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2 Non-target effects of the insecticide methoprene on molting in the 3 estuarine crustacean *Neomysis integer* (Crustacea: Mysidacea)

4 An Ghekiere^{a,*}, Tim Verslycke^b, Nancy Fockedey^c, Colin R. Janssen^a

5 ^a Laboratory of Environmental Toxicology and Aquatic Ecology, Ghent University, J. Plateaustraat 22, B-9000 Ghent, Belgium

6 ^b Biology Department MS#32, Woods Hole Oceanographic Institution, Woods Hole, MA 02543, USA

7 ^c Marine Biology Section, Ghent University, Krijgslaan 281/S8, B-9000 Gent, Belgium

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9

10 Abstract

11 Ecdysteroids, the molting hormones in crustaceans and other arthropods, play a crucial role in the control of growth,
12 reproduction and embryogenesis of these organisms. Insecticides are often designed to target specific endocrine-regulated functions
13 such as molting and larval development such as methoprene, a juvenile hormone analogue.

14 The aim of this study was to examine the effects of methoprene on molting in a non-target species, the estuarine mysid
15 *Neomysis integer* (Crustacea: Mysidacea). Mysids have been proposed as standard test organisms for evaluating the endocrine
16 disruptive effect of chemicals. Juveniles (<24 h) were exposed for 3 weeks to the nominal concentrations 0.01, 1 and 100 µg
17 methoprene/l. Daily, present molts were checked and stored in 4% formaldehyde for subsequent growth measurements. Metho-
18 prene significantly delayed molting at 100 µg/l by decreasing the growth rate and increasing the intermolt period. This resulted in a
19 decreased wet weight of the organism. The anti-ecdysteroidal properties of methoprene on mysid molting were also evaluated by
20 determining the ability of exogenously administered 20-hydroxyecdysone, the active ecdysteroid in crustaceans, to protect against
21 the observed methoprene effects. Co-exposure to 20-hydroxyecdysone did not mitigate methoprene effects on mysid molting. This
22 study demonstrates the need for incorporating invertebrate-specific hormone-regulated endpoints in regulatory screening and
23 testing programs for the detection of endocrine disruption caused by man-made chemicals.

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25 **Keywords:** Endocrine disruption; Juvenile hormone analog; Molting; Methoprene; *Neomysis integer*

26

27 1. Introduction

28 It is increasingly recognized that for the assessment of
29 the ecological impact of potential endocrine disruptors
30 relevant hormonal mechanisms for both invertebrates
31 and vertebrates need to be considered. Invertebrates
32 account for roughly 95% of all animals (Barnes, 1980),
33 yet surprisingly little effort has been invested to under-
34 stand their value in signaling potential environmental

endocrine disruption. Since the hormones produced and 35
used in invertebrates are not directly analogous to those 36
of vertebrates, it is essential to incorporate invertebrate- 37
specific hormone-regulated endpoints in studies aimed 38
at evaluating potential endocrine disruption. 39

Mysid crustaceans have been traditionally used in 40
standard marine/estuarine toxicity testing because of 41
their ecological importance, wide geographic distribu- 42
tion, year-round availability in the field, ease of trans- 43
portation, ability to be cultured in the laboratory, and 44
sensitivity to contaminants. In addition, mysids have 45
been proposed as potential test organisms for the regu- 46

* Corresponding author. Tel.: +32 92643707; fax: +32 92643766.
E-mail address: an.ghekiere@UGent.be (A. Ghekiere).

47 latory screening and testing of endocrine disruptors by
48 several agencies such as USEPA, OECD and the Ministry
49 of the Environment of Japan (Verslycke et al.,
50 2004a,b).

51 Molting is regulated by a multihormonal system, but
52 is under the immediate control of molt-promoting steroid
53 hormones, the ecdysteroids, secreted by the Y-
54 organ (Fig. 1). The Y-organ secretes ecdysone,
55 which upon release in the hemolymph, is converted
56 into active 20-hydroxyecdysone. Ecdysteroids also
57 play a fundamental role in the control of reproduction
58 and embryogenesis (Subramoniam, 2000). One major
59 advantage of using ecdysteroid metabolism as an end-
60 point is that it provides a means of evaluating the impact
61 of environmental chemicals on crustaceans (and poten-
62 tially other arthropods); chemicals which may not nec-
63 essarily affect vertebrates (Verslycke et al., 2004a,b).
64 Juvenile hormones regulate metamorphosis and repro-
65 duction in insects. With the discovery of the chemical
66 structure of insect juvenile hormone in 1967 (Roller et
67 al., 1967), attempts were made to produce synthetic
68 analogs for use as “third generation” insecticides (Wil-
69 liams, 1956). Methoprene is such an insecticide which
70 acts as a juvenile hormone analog and disrupts normal
71 development of insects by inhibiting developing pupae
72 from molting and passing into the adult stage. Metho-
73 prene is one of the most widely used and successful
74 insect growth regulators. One of the main applications

75 of methoprene is mosquito control. Methoprene can
76 enter estuarine environments by either direct application
77 for controlling aquatic-borne pests (such as mosquitoes)
78 or indirectly through land-drainage or erosion from
79 adjacent pesticide-treated agricultural lands (Dhadialla
80 et al., 1998; Retnakaran et al., 1985). Methoprene
81 degrades rapidly in sunlight (Quistad et al., 1975) and
82 in water (Schaefer and Dupras, 1973). Methoprene may
83 have broken down during the bioassay, but methoprene
84 breakdown products are also known to be bioactive
85 (Harmon et al., 1995; LaClair et al., 1998). It was
86 beyond the scope of this study to determine whether
87 the effects observed were mediated by methoprene itself
88 or by its breakdown products such as methoprenic acid.
89 The use of methoprene at recommended application
90 rates is expected to result in environmental concentra-
91 tions of ~10 µg/l (Ingersoll et al., 1999). Methoprene
92 concentrations in natural water of the US ranged from
93 0.39 to 8.8 µg/l (Knuth, 1989), which is in the concen-
94 tration range where laboratory effects were observed on
95 endocrine regulated processes in crustaceans (McKen-
96 ney and Celestial, 1996; McKenney and Matthews,
97 1990; Peterson et al., 2001). However, US-EPA has
98 not reported any specific ecological effects indicating
99 a significant risk associated with methoprene and con-
100 sider methoprene use (US EPA, 2001).

101 Similarities between the endocrinology of molting in
102 crustaceans and insects led to the discovery of a crusta-

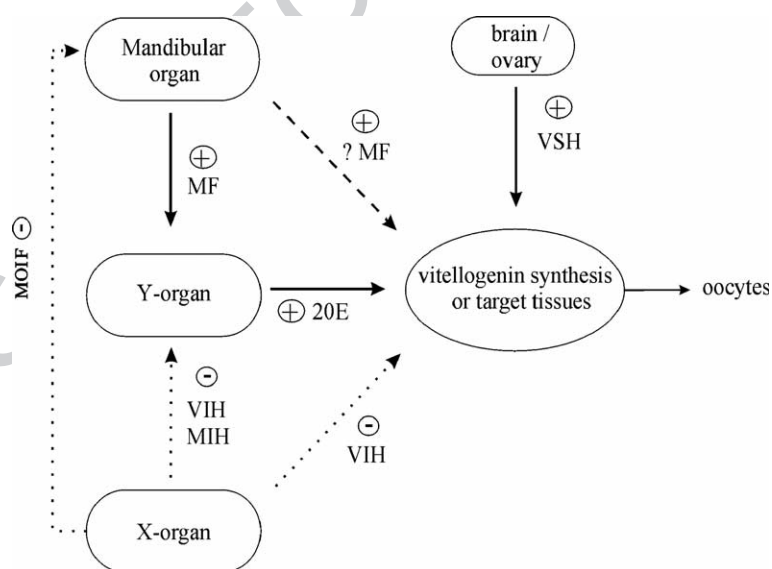


Fig. 1. Schematic representation of the hormonal control of the crustacean molt cycle and vitellogenesis. Adapted from Defur et al., 1999; Meusy and Payen, 1988; Oberdörster and Cheek, 2000. Interrupted arrows (–) represent inhibition and full arrows (+) stimulation. The following hormones play an important role in regulating crustacean molting and vitellogenesis: 20E, 20-hydroxyecdysone, the active molting hormone; MF, methyl farnesoate; MOIF, mandibular organ-inhibiting factor; VIH, vitellogenesis-inhibiting hormone; MIH, molt-inhibiting hormone; VSH, vitellogenesis-stimulating hormone.

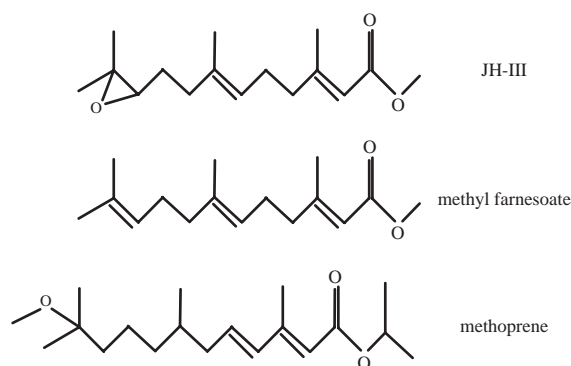


Fig. 2. Chemical structures of juvenile hormone III (JH-III) present in insects, methyl farnesoate in crustaceans and the juvenile hormone analog methoprene.

103 cean analog (methyl farnesoate, the unepoxidated form
104 of juvenile hormone III) to the insect juvenile hormone.
105 Fig. 2 represents the chemical structures of juvenile
106 hormone III, methyl farnesoate and methoprene.

107 We previously developed assays to evaluate chemi-
108 cal effects on steroid and energy metabolism in *Neo-*
109 *mysis integer* (Verslycke et al., 2002; Verslycke and
110 Janssen, 2002). The purpose of this research is to
111 evaluate molting of *N. integer* as invertebrate-specific
112 endpoint. Herefore, we exposed *N. integer* to the test
113 compound methoprene. Methoprene has been shown to
114 reduce mysid fecundity (McKenney and Celestial,
115 1996), interfere with juvenile crustacean development
116 (Celestial and McKenney, 1994; McKenney and Mat-
117 thews, 1990; Olmstead and LeBlanc, 2001; Templeton
118 and Laufer, 1983) and act as an anti-ecdysteroid in
119 daphnids (Mu and LeBlanc, 2004). In a recent study,
120 Mu and LeBlanc (2004) demonstrated that juvenile
121 hormones – and their chemical analogues – interfere
122 with normal ecdysteroid signaling in daphnids, proba-
123 bly via a receptor-based process. Although other crus-
124 taceans most likely have similar cross talk between
125 juvenoid and ecdysteroid signaling pathways, this has
126 not been studied yet in mysids. We performed a co-
127 exposure using the juvenile hormone analog metho-
128 prene and the active ecdysteroid 20-hydroxyecdysone.

129 2. Materials and methods

130 2.1. Chemicals

131 Methoprene (CAS # 40596-69-8) and 20-hydro-
132 xyecdysone were obtained from Sigma-Aldrich (Bor-
133 nem, Belgium). Stock solutions of methoprene and 20-
134 hydroxyecdysone were prepared in absolute ethanol
135 and stored in a dark refrigerator. The ethanol concen-

136 tration in the solvent control and in the different test
137 concentrations was 0.01%.

138 2.2. Test organisms

139 The mysid crustacean, *N. integer*, was collected by
140 handnet in the Braakman, a brackish water (10 psu) near
141 the Schelde estuary in Hoek (The Netherlands). After a
142 24-h acclimation period to the maintenance temperature,
143 the organisms were transferred to 200-l glass aquaria.
144 Culture medium was artificial seawater (Instant
145 Ocean®, Aquarium Systems, France), diluted with aer-
146 ated deionized tap water to a final salinity of 5 psu. A
147 14 h light:10 h dark photoperiod was used during
148 culturing and water temperature was maintained at 15
149 °C. Cultures were fed daily with 24–48 h-old *Artemia*
150 *nauplii* ad libitum. Hatching of the *Artemia* cysts was
151 performed in 1-l cylinder-conical vessels under vigorous
152 aeration and continuous illumination at 25 °C.

153 2.3. Chronic toxicity test

154 Gravid females were collected from the culture and
155 individually transferred to aquaria. The aquaria were
156 examined daily for newly released juveniles. Juveniles
157 <24 h old were placed individually in 80 ml glass
158 recipients containing 50 ml of the desired test concen-
159 tration at a salinity of 5 psu and a temperature of 15 °C.
160 The juveniles were randomly distributed between the
161 different test vessels containing 0–0.01–1–100 µg
162 methoprene/l and 100 µg methoprene/l+0, 24, 77,
163 240 mg/l 20-hydroxyecdysone (=0, 0.05, 0.16, 0.5 µM
164 20-hydroxyecdysone). These concentrations are based
165 on previous studies with *N. integer* (Verslycke et al.,
166 2004a,b) and *Daphnia magna* (Mu and LeBlanc,
167 2002). All concentrations reported in this study are
168 nominal, based on dilutions of the stock solutions.
169 Exposure lasted 5 molts (~3 weeks) and 15 replicates
170 per concentration were used. Exposure solutions were
171 renewed every 48 h and juveniles were fed daily with
172 24- to 48-h old *Artemia* nauplii ad libitum. Daily, dead
173 food was removed and molts were stored in 4% form-
174 aldehyde for subsequent growth measurements.

175 2.4. Growth and molting

176 Toxicological endpoints include time (days) between
177 two successive molts (intermolt period; IMP), and
178 length increases (growth rate, µm/day) during IMPs.
179 The standard length of *N. integer* or the distance from
180 the base of the eyestalks to the posterior end of the
181 last abdominal segment (Fig. 3) cannot be measured

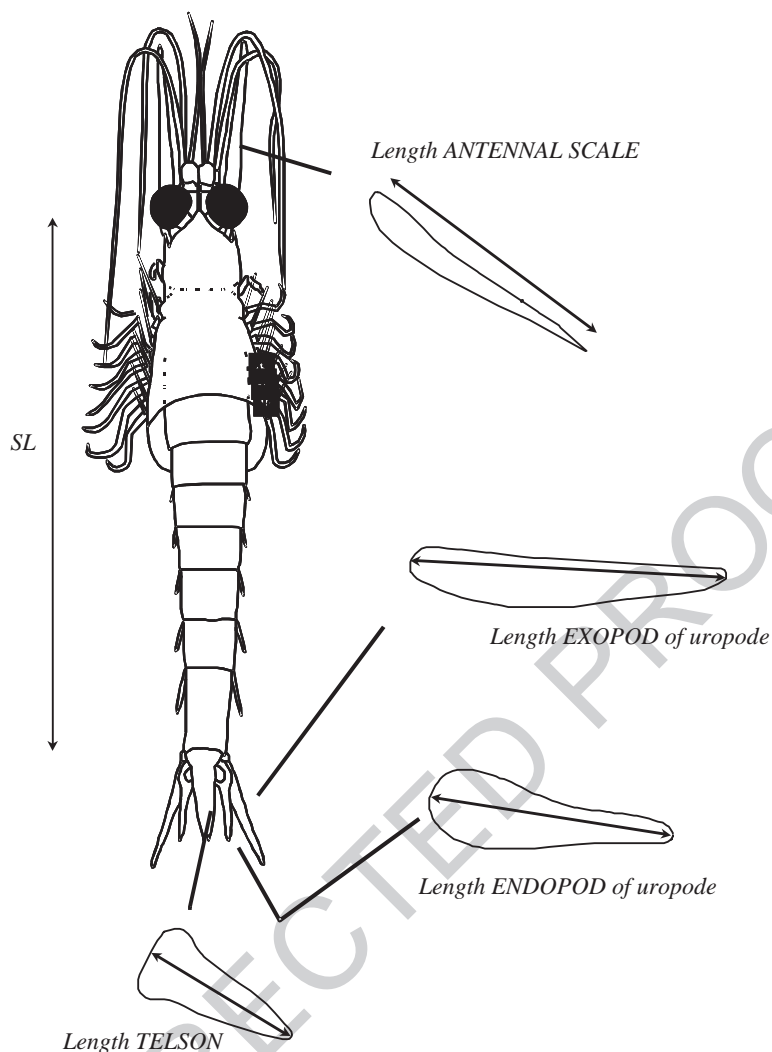


Fig. 3. Schematic representation of *Neomysis integer* with indication of the rigid parts of the molts measured in order to calculate the standard length (SL): length of antennal scale, length of endopod and exopod of the uropod and telson length.

182 directly on the exuvia since the molt is too fragile and
 183 easily breaks during manipulation. Therefore, well-de-
 184 fined rigid parts of the molts were measured using
 185 conventional light microscopy (Fig. 3). Preferably, the
 186 length of the exopodites of the uropod (EXO) were used.
 187 The standard length (SL) can subsequently be calculated
 188 from the exopodite length (EXO) using the linear re-
 189 gression: $SL \text{ (mm)} = 1.085566 + 4.081793 * EXO \text{ (mm)}$;
 190 $R^2 = 0.9569$, $n = 97$ (Fockedeij et al., in press).

191 2.5. Statistics

192 All data were checked for normality and homogene-
 193 ity of variance using Kolmogorov–Smirnov and Leve-
 194 ne’s test respectively, with an $\alpha = 0.05$. The effect of the
 195 treatment was tested for significance using a one-way

analysis of variance (Dunnett’s test; Statistica™, Stat-
 soft, Tulsa, OK, USA). All box-plots were created with
 Statistica™ and show the mean (small square), standard
 error (box), and the standard deviation (whisker).

3. Results

In a preliminary study, we exposed subadults (aver-
 age length 7 mm) to the test compound methoprene
 (0.01, 1, 100 $\mu\text{g/l}$) over the course of 5 molts (data not
 shown). Because of the high individual variability in
 mysid subadult intermolt period (IMP) and growth rate
 (GR), we decided to work with freshly released juve-
 niles (<24 h) to minimize individual variability. The
 duration of the first intermolt stage was equal for all
 animals of the same brood and occurred 3 to 4 days

210 after release from the marsupium (Fockedeey et al., in
 211 press). Animals of the same brood were randomly
 212 distributed over the different exposure treatments
 213 which significantly decreased the individual variability
 214 of the IMP and GR as compared to the preliminary
 215 study with subadults.

216 3.1. Effect of methoprene on mysid intermolt period
 217 (IMP)

218 Fig. 4A shows the effect of methoprene on the IMP
 219 during five successive molts. Generally, the growth of

N. integer is characterized by successively increasing
 220 IMPs (Fockedeey et al., in press). In the controls, the
 221 first IMP (1–2) takes 3.4 ± 0.63 days on average,
 222 whereas the last IMP (4–5) takes about 4.8 ± 1.12
 223 days. Except for IMP (4–5), all the IMPs were signifi-
 224 cantly longer in the highest exposure concentration
 225 ($100 \mu\text{g}$ methoprene/l) compared to the respective con-
 226 trols. Although the first three IMPs appeared to be
 227 larger in the $1 \mu\text{g}$ methoprene/l treatment, these differ-
 228 ences were not statistically significant. Only the third
 229 IMP 185 (3–4) was significantly larger in *N. integer*
 230 exposed to $0.01 \mu\text{g}$ methoprene/l.
 231

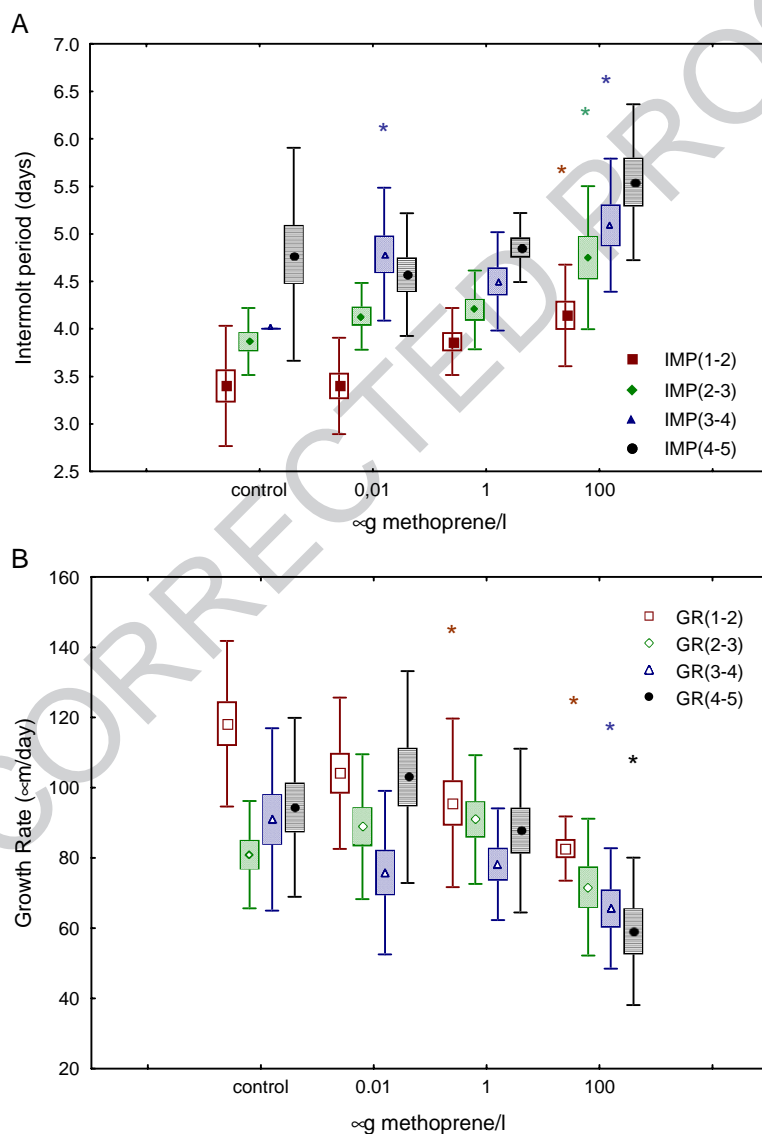


Fig. 4. Effect of methoprene on A) intermolt periods (IMP) and B) growth rates on five successive molts of *Neomysis integer*. * significantly different from control (Dunnett's; $466 p < 0.05$).

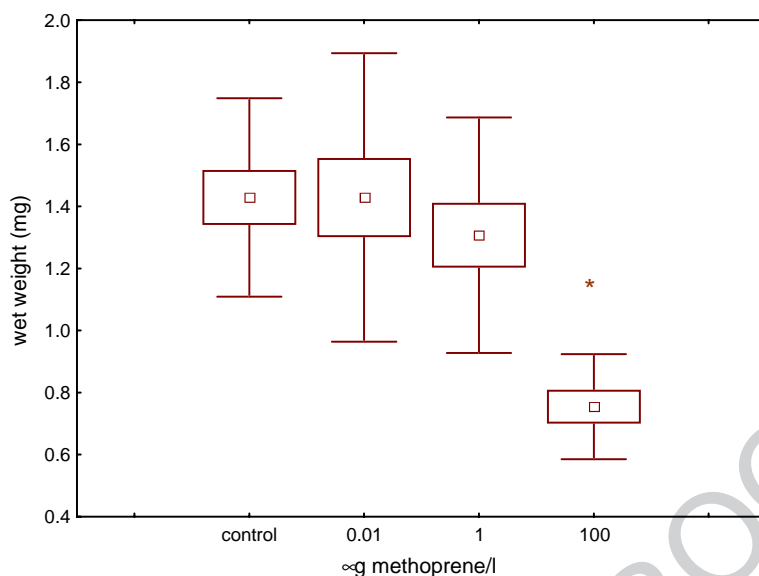


Fig. 5. Wet weight of *Neomysis integer* after the fifth molt, following exposure to methoprene. * significantly different from control (Dunnett's; $p < 0.05$).

232 3.2. Effect of methoprene on mysid growth rate

233 Fig. 4B shows the effect of methoprene on mysid
 234 growth rate during the first five molts. Generally, mysid
 235 growth rate is highest during the first molt GR (1–2)
 236 and subsequently decreases (Fockehey et al., in press).
 237 Significant effects were seen on mysid growth rate of
 238 juveniles exposed to 100 µg methoprene/l for all molts
 239 (GR (1–2), (3–4), (4–5)), except the second GR (2–3).
 240 Exposure to 1 µg methoprene/l reduced the growth rate
 241 at the first molt only. When growth rate is calculated as

total growth (µm) over the total exposure time (day), a
 significant decrease is found in the 100 µg methoprene/
 l treatment (data not shown). 242
 243
 244

3.3. Effect of methoprene on mysid wet weight 245

After the fifth molt, all organisms were weighed. 246
 Fig. 5 shows the effect of methoprene on mysid wet
 weight. There was a significant decrease in wet weight
 at the highest exposure concentration compared to control
 animals. The average wet weight of control organ- 247
 248
 249
 250

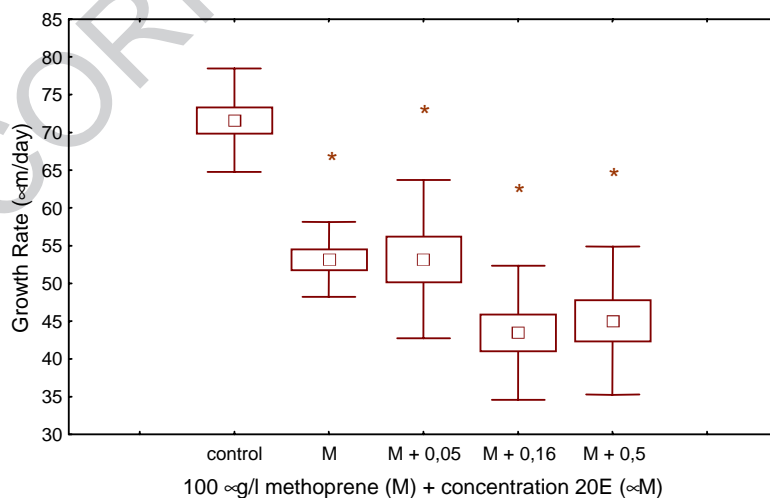


Fig. 6. Growth rate expressed as total growth over the total exposure time (in µm day⁻¹), following exposure to 100 µg methoprene/l (M) spiked with increased concentrations of 20-hydroxyecdysone (0.05, 0.16 and 0.5 µM 20E). * significantly different from control (Dunnett's; $p < 0.05$).

251 isms was 1.43 ± 0.32 mg, almost double of organisms
252 in the 100 μg methoprene/l treatment (average wet
253 weight of 0.75 ± 0.17 mg).

254 3.4. Combined effects of methoprene and 255 20-hydroxyecdysone

256 At 100 $\mu\text{g}/\text{l}$ methoprene significantly reduced mysid
257 growth rate by delaying the IMPs (Fig. 5). To further
258 investigate the anti-ecdysteroidal effects of metho-
259 prene, mysids were co-exposed to the active ecdyster-
260 oid, 20-hydroxyecdysone, to establish whether the
261 observed methoprene effect (IMP delay and decreased
262 growth rate) could be mitigated. Fig. 6 shows the
263 growth rate expressed as total growth (μm) during
264 the total exposure time (day) to 100 μg methoprene/
265 l and increasing concentrations of 20-hydroxyecdysone
266 (0.05, 0.16 and 0.5 μM 210 20E). 20-hydroxyecdysone
267 did not mitigate the putative anti-ecdysteroidal effects
268 on growth rate caused by methoprene. The effects of
269 methoprene on mysid growth reduction were con-
270 firmed in this second study.

271 4. Discussion

272 Ecdysteroids (molting hormones) and juvenoids (ju-
273 venile hormones) represent two classes of hormones in
274 arthropods that regulate many aspects of their develop-
275 ment, growth, and reproduction. Therefore, chemicals
276 that disrupt normal ecdysteroid/juvenoid signaling
277 could have profound effects on many aspects of inver-
278 tebrate function. During their development, insects un-
279 dergo changes at specific times (such as pupation)
280 which are mediated by endogenous hormones. Ecdy-
281 sone, the molting hormone, triggers larva-to-larva molts
282 as long as the juvenile hormone is present. In its
283 absence, ecdysone promotes the pupa-to-adult molt.
284 Thus, juvenile hormone present at specific times during
285 insect development leads to normal metamorphosis,
286 however, if present at other times it will lead to mor-
287 phogenetic abnormalities.

288 This is the basic theory behind the use of metho-
289 prene and other juvenile hormone analogues (e.g. pyr-
290 iproxyfen and fenoxycarb) as insect growth regulators
291 (Dhadialla et al., 1998; Hoffmann and Lorenz, 1998).
292 Methoprene is therefore not directly toxic to insects, but
293 as it disrupts the development of the insect it causes
294 death or reproductive failure at a specific time during
295 the insect life-cycle.

296 A large portion of the aquatic fauna are crustaceans,
297 making the group important for assessing the non-target
298 effects of many pesticides – such as the mosquitocidal

agent methoprene – that end up in aquatic ecosystems 299
(McKenney and Celestial, 1996; Olmstead and 300
LeBlanc, 2001; Peterson et al., 2001; Templeton and 301
Laufer, 1983). As the potential invertebrate-specific 302
endocrine-disruptive effects of chemicals to non-target 303
organisms are presently not specifically addressed in 304
regulatory screening and testing programs, this could 305
lead to significant underestimations of the actual envi- 306
ronmental risk of these chemicals. 307

308 While growth through molting of *Neomysis integer* 309
has been described in the laboratory (Astthorsson and 310
Ralph, 1984; Fockekey et al., in press; Winkler and 311
Greve, 2002), its disruption by chemicals through spe- 312
cific hormone-regulated mechanisms has not been stud- 313
ied. Methoprene effects on growth of *N. integer* were 314
observed after the first molt, which should therefore 315
allow the use of shorter exposure periods in future 316
studies. Methoprene is acutely toxic (96 h) to *Neomysis* 317
integer at 320 $\mu\text{g}/\text{l}$ (Verslycke et al., 2004a,b) and to 318
Americamysis bahia at 125 $\mu\text{g}/\text{l}$ (McKenney and Cele- 319
stial, 1996). McKenzie and Celestial (1996) examined 320
the influence of methoprene on survival, growth and 321
reproduction of *A. bahia* during a complete life cycle, 322
from one-day-old juvenile through juvenile growth and 323
maturation and production of young as an adult. The 324
most sensitive response was a significant reduction in 325
the number of young produced per female at concen- 326
trations $\cdot 2$ $\mu\text{g}/\text{l}$. The mysids weighed significantly less 327
at exposure concentration of 62 μg methoprene/l as 328
compared to the controls, which is in the same range 329
as what we found in this study (*N. integer* weighed 330
significantly less at 100 $\mu\text{g}/\text{l}$). Our results also corrob- 331
orate effect concentrations reported for other non-target 332
crustaceans. Methoprene significantly reduced comple- 333
tion of larval metamorphosis in the estuarine grass 334
shrimp *Palaemonetes pugio* at a concentration of 100 335
 $\mu\text{g}/\text{l}$ (McKenney and Matthews, 1990). Methoprene 336
adversely affected molting and reproduction in 337
D. magna at concentrations higher than 30 nM 338
(~ 10 $\mu\text{g}/\text{l}$) (Olmstead and LeBlanc, 2001). Recently, 339
we found that methoprene adversely affects the energy 340
and steroid metabolism of *N. integer* at 100 $\mu\text{g}/\text{l}$ (Ver- 341
slycke et al., 2004a,b).

342 The present study demonstrates that methoprene 343
significantly affects mysid molting and growth at sub- 344
lethal concentrations. However, previously reported 345
methoprene effect on mysid reproduction was noted 346
at lower concentrations (McKenney and Celestial, 347
1996). As juvenoids and ecdysteroids play a crucial 348
role in the regulation of mysid growth, reproduction 349
and development, comparative approaches that look at 350
a range of ecdysteroid/juvenoid regulated processes in 351

351 crustaceans should be informative in selecting which
352 endpoints are most sensitive. In addition, measuring the
353 hormones and receptors involved in mysid ecdysteroid/
354 juvenoid signaling will provide insights into the mode-
355 of-action of juvenile hormone analogues and other
356 pesticides in non-target arthropods and how this com-
357 pares to what is known in insects. In an effort to
358 improve our understanding of ecdysteroid/juvenoid sig-
359 naling in mysids, we have recently developed assays to
360 study mysid vitellogenesis (Ghekiere et al., in press),
361 embryonic development (Fockedeey et al., in prepara-
362 tion), ecdysteroid receptor interaction (Verslycke, per-
363 sonal communication) and are validating these assays in
364 exposure studies with methoprene and other pesticides
365 (Ghekiere et al., in preparation).

366 Although the ecdysteroid hormone 20-hydroxyecdysone acts as a EcR ligand and activates transcription through EcR/USP heterodimers, the activity of juvenoids and juvenile hormone-analogs such as methoprene remains unclear. Recently, Maki et al. (2004) have demonstrated that JH III- and methoprenic acid-bound USP markedly repressed ecdysone-dependent EcR transcription.

374 In the second part of the present study, we evaluated
375 the anti-ecdysteroidal activity of methoprene in *N. in-*
376 *teger* by exogenously administering 20-hydroxyecdysone. We found no mitigation of the inhibiting effect on
377 growth. This could indicate that methoprene did not
378 exert its effect through ecdysteroid receptor antagonism. Mu and LeBlanc (2002) demonstrated that tes-
380 tosterone had an anti-ecdysteroidal activity in *D. magna*
382 by delaying the molt frequency and this effect was
383 mitigated by co-exposure to 20-hydroxyecdysone.
384 They proposed ecdysteroid receptor antagonism as
385 one possible mechanism by which testosterone caused
386 these effects. We previously tested the effect of 20-
387 hydroxyecdysone on molting of *N. integer* and found
388 no effects on the molting frequency. Raising the con-
389 centration to 10^{-5} M was associated with premature
390 death caused by incomplete ecdysis. These results cor-
391 relate with the findings of Baldwin et al. (2001). Al-
392 though recent studies with daphnids indicate that
393 juvenoids modulate ecdysteroid signaling through a
394 mechanism that may involve reduced availability of
395 the receptor partner protein ultraspiracle (the ecdysone
396 receptor is functional only as a heterodimer with ultra-
397 spiracle), the exact mechanism of action of juvenoids
398 and methoprene remains unclear (Mu and LeBlanc,
399 2004). In this respect, the increasing availability of
400 sequences for the different receptors involved in crus-
401 tacean ecdysteroid/juvenoid signaling may be very
402 valuable.

The endocrine system of an invertebrate differs from
that of a vertebrate organism both in the type of endo-
crine glands present and in the chemical structure of
specific hormones that are produced. As such, assessing
the impact of endocrine disrupting chemicals on inver-
tebrates, requires an approach that is specifically direct-
ed at invertebrates. In this context, we are exploring a
range of endocrine-regulated processes in invertebrates
that could be specifically disrupted by chemicals. This
approach should lead to both a better understanding of
hormone regulation and its disruption by chemicals in
invertebrates.

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References

- Astthorsson, O.S., Ralph, R., 1984. Growth and moulting of *Neomysis integer* (Crustacea: Mysidacea). Mar. Biol. 79, 55–61.
- Baldwin, W.S., Bailey, R., Long, K.E., Klaine, S., 2001. Incomplete ecdysis is an indicator of ecdysteroid exposure in *Daphnia magna*. Environ. Toxicol. Chem. 20 (7), 1564–1569.
- Barnes, R.D., 1980. Invertebrate Zoology. W.B. Saunders, Philadelphia, PA, USA.
- Celestial, D.M., McKenney Jr., C.L., 1994. The influence of an insect growth regulator on the larval development of the mud crab *Rhithropanopeus harrisi*. Environ. Pollut. 85, 169–173.
- Dhadialla, T.S., Carlson, G.R., Le, D.P., 1998. New insecticides with ecdysteroidal and juvenile hormone activity. Annu. Rev. Entomol. 43, 545–569.
- Fockedeey, N., Mees, J., Vangheluwe, M., Verslycke, T., Janssen, C.R., in press. Vincx, M. Growth of *Neomysis integer* (Crustacea: Mysidacea) under different salinity-temperature conditions. J. Exp. Mar. Biol. Ecol.
- Ghekiere, A., Fenske, M., Verslycke, T., Tyler, C., Janssen, C.R., in press. Development of a quantitative enzyme-linked immunosorbent assay to study vitellogenesis in the mysid *Neomysis integer* (Crustacea: Mysidacea). Comp. Biochem. Physiol. A.
- Harmon, M.A., Boehm, M.F., Heyman, R.A., Mangelsdorf, D.J., 1995. Activation of mammalian retinoid X receptors by the insect growth regulator methoprene. Proc. Natl. Acad. Sci. U. S. A. 92, 6157–6160.
- Hoffmann, K.H., Lorenz, M.W., 1998. Recent advances in hormones in insect pest control. Phytoparasitica 26 (4), 1–8.
- Ingersoll, C., Hutchinson, T., Crane, M., Dodson, S., DeWitt, T., Gies, A., Huet, M., McKenney, C., Oberdorster, E., Pascoe, D., Versteeg, D., Warwick, O., 1999. Laboratory toxicity tests

- 457 for evaluating potential effects of endocrine-disrupting com- 499
 458 pounds. In: deFur, P., Crane, M., Ingersoll, C., Tattersfield, 500
 459 L. (Eds.), *Endocrine Disruption in Invertebrates: Endocrinol-* 501
 460 *ogy, Testing, and Assessment*. SETAC Press, Pesacola, FL, 502
 461 pp. 107–197. 503
- 462 Knuth, M.L., 1989. Determination of the insect growth regulator 504
 463 methoprene in natural waters by capillary gas–liquid chromatog- 505
 464 raphy. *Chemosphere* 18, 2275–2281. 506
- 465 LaClair, J.J., Bantle, J.A., Dumont, J., 1998. Photoproducts and 507
 466 metabolites of a common insect growth regulator produce devel- 508
 467 opmental deformities in *Xenopus*. *Environ. Sci. Technol.* 32, 509
 468 1453–1461. 510
- 469 Maki, A., Sawatsubashi, S., Ito, S., Shirode, Y., Suzuki, E., Zhao, Y., 511
 470 Yamagata, K., Kouzmenko, A., Takeyama, K.-I., Kato, S., 2004. 512
 471 Juvenile hormones antagonize ecdysone actions through co-re- 513
 472 pressor recruitment to EcR/USP heterodimers. *Biochem. Biophys.* 514
 473 *Res. Commun.* 320, 262–267. 515
- 474 McKenney Jr., C.L., Celestial, D.M., 1996. Modified survival, 516
 475 growth and reproduction in an estuarine mysid (*Mysidopsis* 517
 476 *bahia*) exposed to a juvenile hormone analogue through a com- 518
 477 plete life cycle. *Aquat. Toxicol.* 35, 11–20. 519
- 478 McKenney Jr., C.L., Matthews, E., 1990. Influence of an insect 520
 479 growth regulator on the larval development of an estuarine 521
 480 shrimp. *Environ. Pollut.* 64 (2), 169–178. 522
- 481 Mu, X., LeBlanc, G.A., 2002. Developmental toxicity of testosterone 523
 482 in the crustacean *Daphnia magna* involves anti-ecdysteroidal 524
 483 activity. *Gen. Comp. Endocrinol.* 129, 127–133. 525
- 484 Mu, X., LeBlanc, G.A., 2004. Cross communication between signal- 526
 485 ing pathways: juvenoid hormones modulate ecdysteroid activity in 527
 486 a crustacean. *J. Exp. Zool.* 301A, 793–801. 528
- 487 Olmstead, A.W., LeBlanc, G.L., 2001. Low exposure concentration 529
 488 effects of methoprene on endocrine-regulated processes in the 530
 489 crustacean *Daphnia magna*. *Toxicol. Sci.* 62, 268–273. 531
- 490 Peterson, J.K., Kashian, D.R., Dodson, S.I., 2001. Methoprene and 532
 491 20-OH-ecdysone affect male production in *Daphnia pulex*. *Envi-* 533
 492 *ron. Toxicol. Chem.* 20 (3), 582–588. 534
- 493 Quistad, G.B., Staiger, L.E., Schooley, D.A., 1975. Environmental 535
 494 degradation of the insect growth regulator methoprene. II. Photo- 536
 495 decomposition. *J. Agric. Food Chem.* 23 (2), 299. (MIRID 537
 496 #5008610). 538
- 497 Retnakaran, A., Granett, G., Ennis, T., 1985. Insect growth regulators. 539
 498 In: Kerkut, G.A., Gilbert, L.I. *Comprehensive Insect Physiology,* 540
 541 *Biochemistry and Pharmacology* vol. 12. Pergamon, Oxford, pp. 499
 529–601. 500
 513 Roller, H., Dahm, K.H., Sweeley, C.C., Trost, B.M., 1967. The 501
 514 structure of the juvenile hormone. *Angew. Chem., Int. Ed. Engl.* 502
 515 6, 179–180. 503
 516 Schaefer, C.H., Dupras, E.F., 1973. Insect development inhibitors. 4: 504
 517 persistence of ZR-515 405 in water. *J. Econ. Entomol.* 66 (4), 505
 518 923–925. MIRID #5008625. 506
 519 Subramoniam, T., 2000. Crustacean ecdysteroids in reproduction and 507
 520 embryogenesis. *Comp. Biochem. Physiol., C* 2, 135–156. 508
 521 Templeton, N.S., Laufer, H., 1983. The effects of a juvenile hormone 509
 522 analog (Altosid ZR-515) on the reproduction and development of 510
 523 *Daphnia magna* (Crustacea: Cladocera). *Int. J. Invertebr. Reprod.* 511
 524 6, 99–110. 512
 525 US EPA., 2001. US Environmental Protection Agency. Pesticide fact 513
 526 sheet. June 2001 update of the March 1991 methoprene R.E.D. 514
 527 fact sheet [online]: available at [http://www.epa.gov/pesticides/](http://www.epa.gov/pesticides/biopesticides/ingredients/factsheets/factsheet_105401.pdf) 515
 528 [biopesticides/ingredients/factsheets/factsheet_105401.pdf](http://www.epa.gov/pesticides/biopesticides/ingredients/factsheets/factsheet_105401.pdf). 516
 529 Verslycke, T., Janssen, C.R., 2002. Effects of a changing abiotic 517
 530 environment on the abiotic environment on the energy metabolism 518
 531 in the estuarine mysid shrimp *Neomysis integer* (Crustacea; Mysi- 519
 532 dacea). *J. Exp. Mar. Biol. Ecol.* 279, 61–72. 520
 533 Verslycke, T., De Wasch, K., De Brabander, H.F., Janssen, C.R., 521
 534 2002. Testosterone metabolism in the estuarine mysid *Neomysis* 522
 535 *integer* (Crustacea; Mysidacea): identification of testosterone 523
 536 metabolites and endogenous vertebrate-type steroids. *Gen.* 524
 537 *Comp. Endocrinol.* 126, 190–199. 525
 538 Verslycke, T., Fockedey, N., McKenney, C.L., Roast, S.D., Jones, 526
 539 M.B., Mees, J., Janssen, C.R., 2004a. Mysid crustaceans as 527
 540 potential test organisms for the evaluation of environmental en- 528
 541 docrine disruption: a review. *Environ. Toxicol. Chem.* 23 (5), 529
 530 1219–1234. 531
 532 Verslycke, T., Poelmans, S., De Wasch, K., De Brabander, H.F., 532
 533 Janssen, C.R., 2004b. Testosterone and energy metabolism of 533
 534 *Neomysis integer* following exposure to endocrine disruptors. 534
 535 *Environ. Toxicol. Chem.* 23 (5), 1289–1296. 535
 536 Williams, C.M., 1956. The juvenile hormone of insects. *Nature* 178, 535
 537 212–213. 536
 538 Winkler, G., Greve, W., 2002. Laboratory studies of the effect of 537
 539 temperature on growth, moulting and reproduction in the co- 538
 540 occurring mysids *Neomysis integer* and *Praunus flexuosus*. *Mar.* 539
 541 *Ecol. Prog. Ser.* 235, 177–188. 540