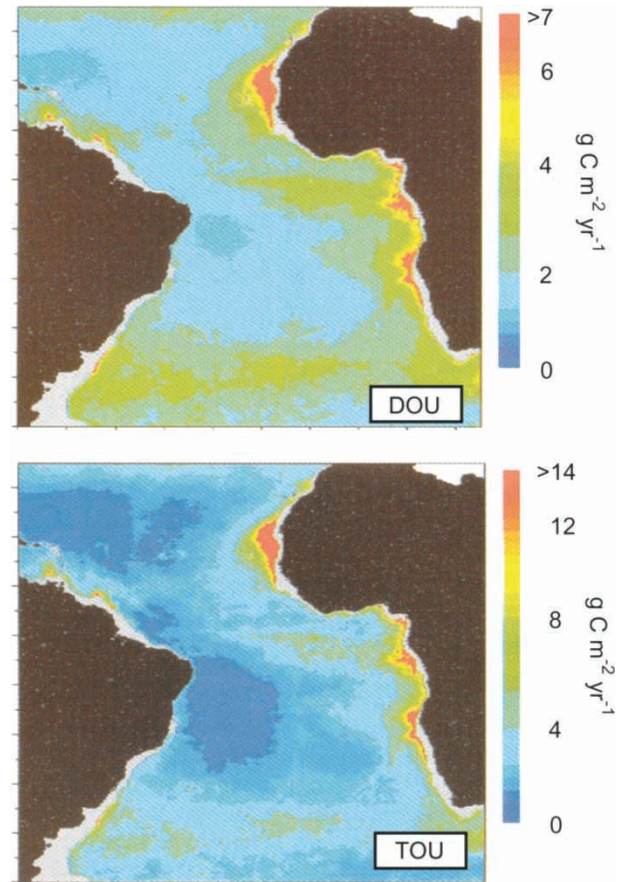


Figure 19. Calculated maps of the DOU and TOU for the sediments of the Atlantic Ocean (redrawn from Wenzhöfer & Glud 2002).

(1996), and 107 and 85% of estimates made by Christensen (2000) using other and different approaches to constrain the rates of benthic mineralization of the Atlantic Ocean.

The particulate organic carbon (POC) fluxes inferred from general empirical relations between primary production and sediment trap data (Berger et al. 1988), or directly measured by sediment traps in the vicinity of our lander deployments (Fischer et al. 2000), were not sufficient to support either the measured or the extrapolated estimates for benthic carbon requirement (Wenzhöfer & Glud 2002). This was especially true close to upwelling areas but also in some of the deeper sites. Inconsistency between sediment trap data, vertical flux estimates and measured benthic requirements has been observed in a number of other studies (e.g. Smith 1987; Smith & Kaufmann 1999; Witte & Pfannkuche 2000; Andersson et al. 2004; Seiter et al. 2005).

An impressive eight-year time series in the eastern North Pacific revealed a consistent deficit in the POC supply as measured by sediment traps compared to the benthic carbon requirement quantified from *in situ* TOU measurements (Smith et al. 2001).

On average, the POC flux only supported 57% of the measured benthic carbon requirement. The discrepancy can be related to consistent underestimation of POC flux by trap approaches, lateral down-slope transport of organic carbon, and/or quantitative important benthic uptake of dissolved organic matter. Alternatively, in some regions the discrepancy can truly represent an imbalance induced by climatic changes that ultimately will lead to a benthic community shift, but as the benthic respiration in deep-sea environments to a large extent often relies on accumulated organic material, the transition can extend over several decades (Smith & Kaufmann 1999; Smith et al. 2001).

When comparing data-sets from sediment traps and benthic community respiration it is important to realize that the two measurements integrate different time-scales. Even though sediment O_2 uptake increases immediately upon carbon enrichment (e.g. Witte et al. 2003a), a fraction of the organic material degraded at any time is more refractory and may have half-life times that scale with months to centuries (e.g. Fenchel et al. 1998). Likewise, a fraction of the benthic O_2 consumption is used for reoxidation of products from anaerobic degradation that might have accumulated on long time-scales (see Introduction). Thus, the total benthic O_2 uptake at a given time integrates a longer time-scale than covered by most sediment trap data (see next section). However, having said that, the compiled data-set in Figure 16 probably only rarely includes events of short-term-intensified activity following bloom settlement.

Despite a steady increase in the database of high-quality benthic O_2 exchange rates, the investigated sites only represent a minute fraction of the global seafloor, and many areas are highly under-explored (e.g. the polar regions, the Indian Ocean and the central gyres). Further, we still know very little about inter-annual or seasonal variability in the benthic O_2 uptake especially in deeper waters.

Seasonality of the benthic oxygen uptake

It is well known that the pelagic primary production shows distinct seasonal patterns governed by nutrient and light availability. In many temperate, coastal environments, the dynamic is characterized by a distinct spring and autumn bloom that is reflected in the vertical POC transport (e.g. Olesen & Lundsgaard 1995). Numerous laboratory investigations have simulated sedimentation of plankton blooms and followed the benthic response that is characterized by a large and immediate increase in the benthic O_2 uptake and a gradual decline to the background level within a number of days to weeks

(e.g. Hansen & Blackburn 1992). However, even though such responses have been observed *in situ*, the response is often less distinct in coastal waters. This is caused by reduced O_2 availability in stratified waters during bloom settlements limiting the potentially high O_2 uptake of the carbon-enriched

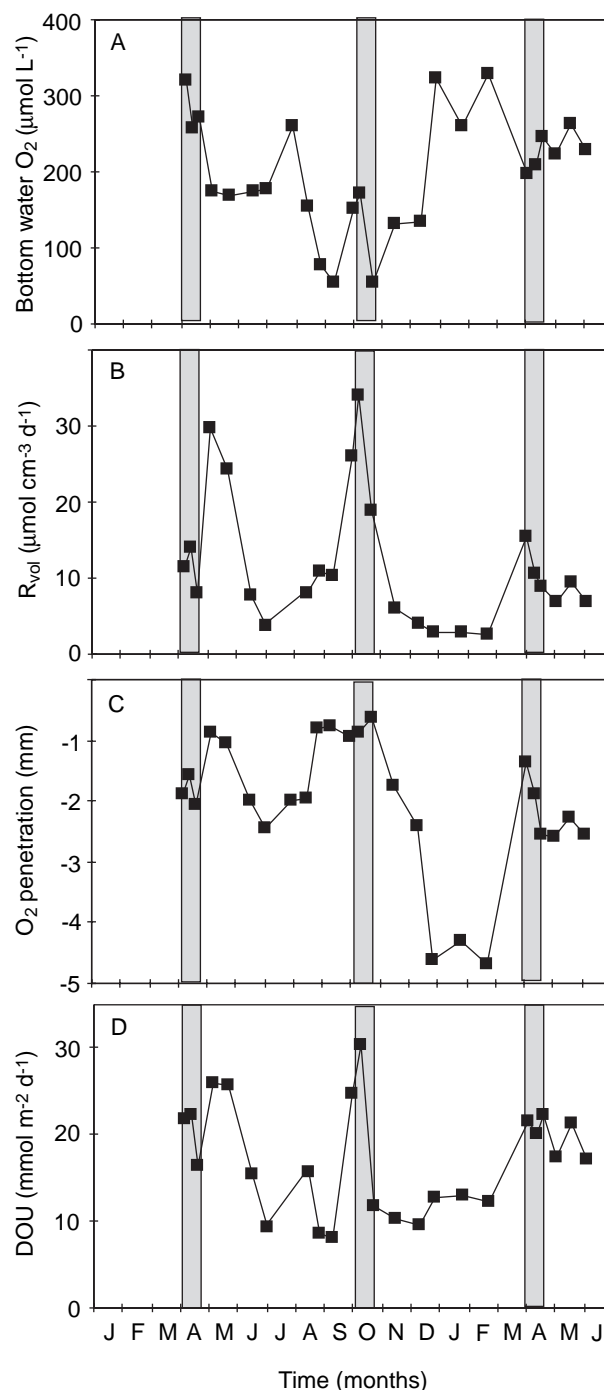


Figure 20. Seasonal variations in the bottom water O_2 concentration (A), the average volume-specific O_2 consumption (R_{vol}) (B), the O_2 penetration (C) and the DOU (D) measured in central Aarhus Bay (1990–1991). Grey bars reflect periods with intensified sedimentation (redrawn from Glud et al. 2003).

sediment (e.g. Rasmussen & Jørgensen 1992; Cowan et al. 1996).

During a 1.5-year seasonal study in Aarhus Bay, the bottom water O_2 concentration decreased by 40–60% following the sedimentation of each of three successive blooms (Figure 20). At each event, there was a distinct increase in the average volume-specific O_2 consumption rate (R_{vol}), while the O_2 penetration depth decreased as a consequence of the intensified activity and the reduced O_2 availability (Figure 20). On a seasonal basis, the two parameters varied by a factor of 13 (from 2.5 to 34 $mmol\ cm^{-3}\ day^{-1}$) and 9 (from 0.5 to 4.5 mm), respectively. In contrast the *in situ* DOU (and TOU) only showed a moderate seasonal variation (from 8 to 30 $mmol\ m^{-2}\ day^{-1}$) due to reduced O_2 availability during periods of carbon enrichment. This led to an accumulation of reduced solids and solutes from anaerobic degradation that first were reoxidized as the O_2 conditions in the bottom water improved, and the labile carbon pool became exhausted. The benthic O_2 uptake for each of the four seasons therefore showed little variation (Figure 20D, Figure 21), even though the seasonality of the carbon mineralization was more extensive as reflected by a larger seasonal variation in the R_{vol} (see above) and the DIC exchange rate as determined in recovered sediment cores (Lomstein & Blackburn 1992).

In polar regions, a single phytoplankton bloom is typically initiated by melting or displacement of the sea-ice cover (e.g. Rao & Platt 1984). This is typically followed by a distinct vertical pulse of organic material (e.g. Rysgaard & Sejr 2007). During a seasonal study in Young Sound, NE Greenland, the event triggered a 2.5-fold increase in the benthic O_2 uptake but the rates declined to a stable background level within a month (Rysgaard et al. 1998). The pulse of labile material reaching the sediment during summer was thus soon consumed, and the benthic O_2 uptake remained at a constant

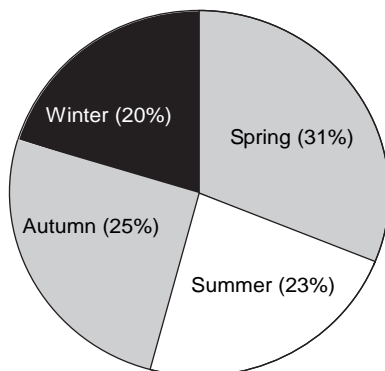


Figure 21. The relative contribution of the four seasons (each of three months) to the annual benthic oxygen uptake at 16 m water depth in central Aarhus Bay (data from Glud et al. 2003).

level of 5–6 $mmol\ m^{-2}\ day^{-1}$ during most of the year (Figure 22A). A more refractory organic carbon pool sustained these background rates. The stimulated benthic O_2 uptake induced by the bloom only

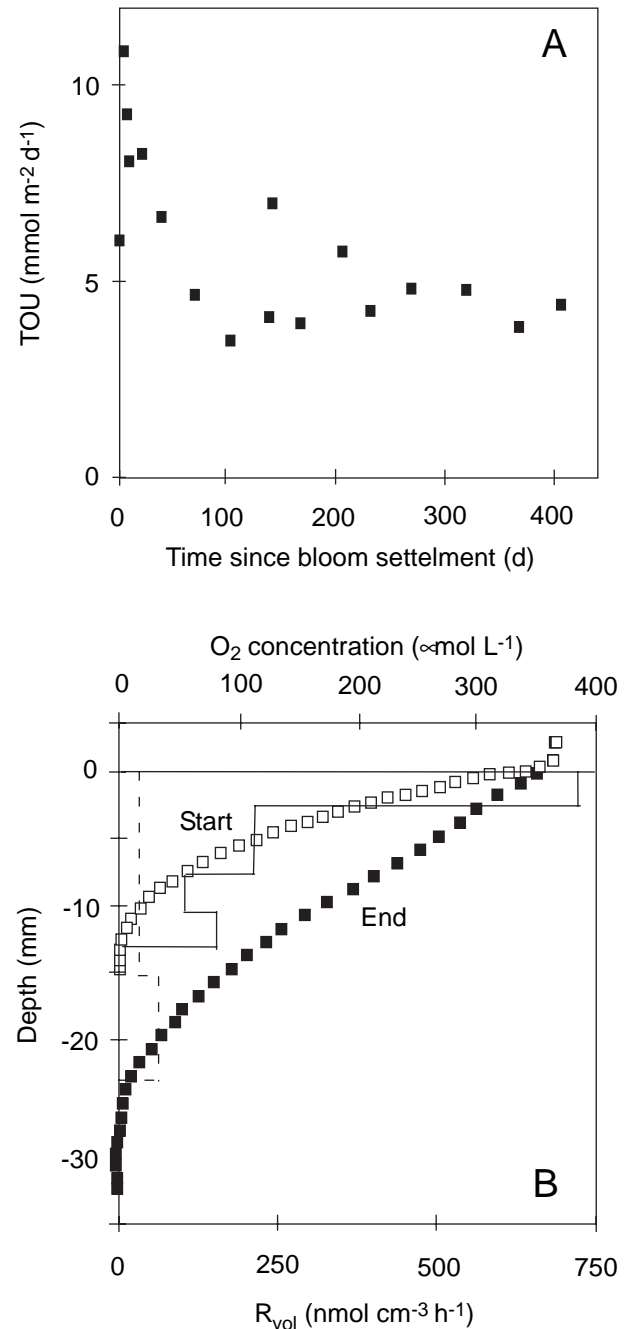


Figure 22. (A) Consecutive measurements of TOU measured in four cores recovered from Young Sound (NE Greenland) at 36 m water depth just after the summer bloom had settled. After the labile organic material was consumed the O_2 uptake remained at 5–6 $mmol\ m^{-2}\ day^{-1}$ for >400 days without addition of any organic carbon. (B) Microprofiles measured just after core recovery and after 400 days in the laboratory. This reflects the gradual downward migration of O_2 as R_{vol} declines along with the pool of organic carbon and reduced inorganic materials (redrawn from Thamdrup et al. 2007).

corresponded to ~8% of the annual activity (Rysgaard et al. 1998), and the dynamic annual response in the benthic O₂ uptake could be modeled assuming two pools of benthic organic material with decay constants of 0.1 and 80 year⁻¹, respectively (Berg et al. 2003b). The results suggest a modest inter-annual and seasonal variability in benthic O₂ uptake in coastal polar settings as a significant part of the heterotrophic activity is fueled by relatively refractory material (Thamdrup et al. 2007).

In general the benthic O₂ uptake of Arctic sediments compares to rates of temperate or even tropical areas (Hulth et al. 1994; Glud et al. 1998a, 2000b and references therein) and the highest biomass densities of macrofauna ever recorded have been found in Arctic regions (e.g. Zenkevitch 1963). Work at the Svalbard archipelago shelf suggests that, by and large, the activity of benthic communities in polar regions is limited by the organic carbon supply rather than the low temperatures (e.g. Kostka 1999; Arnosti & Jørgensen 2003). The long-term carbon burial rates and the relative importance of the respective diagenetic pathways do not seem to be affected by low *in situ* temperature (Glud et al. 1998a; Kostka et al. 1999; Thamdrup et al. 2007). However, despite psychrophilic adaptation of the benthic communities in the Arctic (e.g. Thamdrup & Fleischer 1998), the study in Young Sound indicated a less distinct benthic pelagic coupling due to lower degradation efficiency at the low temperature, i.e. a more outspread release of nutrients and DIC following the settlement of a pelagic bloom as compared to lower latitudes (Thamdrup et al. 2007). In fact it could be speculated that in case the pelagic degradation is impeded by low temperature in seasonally ice-covered areas the importance of benthic mineralization becomes relatively more important due to a higher net deposition. Long-term accumulation of labile organic material at the West Antarctic Peninsula shelf was interpreted as reduced degradation efficiency at low temperatures (Mincks et al. 2005). More work is required to clarify the effect of temperature on benthic diagenetic process rates.

In general, the seasonality of organic carbon production in coastal subtropical and especially tropical areas is less explicit than in temperate or polar regions (Eyre & Balls 1999). The variations in water column production and thus in benthic carbon inputs is mainly driven by changes in wind patterns driving shorter or longer irregular upwelling periods or variations in precipitation regulating local river inputs (e.g. Peterson et al. 1988; Eyre & Ferguson 2005). Such patterns are seasonal, but they can also be erratic and extremely dramatic (Eyre & Ferguson 2006).

The benthic O₂ uptake in coastal and shelf areas expresses a temporal variability related to the input of organic material. Such variability can be logistically difficult to include during *in situ* studies. However, the two examples discussed above suggest that the annual contribution of periods with elevated activity following natural carbon enrichment is minor. Thus proper seasonal representation of benthic O₂ uptake may not be critical for robust O₂ or carbon mineralization budgets in many environments. Episodic resuspension events are probably more important to include in the annual budget. Even short-term resuspension of reduced sediment layers may enhance the O₂ uptake many fold and correspond to a diffusive-mediated O₂ uptake of several weeks (e.g. Jørgensen 1996a). Such events are probably important for reoxidation of reduced surface sediment and in maintaining a high sulfide buffer capacity in coastal environments. The quantitative importance of *in situ* resuspension events for estimates on benthic mineralization ought to receive more attention in the coming years (Tengberg et al. 2003).

For deep-sea environments, it was for a long time anticipated that seasonality in particle flux was significantly dampened before the material reached the sea floor. Seasonal *in situ* studies in the Sargasso Sea supported this (Sayles et al. 1994), and even variable sedimentation of low-reactive organic material will lead to a relatively invariable benthic O₂ demand (Katsev et al. 2006). However, a large body of evidence now documents the fast vertical transport of labile organic material from the photic zone to the deep-sea floor during bloom events (e.g. Billett et al. 1983; Nair et al. 1989). Deep-sea sediments express a similar response to carbon enrichment as coastal sediments and rapidly process labile organic material (Witte et al. 2003b; Moodley et al. 2005) and observations confirm increases in microbial and meiofauna biomass days to weeks after sedimentation events (Graf 1989; Pfannkuche et al. 1999). On longer time-scales, shifts in deep-sea benthic community structure of megafauna has been related to changes in food supply following El Niño and La Niña events (Ruhl & Smith 2004). Deep-sea sediments appear to be more dynamic than previously thought and benthic O₂ uptake in many abyssal areas apparently varies on a seasonal time-scale (Smith & Baldwin 1984; Smith et al. 2001). However, to better understand the quantitative importance of these episodic events for the annual O₂ budget, long-term high-resolution time series are required – data that can only be obtained by long-term deployable transecting vehicles (Smith et al. 1997) or via benthic observatories (Thiel et al. 1994; Pfannkuche & Linke 2003). Such abilities still

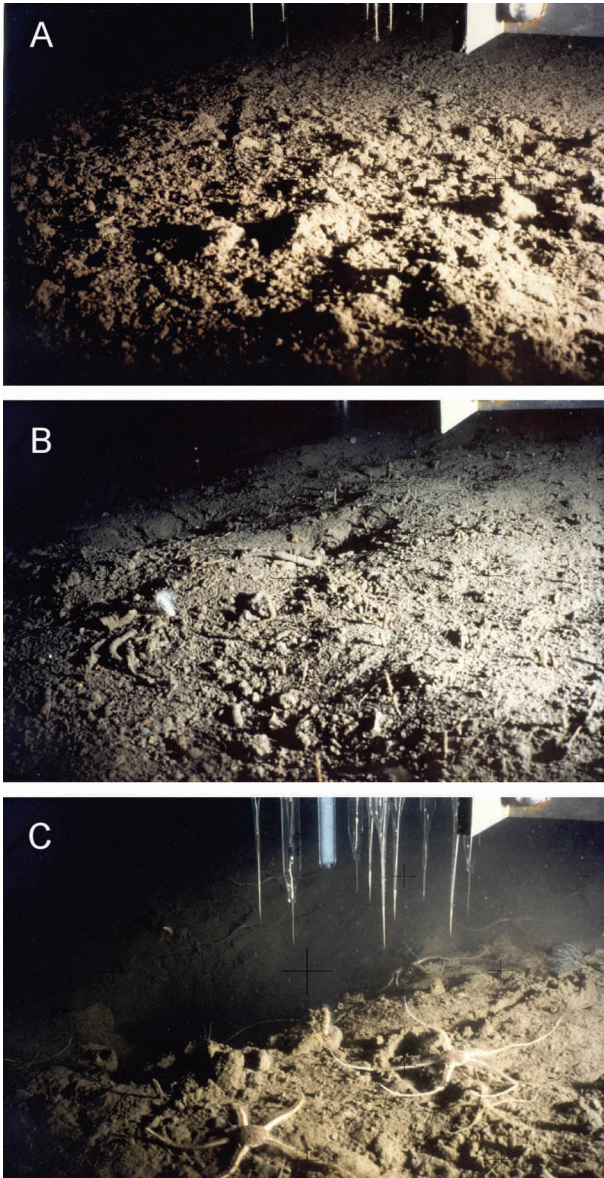


Figure 23. Three *in situ* images taken below an operating profiling lander deployed at water depth between 115 and 175 m around the Svalbard archipelago (photographs by O. Holby and R.N. Glud).

remain to be fully established, but several initiatives for including benthic mineralization as a monitoring parameter on observatories and cable networks have been initiated.

Small-scale spatial variability and microniches

Early laboratory-based O_2 microelectrode measurements documented extensive small-scale variability in microbial mats (e.g. Jørgensen & Revsbech 1983; Jørgensen et al. 1983). For marine sediments, larger funnels, burrow structures and sediment topography have received some attention (see above), but the

consequence of micro patchiness for benthic O_2 distribution has rarely been quantitatively assessed. However, simple visual inspection of the marine sediment surfaces clearly reveals an extensive small-scale heterogeneity; tracks of moving fauna, faecal and detritus deposits, patches of microalgae and small carcasses leave the impression of a varied landscape rather than a homogenous surface (Figure 23).

Microtransects in sediment recovered from water depths >100 m documented a large small-scale variability in the O_2 distribution. Most was ascribed to polychaete activity and burrow structures, but the DOU varied by a factor >2 , within a horizontal distance of a few millimeters even in areas without any visible faunal activity (Jørgensen et al. 2005). Parallel *in situ* measurements also revealed an extensive horizontal variability on the centimeter-scale that complicated area extrapolation of a few point measurements (Glud et al. 1998a). Detailed analysis of *in situ* microelectrode measurements and core incubations from a shallow-water Mediterranean embayment suggested a small-scale spatial variability in O_2 uptake associated with 'organic hot-spots' (Rabouille et al. 2003). *In situ* microprofiles ($n=45$) measured along a 175 m transect in central Sagami Bay (water depth 1450 m) showed that the DOU varied by a factor of >10 (average $2.6 \pm 1.6 \text{ mmol m}^{-2} \text{ day}^{-1}$), while the O_2 penetration depth varied by a factor of 6 (average 3.5 ± 1.3 mm). The variability in DOU and O_2 penetration depth appeared to be associated with a small-scale heterogeneity that was smaller than the distance between two neighboring microelectrodes on the measuring cylinder (~ 3 cm) (Glud et al. 2005). Parallel *in situ* planar O_2 optode measurements ($n=6$) provided additional information on the O_2 distribution and allowed the extraction of ~ 350 vertical microprofiles separated by a horizontal distance down to 1 mm. Detailed statistical analysis of the combined data-set revealed that the O_2 penetration varied at a characteristic patch size of ~ 2 cm (i.e. profiles separated by less than this spatial distance were more similar than profiles separated by larger distances) (Glud et al. 2005). A similar investigation performed in central Øresund, Denmark indicated a characteristic patch size <0.5 cm (Glud et al. 2001). Based on these investigations, variations in the O_2 penetration depth of sediments appear to be structured in small-scale patches with a characteristic size between 5 and 20 mm.

Similar observations were made recently by deploying a transecting microprofiler that resolved 132 microprofiles within an area of 25 cm^2 . Even though the O_2 isolines roughly followed the

sediment relief, the O_2 penetration depth – a proxy for the microbial activity – expressed extreme small-scale variability (Figure 24A). The underlying reason for this pattern remains to be resolved, but it could be associated with characteristic faunal behavior or the size distribution of sedimenting aggregates. Reasonable fits between DOU and O_2 penetration depths as derived from compiled *in situ* O_2 microprofiles have previously been used as a

general indication of: (1) steady-state benthic O_2 distribution; (2) that activity and organic material was evenly distributed in the upper sediment; and (3) that bioirrigation was negligible at least for shelf and slope sediments (Cai and Sayles 1996). The measurements in Sagami Bay suggest otherwise (Figure 24B) and the evidence that the O_2 environment of marine sediments is highly variable and dynamic is increasing.

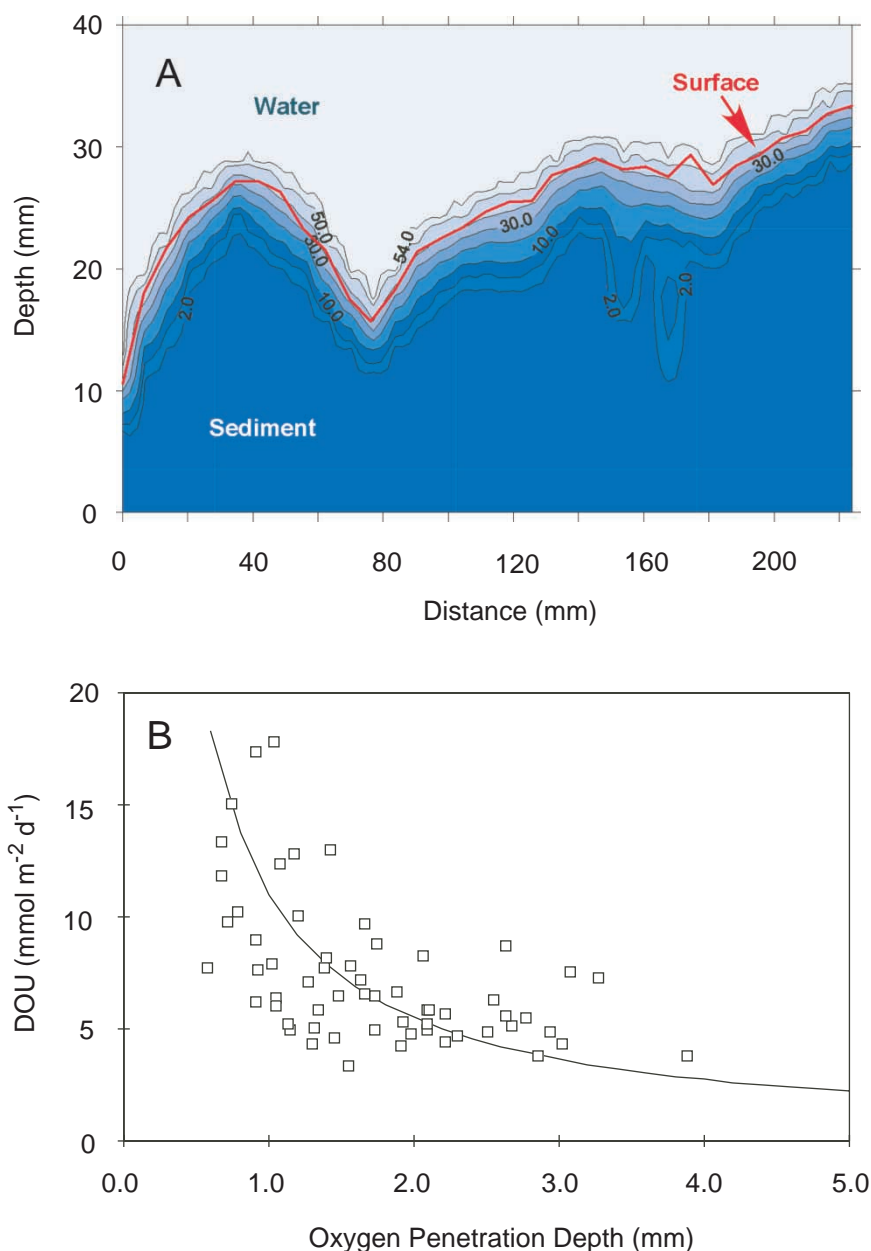


Figure 24. (A) The *in situ* O_2 concentration across the sediment water interface as derived from 33 O_2 microprofiles measured at an equidistant distance of 7.0 mm. The relative position of the sediment surface was derived from a distinct break in the concentration gradients. (B) The DOU depicted as a function of the O_2 penetration depth as derived from individual O_2 microprofiles measured along two microtransects ($n=60$). The solid line indicates $L=2\phi D_s (C_w/DOU)$, which has been proposed as a relation for describing the O_2 penetration (L) as a function of the O_2 concentration (C_w) and the DOU in homogenous sediments at steady state (Cai & Sayles 1996). The equation describes the general trend of the data, but the scatter indicates a substantial heterogeneity in the controls regulating the O_2 distribution.

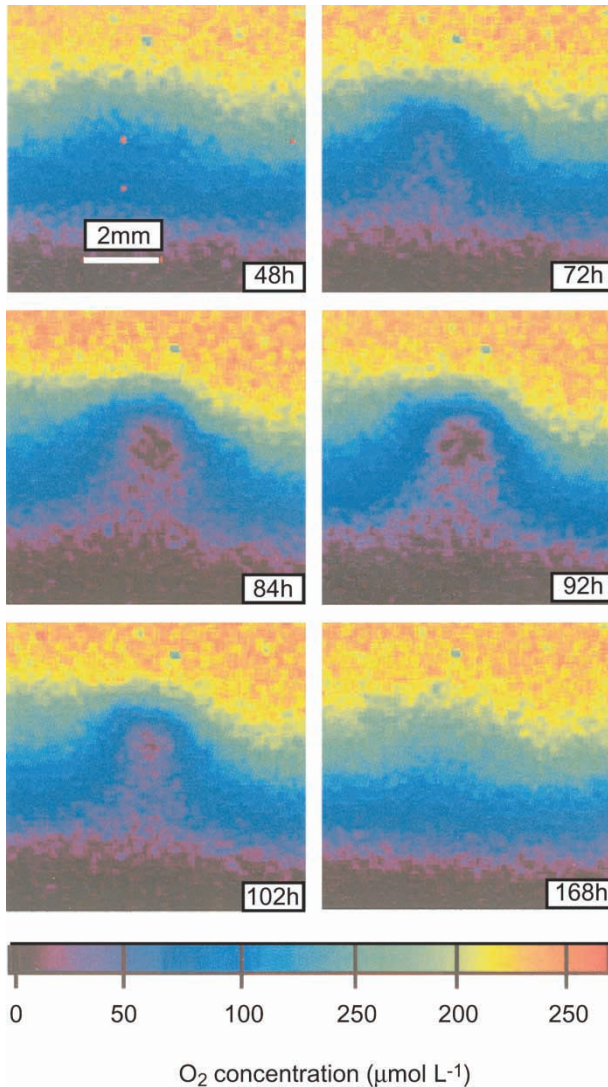


Figure 25. The O_2 dynamic around a 2 mm large skelatonema aggregate placed on a dark-incubated, homogenized sediment. A true anoxic microniche evolves after a lag period as the aggregate gradually decompose (Glud, Nordi, Zang, Sochaczewski & Davidson, manuscript in preparation).

The existence of anoxic microniches in oxic sediments has been repeatedly discussed and postulated for almost half a century in order to explain porewater profiles and organism distributions indicating anaerobic activity in otherwise presumed oxic environments (e.g. Emery & Rittenberg 1956; Jørgensen 1977; Jahnke 1985; Brandes & Devol 1995; Sakita & Kusda 2000). However, whereas reduced microniches in suboxic sediment layers are commonly observed as black enclosures or spots (e.g. Jørgensen 1977), no one has to my knowledge encountered a true anoxic microniche despite measurements of thousands of O_2 microprofiles.

Anoxic microniches have been demonstrated in suspended marine snow aggregates (e.g. Ploug et al. 1997), but it is generally anticipated to be an

ephemeral phenomenon with minor quantitative importance for the pelagic environment (Simon et al. 2002). However, since such aggregates or flocks settle on the sediment, any advective O_2 transport is retarded and the geometry of the diffusive-mediated O_2 supply is changed (e.g. Ploug & Grossart 1999). Theoretical considerations have shown how the potential development of an anoxic microniche depends on the microbial respiration rate in the aggregate, the activity of the surrounding sediment, the aggregate size, the O_2 concentration of the water phase and the O_2 penetration depth of the sediment (e.g. Jørgensen 1977; Jahnke 1985; Brandes & Devol 1995). In areas with relatively shallow O_2 penetration, a potential anoxic microniche will presumably fuse with the zero-isoline of the sediment and merely lead to a change in the topographic relief of the oxic–anoxic interface. Indications of anaerobic processes in the presumed oxic sediment horizon as inferred from porewater profiles (e.g. Brandes & Devol 1995) are in many instances probably the consequence of an area integration of solute concentrations across an oxic–anoxic interface that on the scale of millimeters to centimeters can exhibit an extensive topography (Figure 24). However, planar optodes represent a complementary tool that can provide insights into the microdynamics of O_2 around settling aggregates. Continuous O_2 imaging during a flume study showed that short-lived (<7–8 h) true anoxic microniches developed during the degradation of 1–2 mm large diatom aggregates that settled on recovered sediment (Figure 25). Such events are difficult to capture *in situ* or to study by invasive microelectrode measurements but can be ‘induced’ in front of a planar optode. Planar optode measurements have also documented that larger anoxic niches can develop around degrading *Ulva lactuca* pieces buried in permeable sand with a broad oxic zone induced by advection (Franke et al. 2006). Interpretation of planar optode-derived O_2 images have to account for the fact that measurements are performed along a wall which affects the diffusion geometry (see below) – but experiments using planar optode imaging show that anoxic microniches can be induced in natural sediments and such data will help us to validate and optimize theoretical models allowing us to evaluate the potential importance of anoxic microniches for the benthic diagenesis.

Whereas the existence of anoxic microniches remains controversial, the importance of oxic micro- or mininiches have been documented in a number of studies. Oxic microniches are most commonly induced by faunal activity or burrow flushing (see above), irregular advection of permeable sediments (Cook et al. 2007a) or leakage from plant roots (Caffrey & Kemp 1991). Seagrasses are estimated to

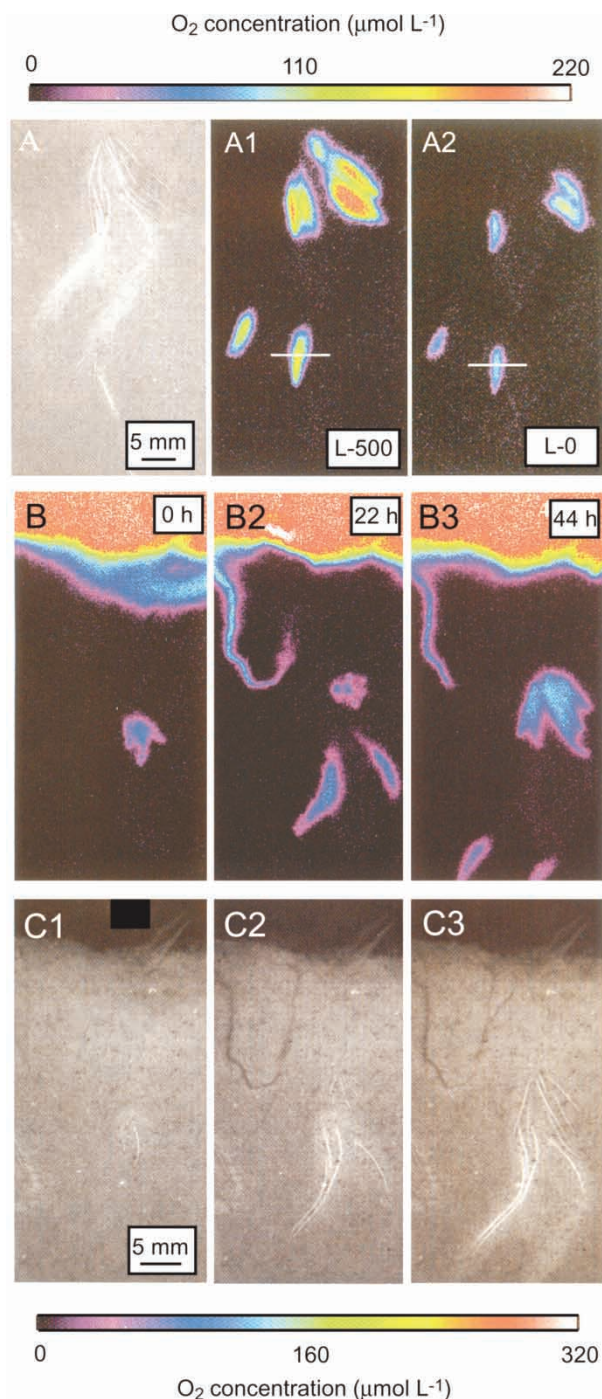


Figure 26. (A) Black and white image of a *Zostera marina* rhizosphere with corresponding O_2 images obtained at $500 \mu\text{mol photons m}^{-2} \text{s}^{-1}$ (A1) and in darkness (A2). (B, C) Series of O_2 images obtained at a 22 h interval (B, B2, B3) with corresponding black and white images (C1, C2, C3) (redrawn from Frederiksen & Glud 2006).

cover $\sim 10\%$ of the global coastal zone and represent an important component in many coastal ecosystems (e.g. Larkum et al. 2006). They are obligate aerobes and thus maintain an efficient O_2 supply to their root system via interconnected gas lacunae. This may

lead to an O_2 loss along the roots and various incubation approaches of entire root systems in artificial media have been applied to estimate the total O_2 loss from seagrass and macrophyte roots (e.g. Sand-Jensen et al. 1982; Caffrey & Kemp 1991). However, such approaches poorly reflect natural conditions and give no insight on where the O_2 leakage occurs.

Microsensor studies have confirmed that some seagrasses develop a barrier toward O_2 permeation along most of the root length (Connell et al. 1999), but only very recently it was documented that for *Zostera marina* only the very root tips – representing less than a few percent of the rhizome area – apparently leak O_2 to the surrounding sediment (Jensen et al. 2005). The radial O_2 loss induced small oxic lenses of $2 \times 6 \text{ mm}$ around the root tips depending on the photosynthetic activity of the plant (Jensen et al. 2005). A parallel planar optode study on the entire rhizosphere growing in intact sediment beds confirmed these findings and documented the extreme dynamic of such microniches (Frederiksen & Glud 2006). Oxygen loss from the root tips depended on the photosynthetic activity of the plant, the O_2 concentration in the overlying water and the root age (Frederiksen & Glud 2006). The roots of the investigated plants grew $\sim 9 \text{ mm day}^{-1}$, and the rhizosphere was thus characterized by a constantly changing mosaic of oxic microniches that developed into reduced sediment leaving an anoxic but oxidized shield around the mature roots (Figure 26). As the roots became older than 7 days they stopped growing and O_2 leakage ceased. For actively growing roots, O_2 leakage only ceased as the O_2 concentration of the overlying water decreased below $70 \mu\text{M}$ (25% air saturation) in darkness – at this point the shielding O_2 niche around the root tips had vanished and potential reduced phytotoxins like H_2S and Fe^{2+} could enter the root system. Extrapolating the findings to a typical seagrass bed with a shoot-density of $\sim 900 \text{ m}^{-2}$, the diurnal O_2 leakage from the root system only corresponded to 10–15% of the O_2 uptake along the primary interface (Jensen et al. 2005; Frederiksen & Glud 2006) and, as such, O_2 leakage from the rhizosphere only played a minor role for the total benthic O_2 consumption rate. Nevertheless, the radial O_2 loss from the root tips might have important biogeochemical consequences locally by increasing the sulfide buffer capacity, but the transient nature of the microniches probably inhibits the development of large densities of aerobic sulfide oxidizing bacteria or nitrifiers. Combined with intense competition for NH_4^+ around active plant roots this explains the apparent lack of coupled nitrification–denitrification in seagrass rhizospheres (e.g. Risgaard-Petersen et al. 1998). It remains to be

shown to what extent O_2 leakage from other seagrasses exhibits similar dynamic as observed for *Zostera marina*. However, preliminary planar optode measurements performed in the rhizosphere of the tropical mangrove tree *Avicennia marina* also showed extensive dynamics where O_2 leakage was regulated by root age, light exposure and ambient O_2 availability as experienced during diel and tidal cycles (R. N. Glud et al., unpublished results).

Experimental considerations of planar optode studies

Planar optode measurements complement traditional microsensor approaches. They overcome the problem of searching 'for the needle in the haystack' and they provide a much better temporal and spatial resolution of the O_2 distribution. Furthermore, they can depict the O_2 dynamics at a given area over many days. As such, they provide new insights into O_2 dynamics around rhizospheres, microbial hot-spots, macro- and meiofauna, structures in permeable sands, and within phototrophic communities (Glud et al. 1999c, 2005; Precht et al. 2004; Wenzhöfer & Glud 2004; Franke et al. 2006; Frederiksen & Glud 2006; Oguri et al. 2006; Polerecky et al. 2006). By combining the technique with inverted periscopes, *in situ* investigations can be performed (Glud et al. 2001, 2005). The number of research groups applying and refining the technique is steadily increasing and other solutes like H^+ , CO_2 and NH_4^+ can now be imaged by similar approaches (e.g. Hulth et al. 2002; Strömberg & Hulth 2005; Zhu et al. 2005, 2006a; Stahl et al. 2007).

A detailed discussion on the limitations and problems of different measuring approaches is beyond the scope of the present review. However, when evaluating planar optode data it is important to realize that the approach has limitations. The main constraint is that the work is being conducted along a wall. In the case of biofilm studies, this can be turned into an advantage as the O_2 dynamics can be studied non-invasively at the base of a biofilm growing directly on the sensor (Glud et al. 1998b). However, in most applications O_2 dynamics are studied around objects or structures positioned along the planar optode wall. Thereby, the sensor acts as an impermeable wall and will distort the three-dimensional O_2 distribution (Polerecky et al. 2006). In case of radial or spherical diffusion geometry the planar optode will provide a skewed image of the O_2 distribution around the structure. Assuming well-defined geometry, simple diffusion models can be used to 'correct' for the presence of an impermeable barrier (Frederiksen & Glud 2006;

Polerecky et al. 2006). However, more elaborate work is in progress quantifying the distortion for various geometries as a function of solute transport and consumption rates. In principle, such models can be used to correct for the impermeable wall, but in most instances that will not be practical – especially during *in situ* work. But the models can be used to bracket the potential distortions. In general, one simply must be aware of the limitations when interpreting two-dimensional O_2 images and deriving quantitative estimates.

Another problem of planar optodes is that a thin water film will stick to the sensor. Along interfaces between regions of advective and diffusive solute transport (i.e. the sediment–water interface, along burrow linings, the interstice of permeable sands) the presence of a wall can lead to distorted concentration gradients. Thus, calculations based on DBL gradients, for example obtained with planar optodes, should be done with caution (Glud et al. 1996b). A different problem is the potential interference with animal behavior. Walls may either attract or detract burrowing infauna dependent upon the sediment type (Dorgan et al. 2005) – and irrigation patterns may change if animals are situated along an impermeable wall that does not consume O_2 or release H_2S .

Specific problems for different planar optode set-ups should be accounted for in order to avoid artefacts. It would be too elaborate to discuss this here, but a general concern is related to light-scattering or -guidance in the various materials between the sensor chemistry and the camera lens (i.e. the support foil and the aquarium or periscope wall; see Franke 2005). This potential problem causes luminescent signals in one part of the image to affect other parts of the image and obtained O_2 images could thus appear smeared. The effect depends on the applied light sources, filter settings, sensors, and materials but can be minimized and maybe even eliminated by proper choice of materials and calibration routines. *In situ* deployments pose a completely different set of problems where changes in local hydrodynamics or solute and particle smearing during periscope insertion potentially can affect the obtained data. Still more work is needed to evaluate the potential problems of using planar optodes, but interferences can often be avoided by proper choice of experimental set-up and – as for many other techniques – avoiding over-interpretation and ignoring inherent caveats. Complementary use of different available techniques often provides the best insight and the best quantitative basis for resolving benthic O_2 dynamics.

Benthic microphytic production and O₂ dynamics of photic sediments

In coastal regions, sunlight may penetrate the water column and reach the sediment surface. The spectral composition of light at the sediment surface differs from sunlight due to the specific absorption spectra of water, dissolved particles and pelagic primary producers, but may still sustain a significant benthic primary production (Cahoon 1999; Gattuso et al. 2006). The maximum water depth for actively growing benthic microalgae rarely exceeds 20–40 m, but active coralline macroalgae have been observed at depths well below 200 m, indicating that the area of benthic primary production can be quite extensive in oligotrophic regions (Steneck 1986).

In extreme environments with reduced metazoan grazing, i.e. hot springs, hypersaline ponds or ice-covered lakes with reduced or no metazoan grazing, the phototrophic communities gradually develop into thick, cyanobacteria-dominated microbial mats bearing resemblance to fossilized Precambrian stromatolitic communities (e.g. Castenholz 1984; Vopel & Hawes 2006). However, in typical coastal settings, benthic microphytes represent an excellent food source, and the fixed carbon is quickly moved up through the different trophic levels (e.g. Middelburg et al. 2000). Nevertheless, even in regions with thriving benthic fauna, the seafloor is often covered by brown or dark green phototrophic communities dominated by diatoms, dinoflagellates or cyanobacteria (Figure 27). The benthic microphytes can take advantage of the nitrogen and phosphorus that is released from the underlying sediment and thereby deprive the pelagic community a nutrient source. In nutrient-limited settings with good light availability, they have a clear advantage over pelagic primary producers. Benthic microalgae compete efficiently with the microbial community for nutrients and their O₂ production has profound effect on other diagenetic processes such as denitrification, nitrification, metal oxidation and sulfide oxidation (Rysgaard et al. 1995; Epping & Jørgensen 1996; Dalsgaard 2003).

The potential importance of the benthic microphytic production in coastal ecosystems has received surprisingly little attention as compared to pelagic production especially in Arctic and tropical regions, and most benthic work in temperate settings has focused on intertidal communities (Cahoon 1999; Middelburg et al. 2005). Recent studies indicate a large importance of subtidal benthic microphytic production for ecosystem carbon cycling (Jahnke et al. 2000; Wenzhöfer et al. 2000; Glud et al. 2002; Dalsgaard 2003; Heil et al. 2004). The relatively few available studies make general conclusions and

extrapolation to larger areas difficult. As an example; to my knowledge at present only five studies have quantified the benthic microphytic photosynthesis of polar regions that host roughly 30% of the global shelf area. To further complicate matters, the few available studies use different measuring techniques such as ¹⁴C incubation, various chamber incubation methods, and microsensor-based methods, that to various degrees each expresses gross or net photosynthetic activity. Studies on benthic primary production have so far been under-appreciated, and more studies on benthic primary production are needed to fully access coastal carbon cycling.

Each of the measuring strategies for accessing benthic primary production has its limitations and problems (e.g. Revsbech et al. 1981). However, microsensor-based measurements offer the possibility to study a given spot in extreme detail and to evaluate factors controlling the photosynthesis and

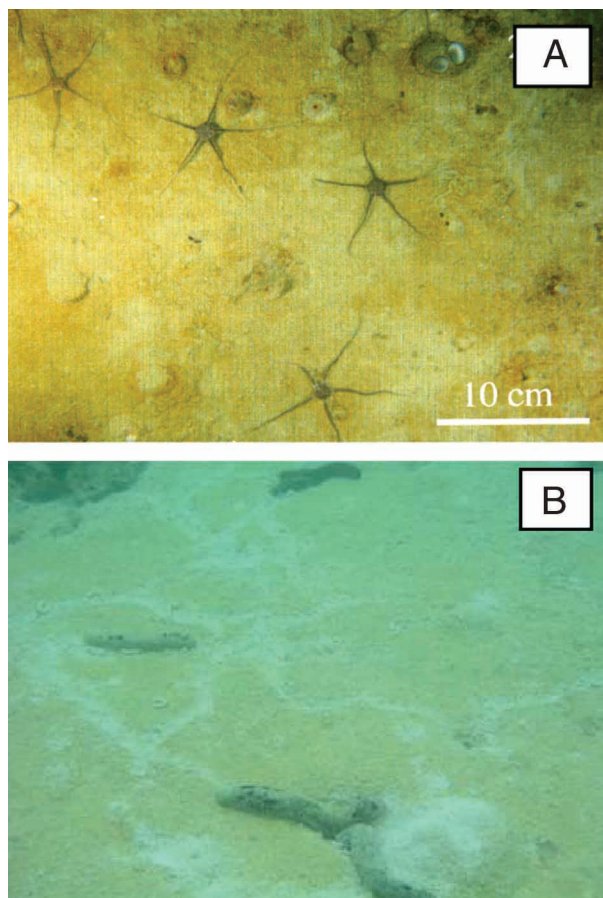


Figure 27. (A) Photograph obtained at 20 m water depth in Young Sound (NE Greenland). The brown patches reflect occurrence of benthic diatoms (photograph by P.B. Christensen and M. Sejr). (B) Photograph taken at 2 m water depth at Heron Island (Australia). The sediment is covered by a layer of benthic dinoflagellates scarred by Holothurians grazing (photograph by N. Patten).

related O_2 -consuming processes (e.g. Revsbech & Jørgensen 1983; Kühl et al. 1996).

O_2 microsensor measurements in benthic microphytic communities

Oxygen microprofiles in light-exposed microphyte communities demonstrate supersaturation just below the sediment–water interface and a relatively large O_2 penetration depth (Figure 28). However, the photic zone is typically less than 1 mm due to the strong light attenuation within the compact communities (Kühl & Jørgensen 1992). The lower boundary for the zone of net O_2 production is typically indicated by a turning-tangent on the concentration profile. From such O_2 concentration profiles, the net photosynthetic activity can be calculated by accounting for the diffusive export of O_2 away from the photic horizon (i.e. the upward flux through the DBL and the downward flux towards the deeper sediment layer) (Figure 28). At light-saturating photosynthesis, the upward flux typically represents >70% of the total O_2 export from the photic zone (Christensen et al. 2000; Wenzhöfer et al. 2000). The depth profile of net O_2 production/consumption can be derived from analysis of the curvature of the

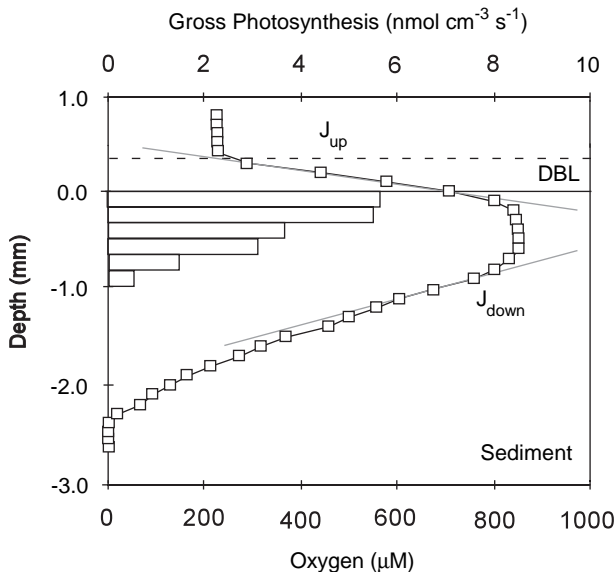


Figure 28. An O_2 microprofile measured in a sediment core recovered from 2 m water depth in Helsingør harbor. The sediment surface was covered by benthic diatoms and was exposed to a downwelling irradiance of $600 \mu\text{mol photon m}^{-2} \text{s}^{-1}$. The upward (J_{up}) and downward flux (J_{down}) amounted to 190 and $63 \text{ mmol m}^{-2} \text{day}^{-1}$, respectively, as calculated from Fick's first law of diffusion. The net photosynthesis was thus $253 \text{ mmol m}^{-2} \text{day}^{-1}$. The depth profile of gross photosynthesis (bars) measured by the light–dark-shift technique equaled $339 \text{ mmol m}^{-2} \text{day}^{-1}$. The respiration of the photic zone was thus $86 \text{ mmol m}^{-2} \text{day}^{-1}$, while the O_2 consumption of the upper millimeters during darkness only was $40 \text{ mmol m}^{-2} \text{day}^{-1}$ (redrawn from Glud 2006).

steady-state O_2 concentration profile as outlined above (Berg et al. 1998) or from transient O_2 measurements at the respective sediment depth during light eclipse (Epping et al. 1999). However, even at moderate light intensities, the O_2 concentration at the sediment surface can become supersaturated and bubble formation can complicate or invalidate flux calculations (Jørgensen et al. 1983).

Depth-integrated benthic net photosynthetic rates increase with light and gradually reach a saturation level but the community rarely expresses photoinhibition (Figure 29). This is due to a gradual expansion of the photic zone with increasing light, and that downward migration allows organisms at the sediment surface to escape inhibiting light conditions. The photosynthetic capacity can be evaluated from plots of photosynthesis versus irradiance (PE relations) and derived key parameters like maximal photosynthetic activity (P_{max}), the light compensation point and the initial increase of the O_2 release (α) often differ markedly among communities (Figure 29). The variability is mainly caused by variations in the adaptation to the local light conditions and the biomass. However, microphytobenthic communities quickly adapt to changing light conditions, and depth transects of *in situ* PE relations reflecting a gradual adaptation to lower light availability merge within hours (or days) when the communities are exposed to the same light conditions (Kühl et al. 2001; Glud et al. 2002).

The depth profile of gross primary production can be determined by the so-called light–dark shift

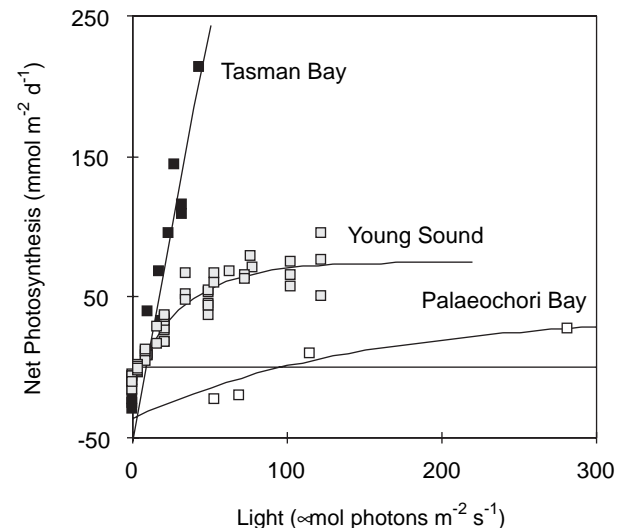


Figure 29. PE relations as measured in three benthic communities all dominated by benthic diatoms. The data-sets were fitted by $P_{\text{net}} = P_{\text{max}}(1 - \exp(-\alpha E_d / P_{\text{max}})) + R$, where E_d is the downwelling irradiance and R expresses the O_2 consumption in darkness (for other constants see text). A few values used for the fitting procedure are off-scale data from Wenzhöfer et al. (2000), Christensen et al. (2003) and Glud et al. (2002).

technique (Revsbech & Jørgensen 1983) (Figure 28). In short, the technique calculates the gross primary production at a given depth from the rate by which O_2 concentration declines after darkening assuming that the rate is equal to the steady-state O_2 production in light. The approach assumes that O_2 consumption and the O_2 concentration gradients are unchanged during darkening and detailed investigations have documented that; in essence this is true as long as one uses the initial rate of O_2 decline (Glud et al. 1992). If longer dark incubations (>1 s) are used, the activity profile will be smeared due to a net diffusion of O_2 down the concentration gradients (Glud et al. 1992; Lassen et al. 1998). When longer dark incubations (i.e. 3–5 s) are applied in communities with steep concentration gradients and high photosynthetic activity the smearing will be extensive and the photosynthetic profiles will only poorly reflect the original activity distribution (Lassen et al. 1998). However, the depth-integrated O_2 decline during darkness remains constant for at least 5 s (i.e. the O_2 consumption rate within the photic zone is constant during this time interval) (Glud et al. 1992).

The depth-integrated gross photosynthetic activity exceeds the net photosynthesis (i.e. the O_2 export from the photic zone). The difference is related to concurrent O_2 consumption within the photic zone. Detailed microsensor investigations have shown that O_2 consumption during light exposure is stimulated from 30 to more than 100% as compared to conditions in darkness (e.g. Glud et al. 1992; Canfield & Des Marais 1993; Kühl et al. 1996; Epping et al. 1999; Fenchel & Glud 2000).

The high gross photosynthetic activity of dense microphytobenthic communities leads to elevated pH and O_2 concentrations, while CO_2 becomes depleted. This is a hostile environment with many O_2 radicals, and conditions are expected to induce intensive photorespiration (i.e. O_2 reduction by RuBisCo) (Glud et al. 1992; Kühl et al. 1996). In such instances, leakage of photosynthates from photorespiration probably serves as an important carbon source for the heterotrophic organisms, and microautoradiography has shown cross-feeding of glycolate from the phototrophic to the heterotrophic community in microbial mats (e.g. Bateson & Ward 1988). Highly active phototrophic communities thereby, to a large extent, recycle the fixed carbon – i.e. photo-excretion stimulates heterotrophic activity which then produces CO_2 and concurrently consumes O_2 . This activity lowers the extreme O_2 levels of the community. Planar optode measurements in microbial mats have documented how light-stimulated photosynthesis enhances net O_2 consumption in neighboring cell clusters especially

during saturating light conditions – presumably as a consequence of photo-excretion (Glud et al. 1999c). At night, the accumulated labile organic material may lead to intensified heterotrophic activity that gradually declines as the labile carbon pool gradually becomes exhausted (Fenchel & Glud 2000; Tang & Kristensen 2007).

Light-stimulated O_2 consumption is therefore mainly ascribed to elevated heterotrophic activity and photorespiration. However, intensified oxidation of reduced inorganic constituents that have accumulated in darkness can also be of quantitative importance. With a complementary microsensor-core-incubation study a budget for the 162 mmol $m^{-2} day^{-1}$ of O_2 produced during 12 h light by microphytobenthic photosynthesis was established. The stimulated heterotrophic activity/photorespiration was by far the dominant O_2 -consuming process, while export and oxidation processes were minor O_2 sinks (Figure 30). During the night, the mat went almost anoxic, and the relatively low O_2 consumption rate (44 mmol $m^{-2} day^{-1}$) was presumably dominated by oxidation of H_2S produced by the sulfate-reducing bacteria (Fenchel & Glud 2000). Even though it is a common procedure, it cannot be recommended to estimate O_2 consumption rates during light exposure from dark profiles. Similarly for chamber experiments, dark-incubated chambers do not express the O_2 consumption rate in light and, furthermore, net rates in light and darkness compare the activity of different depth intervals as the O_2

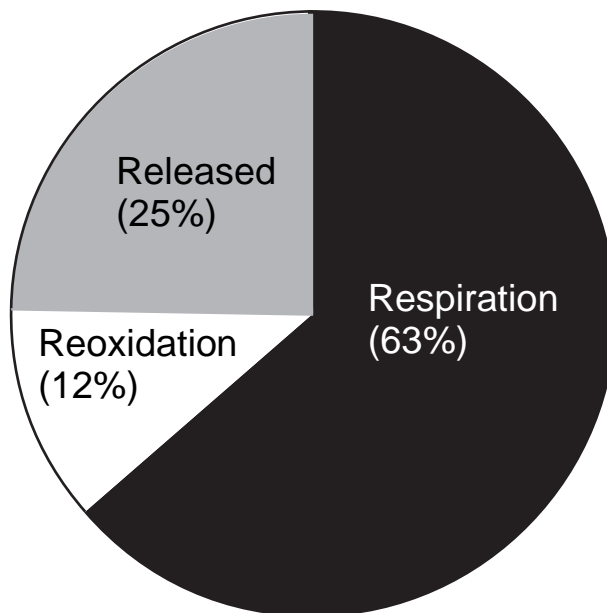


Figure 30. The relative importance of processes responsible for consuming the O_2 produced during 12 h light exposure of a microphytobenthic community. The 'respiration' term includes photorespiration (data from Fenchel & Glud 2000).

penetration depth differs between the two situations (Epping & Jørgensen 1996).

In cohesive phototrophic communities, the DBL thickness affects to what extent the produced O_2 is consumed or exported (Kühl et al. 1996; Garcia-Pichel et al. 1999; Larkum et al. 2003). By increasing the free flow velocity, the DBL thickness will decrease, and the upward O_2 flux will increase. To the extent the gross photosynthesis is constant, the downward flux and the O_2 consumption rate will decrease correspondingly (Figure 31). If gross photosynthetic rates are limited by DIC availability, a reduction in the DBL thickness could be expected to increase the gross photosynthesis.

Temperature is another important factor controlling benthic photosynthesis and coupled O_2 consumption (Wieland & Kühl 2000; Hancke & Glud 2004). Measurements in diatom-covered sediments expressed a gross photosynthetic Q_{10} response of 2.2–2.6 while the O_2 -consuming processes had Q_{10} values of 2.6–5.6 in the temperature range from 0 to $\sim 20^\circ\text{C}$. The sediment inhabited by the benthic microphytes and exposed to constant light conditions thus gradually turned net heterotrophic with increasing temperature (Hancke & Glud 2004). Whether this pattern only represents a short-term (diurnal) temperature response or also is induced by seasonal temperature changes still remains to be investigated. Salinity appears to be of minor importance as an environmental control for benthic photosynthesis (Garcia-Pichel et al. 1999).

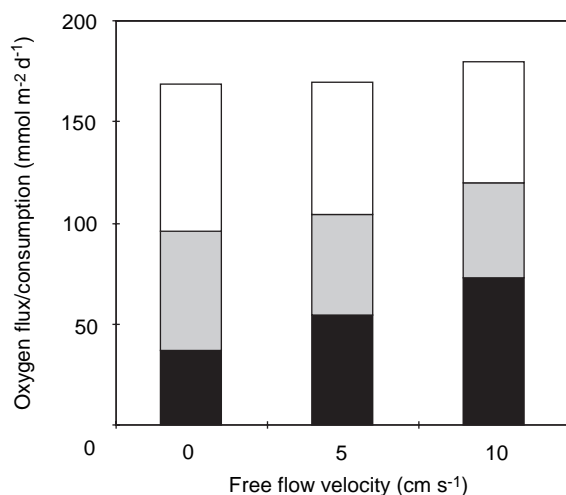


Figure 31. The upward (black) and downward (grey) O_2 fluxes along with the inherent O_2 consumption rate of the photic zone (white) as measured in a cyanobacteria-dominated biofilm exposed to different flow velocities. With increasing flow the upward O_2 flux increases while the downward flux and respiration decreases. The total O_2 production increases slightly with the flow presumably as a consequence of a higher DIC supply from the overlying water phase (data from Kühl et al. 1996).

As shown above, O_2 microsensors can provide very detailed insights into the O_2 dynamics of microphytobenthic communities including rates of O_2 consumption and net and gross photosynthetic rates. However, the main problem in using such measurements for estimating benthic primary production is to extrapolate the findings beyond the investigated spot. Marine phototrophic sediments are extremely heterogeneous, and the resolved O_2 exchange rates and O_2 penetration depths vary extensively even within a few centimeters. It takes at least 30–40 min to obtain one data-set of net and gross photosynthesis which sets a limit to how many parallel measurements can be obtained, even when using sensor arrays (Holst et al. 1997). Planar optodes can complement traditional microsensor approaches, but the limitations of working along a wall and potential smearing during sensor insertion makes this approach less suited for accessing *in situ* benthic productivity.

The fluorescence emitted by the photosynthetic apparatus has been used as a proxy for biomass and photosynthetic activity of benthic microphytes (Schreiber et al. 1994; Hartig et al. 1998; Barranguet & Kromkamp 2000; Kühl et al. 2001; Glud et al. 2002). Using a diver-operated underwater Pulse Amplitude Modulated fluorometer (Diving PAM) that applies the 'pulse saturation technique', two relatively robust proxies for microphytobenthic biomass and photosynthesis can be measured *in situ* within a few minutes: the minimal fluorescence, F_0 , and the electron transport rate between photosystem II and I, ETR (Schreiber et al. 1986; Kühl et al. 2001). Empirical relationships between ETR and net photosynthesis derived from microprofiles have been established facilitating area extrapolation of microsensor measurements (Glud et al. 2002). Such relationships are far from universal, and frequent intercalibration during a study is required to obtain trustworthy results, but the technique may overcome the prime limitations of using microsensors for quantification of benthic primary production. PAM imaging applied either vertically across the sediment surface *via* inverted periscopes or horizontally above the sediment surface represents a promising approach to resolve and overcome small-scale variability in microphytobenthic distribution and activity (Grunwald & Kühl 2004).

A study in Young Sound NE Greenland showed that benthic diatoms covered the sediment surface at water depths < 30 m (with a maximum biomass at 20 m water depth; Glud et al. 2002). The area hosting benthic productivity was equivalent to 15.1 km² or 12% of the outer sound. By using a microsensor-derived PE relationship for net photosynthesis and accounting for light conditions and the

diatom coverage at the respective depths, the net benthic photosynthetic rate for the sound was quantified. The estimated integrated net photosynthetic rate during the 80-day sea-ice-free period amounted to 311 tons C (Glud et al. 2002). For comparison, the gross primary production of the pelagic community for the same area and period estimated from short-term ^{14}C incubations amounted to 113 tons C (Rysgaard et al. 1999; Glud et al. 2002). Thus, benthic primary production completely dominated the ecosystem production at water depth above 30 m (Figure 32). Accounting for the entire sound, with an average water depth of 100 m, benthic microphytic net photosynthesis still amounted to 27% of the gross primary production of the water column. Even though the contribution from vascular macroalgae dominated, the budget underlines the potential importance of benthic microphytes for coastal ecosystem function and the need for further studies evaluating and quantifying benthic photosynthetic activity (Krause-Jensen et al. 2007).

Most assessments of microphytobenthic primary production have been done by ^{14}C -labeled DIC incubations or from chamber-derived benthic O_2 exchange rates. The main problem with ^{14}C incubations is estimating the specific activity of the DIC available for the phototrophic community during the incubation. This is complicated due to the extreme temporal and spatial variability and the steep DIC

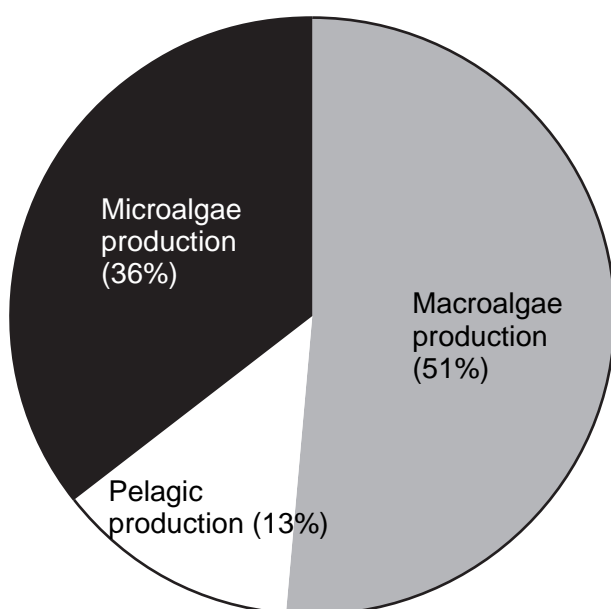


Figure 32. The relative contribution of the respective primary producers to system production integrating areas having <30 m water depth in Young Sound (NE Greenland). The pelagic production expresses gross rate while the benthic compartments are net rates. An insignificant contribution from sea-ice algae has been ignored (data from Glud et al. 2002).

concentration gradient of intact communities. Slurry approaches that evenly distribute the tracer have been applied, but these drastically change the microenvironmental controls that regulate the activity. It is therefore difficult to quantify *in situ* benthic production from ^{14}C incubations (Revsbech et al. 1981). Net photosynthesis calculated from chamber-derived total exchange rates of O_2 are confounded by fauna-related O_2 consumption and exchange across enclosed patches of bare sediment without microphyte coverage (Glud et al. 2002). None of the generally applied techniques are ideal, and to obtain trustworthy estimates, a combination of several complementary approaches is recommended.

Primary production in permeable sediments

Chamber incubations have shown that permeable sand holds a large potential for benthic primary production (Jahnke et al. 2000). Enhanced percolation as induced by increased stirring in cylindrical chambers (Janssen et al. 2005) often stimulates the O_2 release rate of permeable sediments hosting microphytobenthic communities (Figure 33). A recent study in sand demonstrated that percolation lowered the interstitial pH and overcame the diffusion-limited CO_2 supply to benthic microphytes. Thereby, advection stimulated the depth-integrated benthic photosynthesis by 20% (Cook & Røy 2006).

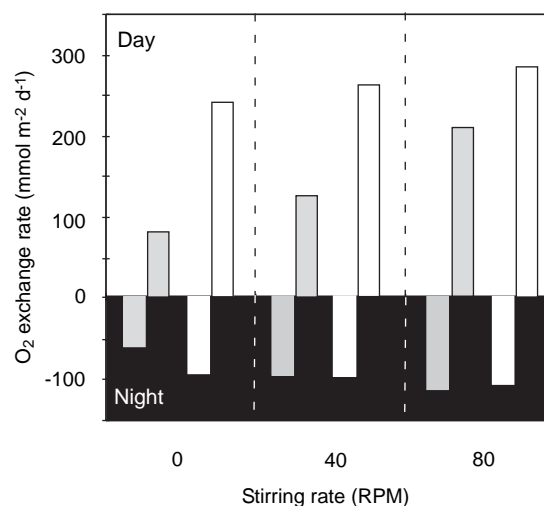


Figure 33. The O_2 exchange rates measured *in situ* by circular chambers kept at different stirring/percolation rates. Measurements were performed at day and night and before (grey bars) and after (white bars) a massive bloom of dinoflagellates had covered the sand surface (Figure 27B). Enhanced percolation clearly stimulated the benthic primary production and O_2 consumption at night time prior to the bloom. The bloom lowered the permeability of the sediment and thus the imposed pressure gradient did not induce percolation and the O_2 exchange rates became less sensitive to stirring. 0 RPM mean that the overlying water was gently mixed without inducing a partial pressure gradient (redrawn from Glud et al. 2008).

In permeable sediments, the enhanced solute transport may thus lower the photosynthetic stress (i.e. the O_2 /DIC ratio) and overcome the diffusion-limited DIC supply experienced by cohesive microphytic communities. The very few available O_2 release rates of light-exposed sandy sediments are comparable to rates from cohesive sediments. This underlines the potential importance of primary production in coastal settings hosting permeable sand (Jahnke et al. 2000).

Unless grazing or hydrodynamic forces keep the microphytobenthic biomass low, the sediment will gradually plug up, and advection will be reduced and eventually cease. A shift from an advective to a diffusive-dominated benthic system was observed as coral mass-spawning products were advected into carbonate sand at Heron Island, Australia. This induced a benthic bloom of dinoflagellates that took advantage of the released nutrients. Their coverage plugged up the sand until intensified grazing had re-set the system after 8–10 days (Figure 33).

Concluding remarks

As described above, O_2 exchange measurements represent a widely applied proxy for quantifying the benthic carbon mineralization and the primary production. During the last decades, much has been learned about benthic O_2 dynamics. The introduction of *in situ* technology and O_2 microsensors to aquatic biology represented milestones in this. Microsensors provided new and much better insights into benthic O_2 dynamics, and new questions in relation to cycling of organic material in sediment were raised and addressed. This is a good example of how science benefits and progresses from the introduction of new technologies. However, there still exist many open questions that will require our attention in the coming years and many environments like permeable sands, the polar regions, the open oceans and the deep trenches are grossly under-studied. New exciting technologies like ‘eddy correlation’ and ‘planar optodes’ are now providing better insights, but also allow us to raise new fundamental questions about the function and activity of marine sediments; processes and benthic systems that previously were difficult to address quantitatively are now within our reach. The recent developments of benthic observatories and cable networks represent major advances. There is a huge potential in linking *in situ* operating systems for monitoring benthic O_2 dynamics to such platforms, allowing a much better understanding of the temporal and spatial variability in benthic O_2 dynamics and carbon cycling.

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