



Seasonal response of air–water CO₂ exchange along the land–ocean aquatic continuum of the northeast North American coast.

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Abstract. This regional study quantifies the CO₂ exchange at the air–water interface along the land–ocean aquatic continuum (LOAC) of the northeast North American coast, from streams to the shelf break. Our analysis explicitly accounts for spatial and seasonal variability in the CO₂ fluxes. The yearly integrated budget reveals the gradual change in the intensity of the CO₂ exchange at the air–water interface, from a strong source towards the atmosphere in streams and rivers ($3.0 \pm 0.5 \text{ TgC yr}^{-1}$) and estuaries ($0.8 \pm 0.5 \text{ TgC yr}^{-1}$) to a net sink in continental shelf waters ($-1.7 \pm 0.3 \text{ TgC yr}^{-1}$). Significant differences in flux intensity and their seasonal response to climate variations is observed between the North and South sections of the study area, both in rivers and coastal waters. Ice cover, snowmelt, and intensity of the carbon removal efficiency through the estuarine filter are identified as important control factors of the observed spatiotemporal variability in CO₂ exchange along the LOAC.

aries outgas 1.1 and 0.25 PgC yr⁻¹, respectively, while continental shelf seas take up 0.2 PgC yr⁻¹. However, CO₂ data are too sparse and unevenly distributed to provide global coverage and large uncertainties remain associated with these estimates. The inland water outgassing could for instance reach 2.1 PgC yr⁻¹ with 86 % coming from streams and rivers (Raymond et al., 2013), a value which is about twice that reported in Regnier et al. (2013a) and in the IPCC Fifth Assessment Report (Ciais et al., 2013). The most recent global budgets for the estuarine CO₂ source and the continental shelf CO₂ sink also reveal significant discrepancies, both falling within the 0.15–0.4 PgC yr⁻¹ range (Laruelle et al., 2010, 2013; Cai, 2011; Bauer et al., 2013; Dai et al., 2013). None of these estimates, however, fully resolves the seasonality in CO₂ fluxes because temporal coverage of the global data is insufficient. Complex seasonal dynamics of CO₂ exchanges between the atmosphere and individual components of the LOAC have been reported in previous studies which have highlighted the potential importance of the intra-annual variability for local and regional CO₂ budgets (e.g., Kempe, 1982; Frankignoulle et al., 1998; Jones and Mulholland, 1998; Degrandpré et al., 2002; Thomas and Schneider, 1999; Wallin et al., 2011; Regnier et al., 2013a; Rawlins et al., 2014). Here, we extend the analysis to the sub-continental scale, and present the spatial and seasonal variability of CO₂ fluxes at the air–water interface (F_{CO_2}) for the entire northeast North American LOAC, from streams to the shelf break. This region of unprecedented data coverage allows us to pro-

1 Introduction

Over the past decade, several syntheses have highlighted the significant contribution of the land–ocean aquatic continuum (LOAC) to the global atmospheric CO₂ budget (Cole et al., 2007; Battin et al., 2009; Mackenzie et al., 2012; Bauer et al., 2013; Ciais et al., 2013; Raymond et al., 2013; Regnier et al., 2013a). In a recent review, Regnier et al. (2013a) proposed that inland waters (streams, rivers and lakes) and estu-

duce, for the first time, empirically derived monthly maps of CO₂ exchange at 0.25° resolution. Our results allow us to investigate the seasonal CO₂ dynamics across the interconnected systems of the LOAC and elucidating their response to contrasting intra-annual changes in climate conditions.

2 Methods

Our study area is located along the Atlantic coast of the northern US and southern Canada and extends from the Albemarle Sound in the South section to the eastern tip of Nova Scotia in the North section. It corresponds to COSCAT 827 (for Coastal Segmentation and related CATchments) in the global coastal segmentation defined for continental land masses by Meybeck et al. (2006) and extrapolated to continental shelf waters by Laruelle et al. (2013). COSCATs are homogenous geographical units that divide the global coastline into homogeneous segments according to lithological, morphological, climatic, and hydrological properties. The area corresponding to COSCAT 827 comprises 447 × 10³ km² of watersheds and 357 × 10³ km² of coastal waters, amongst which 15 × 10³ km² of estuaries. It is one of the best monitored regions in the world with several regularly surveyed rivers (Hudson, Susquehanna, York, Connecticut) and some of the most extensively studied coastal waters (Degrandpré et al., 2002; Chavez et al., 2007; Fennel et al., 2008; Fennel and Wilkin, 2009; Previdi et al., 2009; Fennel, 2010; Shadwick et al., 2010, 2011; Signorini et al., 2013). For the purpose of this study, the area was divided into a North and a South section (Fig. 1). The boundary is set on land to distinguish the regions subject to seasonal ice freeze and snowfalls from those that are not (Armstrong and Brodzik, 2001). This delineation attributes 96 % of the estuarine surface area to the South section due, for the most part, to the contribution of Chesapeake Bay which accounts for about two thirds of the estuarine area. The delineation extends further into the coastal waters in such a way that the Scotian Shelf and the Gulf of Maine correspond to the North section and the Mid-Atlantic Bight and Georges Bank to the South section. The riverine data are calculated from pH and alkalinity measurements extracted from the GLOBAL RIVER CHEMISTRY Database (GLORICH (Hartmann et al., 2014), previously used in Lauerwald et al., 2013), while continental shelf values are calculated from the Surface Ocean CO₂ Atlas (SOCAT v2.0) database which contains quality controlled direct pCO₂ measurements (<http://www.socat.info/>, Bakker et al., 2014).

2.1 Rivers

CO₂ evasion from rivers (F_{CO_2}) was calculated monthly per 15 s grid cell (resolution of the hydrological routing scheme HydroSHEDS 15 s, Lehner et al., 2008) from estimates of the

effective stream/river surface area A_{eff} [m²], gas exchange velocity k [m d⁻¹], and water–atmosphere CO₂ concentration gradient $\Delta[\text{CO}_2]$ [$\mu\text{mol l}^{-1}$]:

$$F_{CO_2} = A_{\text{eff}} \times k \times \Delta[\text{CO}_2]. \quad (1)$$

The calculation of A_{eff} first requires estimation of the total stream/river surface area, A . The latter was calculated from the linear stream network derived from the HydroSHEDS 15 s routing scheme using a minimum threshold on the catchment area of 10 km², and estimates of stream width derived from the annual mean discharge Q_{ann} using the equations of Raymond et al. (2012, 2013) (Eqs. 2, 3). Values of A were not calculated for each individual month, as the discharge–stream width relationship only hold true for Q_{ann} (Raymond et al., 2013). Q_{ann} was obtained using HydroSHEDS 15 s to route the gridded data of average annual runoff from the UNH/GRDC composites (Fekete et al., 2002).

$$\ln(B[\text{m}]) = 2.56 + 0.423 \times \ln(Q_{\text{ann}}[\text{m}^3 \text{s}^{-1}])$$

(Eq. 2 after Raymond et al., 2012),

$$\ln(B[\text{m}]) = 1.86 + 0.51 \times \ln(Q_{\text{ann}}[\text{m}^3 \text{s}^{-1}])$$

(Eq. 3 after Raymond et al., 2013),

where B is stream width [m] and Q_{ann} is annual average discharge [$\text{m}^3 \text{s}^{-1}$].

For each 15 s raster cell covered by lake and reservoir areas as represented in the global lake and wetland database of Lehner and Döll (2004), A was set to 0 km². A_{eff} was then derived from A to account for seasonal stream drying and ice cover inhibiting F_{CO_2} . Seasonal stream drying was assumed for each 15 s cell and month when the monthly average discharge Q_{month} is 0 m³ s⁻¹. Values of Q_{month} were calculated similarly to that of Q_{ann} using the gridded data of average monthly runoff from the UNH/GRDC composites (Fekete et al., 2002). Ice cover was assumed for each 15 s cell and month when the mean air temperature (T_{air}), derived from the WorldClim data set of Hijmans et al. (2005), is below −4.8° C (Lauerwald et al., 2015). In case of ice cover and/or stream drying, A_{eff} is set to 0 m². Otherwise A_{eff} equals A .

Values of k were first calculated as standardized values for CO₂ at a water temperature (T_{water}) of 20° C (k_{600}), from stream channel slope CS and estimates of flowing velocity V (Eq. 4). Using the Strahler order (Strahler, 1952) to perform the segmentation of the stream network, CS was calculated for each segment by dividing the change in its altitude by its length. Information on altitude was derived from the HydroSHEDS elevation model. V was calculated from Q_{ann} based on the equations of Raymond et al. (2012, 2013) (Eqs. 5, 6). Similarly to the stream width, the $V - Q$ relationship only holds true for Q_{ann} (Raymond et al., 2013), and this is why only annually average values for V and k_{600} could be

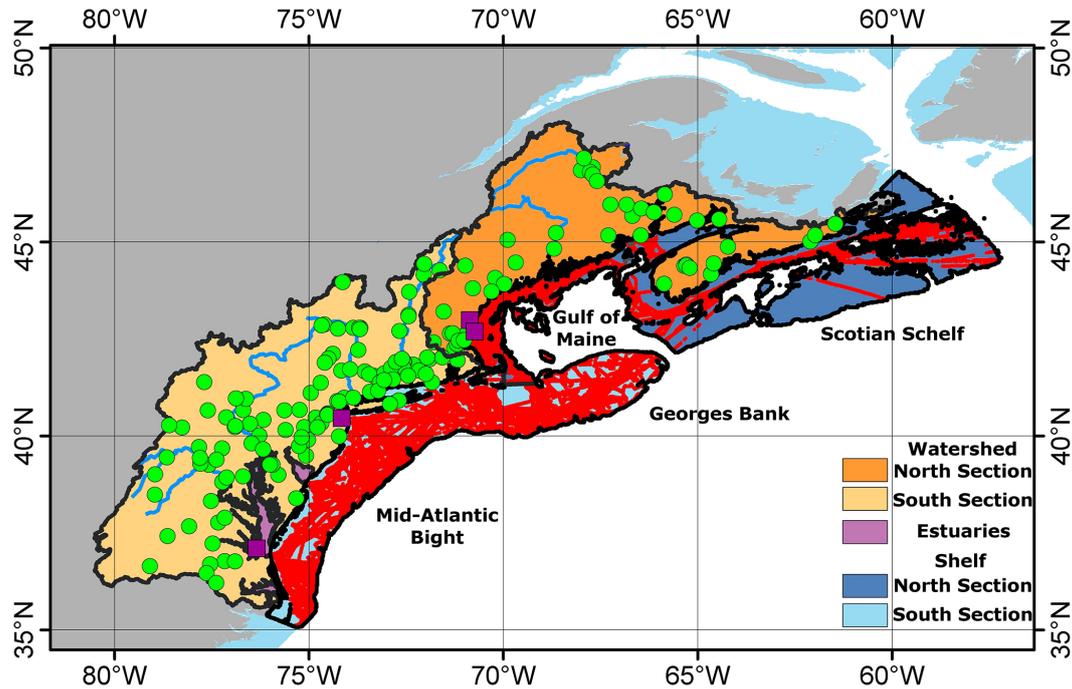


Figure 1. Geographic limits of the study area with the location of the riverine (GLORICH database, in green; Lauerwald et al., 2013) and continental shelf waters data used for our calculations (SOCAT 2.0 database, in red; Bakker et al., 2014). The location of the estuarine studies used is indicated by purple squares.

calculated. The k value for each month was calculated from k_{600} , an estimate of the average monthly water temperature T_{water} (Lauerwald et al., 2015; Raymond et al., 2012).

$$k_{600}[\text{m d}^{-1}] = V[\text{m s}^{-1}] \times \text{CS}[1] \times 2841 + 2.02$$

(Eq. 4 after Raymond et al., 2012),

$$\ln(V[\text{m s}^{-1}]) = -1.64 + 0.285 \times \ln(Q_{\text{ann}}[\text{m}^3 \text{s}^{-1}])$$

(Eq. 5 after Raymond et al., 2012),

$$\ln(V[\text{m s}^{-1}]) = -1.06 + 0.12 \times \ln(Q_{\text{ann}}[\text{m}^3 \text{s}^{-1}])$$

(Eq. 6 after Raymond et al., 2013),

where k_{600} is the standardized gas exchange velocity for CO₂ at 20°C water temperature [m d^{-1}], Q_{ann} is annual average discharge [$\text{m}^3 \text{s}^{-1}$], V stream flow velocity [m s^{-1}], and CS channel slope [dimensionless].

Values of $\Delta(\text{CO}_2)$ were derived from monitoring data with calculated $p\text{CO}_{2\text{river}}$ (12 300 water samples, from 161 locations, Lauerwald et al., 2013), and an assumed $p\text{CO}_{2\text{atmosphere}}$ of 390 μatm . Lauerwald et al. (2013) calculated $p\text{CO}_{2\text{river}}$ values from pH, alkalinity, water temperature, and, where available, major ion concentrations, using the hydrochemical modeling software Phreeqc v2 (Parkhurst and Appelo, 1999). The $p\text{CO}_2$ values were converted into concentrations, $[\text{CO}_2]$, using Henry's constant (Henry, 1803) for each sample at its observed temperature T_{water} using the

equation of Telmer and Veizer (1999). In order to minimize the influence of extreme values, the results were aggregated to median values per sampling location and month for which at least three values were available. These median values per sampling location and month were then used to calculate maps of $\Delta[\text{CO}_2]$ at a 15 s resolution. To this end, an inverse distance-weighted interpolation was applied. This method allows us to predict a value for each grid cell from observed values at the four closest sampling locations, using the inverse of the squared distance between the position on the grid and each sampling locations as weighting factors. To account for downstream decreases in $p\text{CO}_{2\text{river}}$, which are often reported in the literature (Finlay, 2003; Teodoru et al., 2009; Butman and Raymond, 2011), the interpolation was applied separately to three different classes of streams and rivers defined by Q_{ann} , for which sufficiently large subsets of sampling locations could be retained: (1) $Q_{\text{ann}} < 10 \text{ m}^3 \text{ s}^{-1}$ ($n = 76$), (2) $10 \text{ m}^3 \text{ s}^{-1} \leq Q_{\text{ann}} < 100 \text{ m}^3 \text{ s}^{-1}$ ($n = 47$), and (3) $Q_{\text{ann}} \geq 100 \text{ m}^3 \text{ s}^{-1}$ ($n = 38$). The three maps of $\Delta[\text{CO}_2]$ per month were then recombined according to the spatial distribution of Q_{ann} values. The $F\text{CO}_2$ values were first calculated using Eq. (1) at the high spatial resolution of 15 s for each month. The results were then aggregated to a 0.25° resolution and 3-month period and reported as area-specific values referring to the total surface area of the grid cell. At the outer boundaries, only the proportions of the cell covered by our study area are taken into account. The difference between the

FCO_2 values calculated using the equations of Raymond et al. (2012) and Raymond et al. (2013) was used as an estimate of the uncertainty of the mean yearly FCO_2 . The aforementioned method is consistent with the approach of Raymond et al. (2013), which used two distinct sets of equations for k and A to estimate the uncertainty in these parameters and their combined effect on the estimated FCO_2 .

2.2 Estuaries

The yearly averaged CO₂ exchange at the air–water interface was obtained from local estimations of emission rates in seven estuaries located within the study area (see Table 1). The limited number of observations does not allow us to resolve the seasonality in CO₂ emissions. The yearly average local CO₂ emission rates range from 1.1 molC m⁻² yr⁻¹ in the Parker River to 9.6 molC m⁻² yr⁻¹ in the Hudson River estuary, for a mean value of 4.2 molC m⁻² yr⁻¹ for the seven systems. This value was then multiplied by the estuarine surface areas extracted from the SRTM water body data set (NASA/NGA, 2003), to estimate the bulk outgassing for the North and South sections of COSCAT 827. It should be noted that the methods used to estimate the CO₂ emission rates differ from one study to the other (i.e., different relationships relating wind speed to the gas transfer coefficient). However, in the absence of a consistent and substantial estuarine pCO_2 database for the region, we believe that our method is the only one which allows one to derive a regional data driven estimate for the CO₂ outgassing from estuaries which would otherwise require the use of reactive transport models (Regnier et al., 2013b). Similar approaches have been used in the past to produce global estuarine CO₂ budgets (Borges et al., 2005; Laruelle et al., 2010, 2013; Cai, 2011; Chen et al., 2013). The standard deviation calculated for the emission rates of all local studies was used as an estimate of the uncertainty of the regional estuarine FCO_2 .

2.3 Continental shelf waters

Monthly CO₂ exchange rates at the air–water interface were calculated in continental shelf waters using 274 291 pCO_2 measurements extracted from the SOCAT 2.0 database (Bakker et al., 2014). For each measurement, an instantaneous local CO₂ exchange rate with the atmosphere was calculated using Wanninkhof's equation (Wanninkhof, 1992), which is a function of a transfer coefficient (k), dependent on the square of the wind speed above sea surface, the apparent solubility of CO₂ in water (K'_0) [moles m⁻³ atm⁻¹], which depends on surface water temperature and salinity, and the gradient of pCO_2 at the air–water interface (ΔpCO_2) [μ atm].

$$FCO_2 = A_s \times k \times K'_0 \times \Delta pCO_2 \quad (2)$$

The parameterization used for k is that of Wanninkhof et al. (2013), and all the data necessary for the calculations are

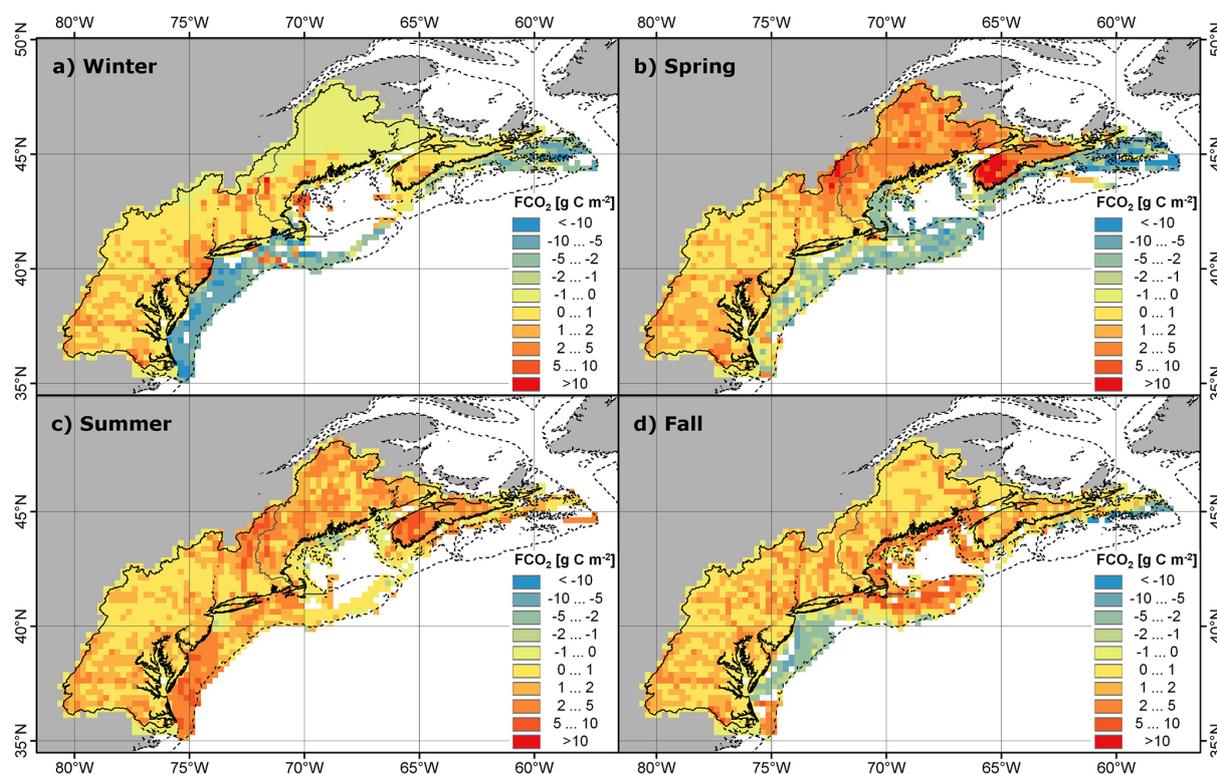
available in SOCAT 2.0 except for wind speed, which was extracted from the CCMP database (Atlas et al., 2011). The resulting CO₂ exchange rates were then averaged per month for each 0.25° cell in which data were available. Average monthly CO₂ exchange rates were calculated for the North and South sections using the water surface area and weighted rate for each cell, and those averages were then extrapolated to the entire surface area A_s of the corresponding section to produce FCO_2 . In effect, this corresponds to applying the average exchange rate of the section to the cells devoid of data. To refine further the budget, a similar procedure was also applied to 5 depth segments (S1 to S5) corresponding to 0–20, 20–50, 50–80, 80–120 and 120–150 m, respectively, and their respective surface areas were extracted from high resolution bathymetric files (Laruelle et al., 2013). The choice of slightly different methodologies for FCO_2 calculations in rivers and continental shelf waters stems from the better data coverage in the continental shelf, which allows the capturing of spatial heterogeneity within the region without using interpolation techniques. The standard deviation calculated for all the grid cells of the integration domain was used as the uncertainty of the yearly estimates of FCO_2 . A more detailed description of the methodology applied to continental shelf waters at the global scale is available in Laruelle et al. (2014).

3 Results and discussion

Figure 2 shows the spatial distribution of FCO_2 along the LOAC integrated per season. Throughout the year, river waters are a strong source of CO₂ for the atmosphere. Significant differences in the intensity of the CO₂ exchange at the air–water interface can nevertheless be observed between the North and South sections, both in time and space. During winter, there is nearly no CO₂ evasion from rivers in the North section due to ice coverage and stream drying. Over the same period, the CO₂ emissions from the South section range from 0 to 5 gC m⁻² season⁻¹. During spring, the pattern is reversed and rivers in the North exhibit higher outgassing rates than in the South section with maximum emissions rates of > 10 gC m⁻² season⁻¹. This trend is maintained throughout summer while during fall, the entire COSCAT displays similar emission rates without a clear latitudinal signal. Continental shelf waters display a very different spatial and seasonal pattern than that of rivers. During winter, the North section is predominantly a mild CO₂ sink, with rates between +2 and –5 gC m⁻² season⁻¹, which intensifies significantly in the South section (–2 to > –10 gC m⁻² season⁻¹). During spring, an opposite trend is observed, with a quasi-neutral CO₂ uptake in the South section and a strong uptake in the North section, especially on the Scotian Shelf. The entire COSCAT becomes a net CO₂ source in summer with emission rates as high as 5 gC m⁻² season⁻¹ in the Mid-Atlantic Bight. During fall, the Gulf of Maine and Georges Bank remain CO₂ sources,

Table 1. Summary of the data used for the FCO_2 calculations in compartment of the LOAC.

Compartment	Parameter	Description	Source	Reference
Rivers	pCO_2	CO ₂ partial pressure	GLORICH	Hartmann et al. (2014)
	–	River network, digital elevation model (DEM)	HydroSHEDS 15 s	Lehner et al. (2008)
	–	Runoff	UNH/GRDC	Fekete et al. (2002)
	T	Air temperature	–	Hijmans et al. (2005)
	–	Lake surface area	Global Lake and Wetland Database	Lehner and Döll (2004)
Estuaries	A_s	Surface Area	SRTM water body data set	NASA/NGA (2003)
	–	CO ₂ exchange rate	Average of local estimates	Raymond et al. (1997) Raymond et al. (2000) Raymond and Hopkinson (2003) Hunt et al. (2010)
Shelves	A_s	Surface area	COSCAT/MARCATS Segmentation	Laruelle et al. (2013)
	ΔpCO_2	pCO_2 gradient at the air–water interface	SOCAT database	Bakker et al. (2014)
	k	Calculated using wind Speed	CCMP database	Altas et al. (2011)
	K'_0	Solubility, calculated using salinity, water temperature	SOCAT database	Bakker et al. (2014)

**Figure 2.** Spatial distribution of the CO₂ exchange with the atmosphere in rivers and continental shelf waters aggregated by seasons. The fluxes are net FCO_2 rates averaged over the surface area of each 0.25° cell and a period of 3 months. Positive values correspond to fluxes towards the atmosphere. Winter is defined as January, February, and March, Spring as April, May, and June, and so forth.

while the Scotian Shelf and the Mid-Atlantic Bight become again regions of net CO₂ uptake.

The monthly integrated F_{CO_2} for the North and South sections provides further evidence of the contrasting seasonal dynamics for the two areas (Fig. 3a and b). In the North section, CO₂ evasion from rivers is almost zero in January and February, rises to a maximum value of $0.26 \pm 0.05 \text{ TgC month}^{-1}$ in May and then progressively decreases until the end of the year. These low winter values are explained by the ice cover inhibiting the gas exchange with the atmosphere. The steep increase and F_{CO_2} maximum in spring could be related to the flushing of water from the thawing top soils, which are rich in dissolved organic carbon (DOC) and CO₂. Additionally, the temperature rise also induces an increase in respiration rates within the water columns (Jones and Mulholland, 1998; Striegl et al., 2012). Rivers and the continental shelf in the North section present synchronized opposite behaviors from winter through spring. In the shelf, a mild carbon uptake takes place in January and February ($-0.04 \pm 0.25 \text{ TgC month}^{-1}$), followed by a maximum uptake rate in April ($-0.50 \pm 0.20 \text{ TgC month}^{-1}$). This CO₂ uptake in spring has been attributed to photosynthesis associated with the seasonal phytoplankton bloom (Shadwick et al., 2010). Continental shelf waters behave quasi-neutral during summer ($<0.05 \pm 0.09 \text{ TgC month}^{-1}$) and emit CO₂ at a high rate in November and December ($>0.15 \pm 0.21 \text{ TgC month}^{-1}$). Overall, the rivers of the North section emit $1.31 \pm 0.24 \text{ TgC yr}^{-1}$, while the continental shelf waters take up $0.47 \pm 0.17 \text{ TgC yr}^{-1}$. The very limited estuarine surface area ($0.5 \times 10^3 \text{ km}^2$) only yields an annual outgassing of $0.03 \pm 0.02 \text{ TgC yr}^{-1}$. The shelf sink calculated for the region differs from that of Shadwick et al. (2011) which reports a source for the Scotian Shelves, in contrast to the current estimate. Our seasonally resolved budget is however in line with the -0.6 TgC yr^{-1} sink calculated by Signorini et al. (2013) using an 8-year data set as well as with the simulations of Fennel and Wilkin (2009), which also predict sinks of -0.7 and -0.6 TgC yr^{-1} for 2004 and 2005, respectively. No similar analysis was so far performed for inland waters.

In the South section of the COSCAT, the warmer winter temperature leads to the absence of ice cover (Armstrong and Brodzik, 2001). Our calculations predict that the riverine surface area remains stable over time, favoring a relatively constant outgassing between 0.1 and $0.2 \text{ TgC month}^{-1}$ throughout the year, adding up to a yearly source of $1.69 \pm 0.31 \text{ TgC yr}^{-1}$. Estuaries emit $0.73 \pm 0.45 \text{ TgC yr}^{-1}$, because of their comparatively large surface area ($14.5 \times 10^3 \text{ km}^2$), about 1 order of magnitude larger than that of rivers ($1.2 \times 10^3 \text{ km}^2$, Table 2). It should be noted that our estimate of the estuarine outgassing is derived from a limited number of local studies, none of which were performed in the two largest systems of COSCAT827, which are the Chesapeake and Delaware bays ($>80\%$ of the total estuarine surface area in COSCAT 827). These estuaries are highly eutrophic (Cai, 2011), which suggests that they might be character-

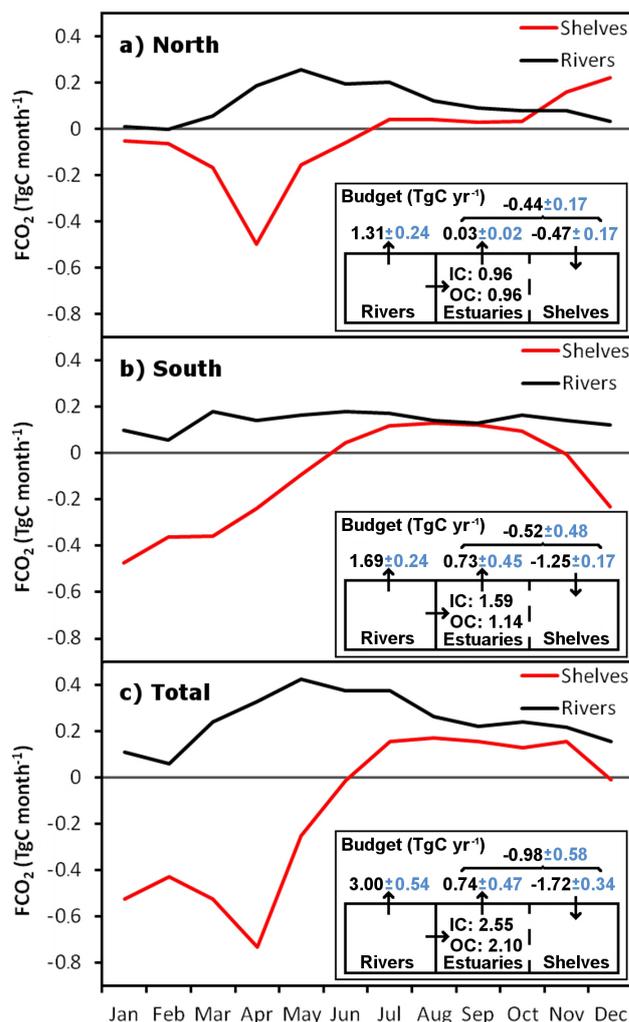


Figure 3. Areal-integrated monthly air–water CO₂ flux for rivers and the continental shelf waters in the North section (a), South section (b), and entire study area (c). Positive values correspond to fluxes towards the atmosphere. The boxes inside each panel correspond to the annual carbon budgets for the region including the lateral carbon fluxes at the river–estuary interface, as inorganic (IC) and organic carbon (OC). The values in grey represent the uncertainties of the annual fluxes.

ized by lower p_{CO_2} values and subsequent CO₂ exchange than the other systems in the region. On the other hand, our regional outgassing of $50 \text{ gC m}^{-2} \text{ yr}^{-1}$ is already well below the global average of $218 \text{ gC m}^{-2} \text{ yr}^{-1}$ calculated using the same approach by Laruelle et al. (2013) for tidal estuaries. The continental shelf CO₂ sink is strongest in January ($-0.47 \pm 0.30 \text{ TgC month}^{-1}$) and decreases until June, when a period of moderate CO₂ emission begins (max of $0.13 \pm 0.08 \text{ TgC month}^{-1}$ in August) and lasts until October. Finally, November and December are characterized by mild CO₂ sinks. Such seasonal signal, following that of water temperature, is consistent with the hypothesis of a CO₂ exchange in the South section regulated by variations in gas solubility,

Table 2. Surface areas, CO₂ exchange rate with the atmosphere, and surface integrated $F\text{CO}_2$ for the North and South sections of COSCAT 827, subdivided by river discharge classes and continental shelf water depth intervals.

	North			South			Total		
	Surface Area 10 ³ km ²	Rate gCm ⁻² yr ⁻¹	$F\text{CO}_2$ 10 ⁹ gC yr ⁻¹	Surface Area 10 ³ km ²	Rate gCm ⁻² yr ⁻¹	$F\text{CO}_2$ 10 ⁹ gC yr ⁻¹	Surface Area 10 ³ km ²	Rate gCm ⁻² yr ⁻¹	$F\text{CO}_2$ 10 ⁹ gC yr ⁻¹
Rivers									
Q1 ($Q < 1 \text{ m s}^{-1}$)	0.14	2893 ± 521	391 ± 70	0.27	1961 ± 353	532 ± 96	0.41	2271 ± 409	924 ± 166
Q2 ($1 \text{ m s}^{-1} < Q < 10 \text{ m s}^{-1}$)	0.21	2538 ± 457	525 ± 95	0.32	1570 ± 283	506 ± 91	0.53	1948 ± 351	1032 ± 186
Q3 ($10 \text{ m s}^{-1} < Q < 100 \text{ m s}^{-1}$)	0.16	1476 ± 267	237 ± 43	0.30	1307 ± 235	392 ± 71	0.46	1366 ± 246	629 ± 113
Q4 ($100 \text{ m s}^{-1} < Q$)	0.17	891 ± 160	152 ± 27	0.36	729 ± 131	261 ± 47	0.52	781 ± 141	412 ± 74
Sub-total	0.67	1939 ± 349	1305 ± 235	1.25	1351 ± 243	1692 ± 305	1.92	1557 ± 280	2997 ± 539
Estuaries	0.53	50 ± 31	27 ± 19	14.51	50 ± 31	731 ± 453	15.04	50 ± 31	758 ± 469
Shelf									
S1 (depth < 20 m)	11.21	5 ± 1	53 ± 19	24.28	-3 ± 1	-79 ± 11	35.49	-1 ± 1	-27 ± 5
S2 (20 m < depth < 50 m)	26.25	-1 ± 1	-35 ± 12	63.88	-8 ± 1	-521 ± 70	90.13	-6 ± 1	-556 ± 108
S3 (50 m < depth < 80 m)	39.28	-3 ± 1	-128 ± 45	48.63	-7 ± 1	-359 ± 126	87.91	-6 ± 1	-488 ± 95
S4 (80 m < depth < 120 m)	60.69	-3 ± 1	-209 ± 73	25.18	-8 ± 1	-199 ± 27	85.87	-5 ± 1	-409 ± 80
S5 (120 m < depth < 150 m)	34.73	-4 ± 1	-151 ± 18	7.63	-12 ± 1	-91 ± 12	42.36	-6 ± 1	-242 ± 47
Sub-total	172.17	-3 ± 1	-472 ± 166	169.59	-7 ± 1	-1250 ± 169	341.77	-5 ± 1	-1722 ± 335

as suggested by Degrandpré et al. (2002) for the Mid-Atlantic Bight.

The analysis of the intensity of the river CO₂ outgassing reveals that the smallest streams ($Q < 1 \text{ m}^3 \text{ s}^{-1}$, Q1 in Table 2) display the highest emission rates per unit surface area, with values ranging from 1961 gC m⁻² yr⁻¹ in the South section to 2893 gC m⁻² yr⁻¹ in the North section. These values gradually decrease with increasing river discharge to 729 gC m⁻² yr⁻¹ in the South section and 891 gC m⁻² yr⁻¹ in the North section for $Q > 100 \text{ m}^3 \text{ s}^{-1}$ (Q4, Table 2). The emission rates for this latter class of rivers are consistent with the median emission rate of 720 gC m⁻² yr⁻¹ proposed by Aufdenkampe et al. (2011) for temperate rivers with widths larger than 60–100 m. Aufdenkampe et al. (2011) also report a median emission rate of 2600 gC m⁻² yr⁻¹ for the smaller streams and rivers, which falls on the high end of the range calculated for Q1 in the present study. The surface area of the river network is relatively evenly distributed amongst the four discharge classes of rivers (Table 2). Yet, river sections for which $Q < 10 \text{ m}^3 \text{ s}^{-1}$ (Q1+Q2) contribute to 65 % of the total CO₂ outgassing although they only represent 51 % of the surface area. This result therefore highlights that streams and small rivers are characterized by the highest surface-area-specific emission rates. The higher outgassing rates in the North section are a consequence of higher ΔCO_2 values since average k values are similar in both sections. In rivers with $Q_{\text{ann}} < 10 \text{ m}^3 \text{ s}^{-1}$, the ΔCO_2 is about twice as high in the North than in the South section from April to August (Table 2). The calculation of $p\text{CO}_2$ from alkalinity and pH presumes however that all alkalinity originates from bicarbonate and carbonate ions and thus tends to overestimate $p\text{CO}_2$ because non-carbonate contributions to alkalinity, in particular organic acids, are ignored in this approach. The rivers in Maine and New Brunswick, which drain most of the northern part of COSCAT 827, are characterized by relatively low

mineralized, low pH waters, rich in organic matter. In these rivers, the overestimation in $p\text{CO}_2$ calculated from the alkalinity attributed to the carbonate system only was reported to be in the range of 13–66 % (Hunt et al., 2011). Considering that rivers in the southern part of COSCAT 827 have lower DOC concentrations and higher dissolved inorganic carbon (DIC) concentrations, the higher $F\text{CO}_2$ rates per surface water area reported in the northern part could partly be due to an overestimation of their $p\text{CO}_2$ values. However, a direct comparison of average $p\text{CO}_2$ values does not confirm this hypothesis. For the two Maine rivers (Kennebec and Androscoggin rivers), Hunt et al. (2014) report an average $p\text{CO}_2$ calculated from pH and DIC of 3064 μatm . In our data set, three sampling stations are also located in these rivers and present lower median $p\text{CO}_2$ values of 2409, 901 and 1703 μatm for Kennebec River at Bingham and North Sidney and for Androscoggin River at Brunswick, respectively. A probable reason for the discrepancy could be that we report median values per month while Hunt et al. (2014) report arithmetic means, which are typically higher.

On the continental shelf, the shallowest depth interval is a CO₂ source in the North section while all other depth intervals are CO₂ sinks (Table 2). The magnitude of the air–sea exchange for each segment is between the values calculated for estuaries (50 gC m⁻² yr⁻¹) and the nearby open ocean ($\sim 20 \text{ gC m}^{-2} \text{ yr}^{-1}$, according to Takahashi et al., 2009). This trend along a depth transect, suggesting a more pronounced continental influence on nearshore waters and a strengthening of the CO₂ shelf sink away from the coast was already discussed in the regional analysis of Chavez et al. (2007) and by Jiang et al. (2013), specifically for the South Atlantic Bight. Modeling studies over a larger domain including the upper slope of the continental shelf also suggest that the coastal waters of the Northeast US are not a more intense CO₂ sink than the neighboring open ocean (Fennel

and Wilkin, 2009; Fennel, 2010). Our analysis further suggests that the continental influence is more pronounced in the North section. Here, the shallowest waters (S1) are strong net sources of CO₂ while the intensity of the CO₂ sink for the other depth intervals gradually decreases, but only to a maximum value of $-4 \text{ gC m}^{-2} \text{ yr}^{-1}$ for S5. This value is about 3 times smaller than in the South section ($-12 \text{ gC m}^{-2} \text{ yr}^{-1}$).

Annually, river and estuarine waters of the entire COSCAT 827 outgas 3.0 ± 0.5 and $0.8 \pm 0.5 \text{ TgC yr}^{-1}$, respectively, while continental shelf waters take up $1.7 \pm 0.3 \text{ TgC yr}^{-1}$ (Fig. 3c). The total riverine carbon load exported from rivers to estuaries for the same area has been estimated to be 4.65 TgC yr^{-1} , 45 % as dissolved and particulate organic carbon (2.10 TgC yr^{-1} , Mayorga et al., 2010) and 55 % as dissolved inorganic carbon (2.55 TgC yr^{-1} , Hartmann et al., 2009). The ratio of organic to inorganic carbon in the river loads is about 1 in the North section and 1.4 in the South section. This difference stems mainly from a combination of different lithogenic characteristics in both sections and the comparatively higher occurrence of organic soils in the North section (Hunt et al., 2013; Hossler and Bauer, 2013). Estimates of the total amount of terrestrial carbon transferred to the riverine network are not available, but the sum of the river export and the outgassing, which ignores the contribution of carbon burial and lateral exchange with wetlands, provides a lower bound estimate of 7.65 TgC yr^{-1} . Under this hypothesis, $\sim 40\%$ of the terrestrial carbon exported to rivers is emitted to the atmosphere before reaching estuaries. In spite of higher emission rates per unit surface area in the North section (Table 2), the overall efficiency of the riverine carbon filter is essentially the same in the two sections (40 and 38 % outgassing for the North and the South sections, respectively). On the shelf, however, the South section exhibits a significantly more intense CO₂ sink ($-1.25 \pm 0.2 \text{ TgC yr}^{-1}$) than in the North section ($-0.47 \pm 0.2 \text{ TgC yr}^{-1}$). A possible reason for this difference can be found in the contribution of the estuarine carbon filter. In the South section, where 96 % of the estuarine surface area is located, these systems contribute to an outgassing of 0.73 TgC yr^{-1} while in the North section, their influence is negligible. Cole and Caraco (2001) estimated that 28 % of the DOC entering the relatively short Hudson River estuary is respired in situ before reaching the continental shelf and it is thus likely that the estuarine outgassing in the South section is fueled by the respiration of the organic carbon loads from rivers. In contrast, the absence of estuaries in the North section favors the direct export of terrestrial organic carbon onto continental shelf waters where it can be buried and decomposed. The respiration of terrestrial organic carbon could therefore explain why the strength of the shelf CO₂ sink is weaker in this portion of the domain. Such filtering of a significant fraction of the terrestrial carbon inputs by estuaries has been evidenced in other systems (Amann et al., 2010; 2015). This view is further substantiated by the similar cumulated estuarine and continental shelf FCO_2 fluxes in both sections (Fig. 3a and b). Naturally, other

environmental and physical factors also influence the carbon dynamics in shelf waters and contribute to the difference in CO₂ uptake intensity between both sections. For instance, in the North section, the Gulf of Maine is a semi-enclosed basin characterized by specific hydrological features and circulation patterns (Salisbury et al., 2008; Wang et al., 2013) which could result in longer water residence times promoting the degradation of shelf-derived organic carbon. Other potential factors include the plume of the Saint Lawrence Estuary, which has also been shown to transiently expand over the Scotian Shelf (Kang et al., 2013), the strong temperature gradient, and the heterogeneous nutrient availability along the region which may result in different phytoplankton responses (Vandemark et al., 2011; Shadwick et al., 2011). Additionally, modeling studies evidenced the potential influence of sediment denitrification on water pCO_2 through the removal of fixed nitrogen in the water column and consequent inhibition of primary production (Fennel et al., 2008; Fennel, 2010). This removal was estimated to be of similar magnitude as the lateral nitrogen loads, except for estuaries of the Mid-Atlantic Bight (MAB) region (Fennel, 2010). It can nonetheless be suggested that the estuarine carbon filter in the South section of COSCAT 827 is an important control factor of the CO₂ sink in the Mid-Atlantic Bight, which is stronger than in any other area along the entire Atlantic coast of the US (Signorini et al., 2013).

4 Conclusions

Our data-driven spatially and seasonally resolved budget analysis captures the main characteristics of the air–water CO₂ exchange along the LOAC of COSCAT 827. It evidences the contrasting dynamics of the North and South sections of the study area and an overall gradual shift from a strong source in small streams oversaturated in CO₂ towards a net sink in continental shelf waters. Our study reveals that ice and snow cover are important controlling factors of the seasonal dynamics of CO₂ outgassing in streams and rivers and account for a large part of the difference between the North and South sections. The close simultaneity of the snowmelt on land and of the phytoplankton bloom on the continental shelf leads to opposite temporal dynamics in FCO_2 in these two compartments of the LOAC. In addition, our results reveal that estuaries filter significant amounts of terrestrial carbon inputs, thereby influencing the continental shelf carbon uptake. Although this process likely operates in conjunction with other regional physical processes, it is proposed that the much stronger estuarine carbon filter in the South section contributes to a strengthening of the CO₂ sink in the adjacent continental shelf waters.

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