



Royal Netherlands Institute for Sea Research

This is a pre-copyedited, author-produced version of an article accepted for publication, following peer review.

Kraal, P.; van Genuchten, C.M.; Lenstra, W.K. & Behrends, T. (2020). Coprecipitation of phosphate and silicate affects environmental iron (oxyhydr)oxide transformations: A gel-based diffusive sampler approach. *Environmental Science and Technology*, 54, 12795-12802

Published version: <https://dx.doi.org/10.1021/acs.est.0c02352>

NIOZ Repository: <http://imis.nioz.nl/imis.php?module=ref&refid=330302>

[Article begins on next page]

The NIOZ Repository gives free access to the digital collection of the work of the Royal Netherlands Institute for Sea Research. This archive is managed according to the principles of the [Open Access Movement](#), and the [Open Archive Initiative](#). Each publication should be cited to its original source - please use the reference as presented.

When using parts of, or whole publications in your own work, permission from the author(s) or copyright holder(s) is always needed.

1 Coprecipitation of phosphate and silicate affects
2 environmental iron (oxyhydr)oxide transformations:
3 a gel-based diffusive sampler approach

4 *Peter Kraal*^{†§}, Case M. van Genuchten[†], Wytze K. Lenstra[†], Thilo Behrends[†]*

5 [†] Department of Earth Sciences – Geochemistry, Faculty of Geosciences, Utrecht University, PO
6 Box 80021, 3508 TA Utrecht, The Netherlands

7 [§] Royal Netherlands Institute for Sea Research, Department of Ocean Systems, and Utrecht
8 University, PO Box 59, 1790 AB Den Burg, The Netherlands

9 * Corresponding author, peter.kraal@nioz.nl

10 Keywords: ferrihydrite, lepidocrocite, crystallization, sulfidation, DGT-DET, eutrophication

ABSTRACT

Sorption of nutrients such as phosphate (P) and silicate (Si) by ferric iron (oxyhydr)oxides (FeOx) modulates nutrient mobility and alters the structure and reactivity of the FeOx. We investigated the impact of these interactions on FeOx transformations using a novel approach with samplers containing synthetic FeOx embedded in diffusive hydrogels. The FeOx were prepared by Fe(III) hydrolysis and Fe(II) oxidation, in the absence and presence of P or Si. Coprecipitation of P or Si during synthesis altered structure of Fe precipitates and, in case of Fe(II) oxidation, lepidocrocite was (partly) substituted by poorly ordered FeOx. The pure and P- or Si-bearing FeOx were deployed in (i) freshwater sediment rich in dissolved Fe(II) and P and (ii) marine sediment with sulfidic pore-water. Iron(II)-catalyzed crystallization of poorly ordered FeOx was negligible, likely due to surface passivation by adsorption of dissolved P. Reaction with dissolved sulfide was modulated by diffusion limitations and therefore the extent of sulfidation was lowest for poorly ordered FeOx with high reactivity towards sulfide that created temporary, local sulfide depletion ($F_h < L_p$). We show that coprecipitation-induced changes in FeOx structure affect coupled iron-nutrient cycling in aquatic ecosystems. The gel-based method enriches our geochemical toolbox by enabling detailed characterization of target phases under natural conditions.

INTRODUCTION

Poorly crystalline ferric iron (oxyhydr)oxides (FeOx) have high sorption capacity and therefore modulate the mobility and bioavailability of (trace) contaminants and nutrients in aquatic systems. Furthermore, laboratory studies have shown that adsorption or coprecipitation of such species can alter the transformation and dissolution of FeOx such as ferrihydrite (Fh) and lepidocrocite (Lp). Strong effects have been observed in laboratory studies for nutrients such as phosphate (PO_4 , denoted here as P) and silicate (SiO_4 , denoted here as Si).¹⁻⁶

Occupation of FeOx surface sites by oxyanions such as PO_4 or SiO_4 during Fe precipitate formation decreases structural order. The extent of the impact depends on the molar oxyanion/Fe ratio and the FeOx formation pathway. Sorption of oxyanions during Fe(III) hydrolysis and Fe(II) oxidation disrupts Fe-Fe polymerization and under high oxyanion/Fe ratios, amorphous precipitates such as Fe(III) phosphate can form.⁷⁻⁸ Furthermore, in the case of Fe(II) oxidation, sorption of oxyanions inhibits Fe(II)-catalyzed transformation of freshly formed nano-scale FeOx into more crystalline Fe phases.^{3, 7-10} The environmental implications of these FeOx alterations are poorly known. In particular, it is unclear how key FeOx transformation pathways such as crystallization and sulfidation are affected under natural conditions. These transformations alter the nutrient and contaminant sorption capacity of Fe minerals and represent key processes in the coupling between nutrient dynamics and Fe redox cycling and transformation.¹¹⁻¹³

Detailed tracking of the environmental fate of target Fe phases can be complicated by interference from a large and diverse pool of background Fe in natural sediments. Laboratory-based batch and column experiments¹⁴⁻¹⁸ have provided detailed insight into FeOx transformations but may not adequately represent natural conditions, in part because of the use of FeOx as single solid phase or synthetic sediment. Here, we aim to advance our understanding of the environmental

fate of Fe by adapting an established pore-water method (diffusive gradients/equilibrium in thin films, DGT-DET)¹⁹⁻²⁰ to deploy well-characterized (P-bearing) FeOx, embedded in diffusive hydrogels, in natural sediments. We show that the target FeOx undergo geochemical transformations (with a prominent role for P and Si coprecipitation) that can be investigated in (micro-scale) detail because the hydrogels only contain FeOx and its transformation products.

MATERIALS AND METHODS

Synthesis of iron precipitates. Two-line ferrihydrite (Fh; $\text{Fe(II)}_{10}\text{O}_{14}(\text{OH})_2 \cdot n\text{H}_2\text{O}$ with $n < 1$)²¹⁻²² and lepidocrocite (Lp; $\gamma\text{-Fe(III)OOH}$), referred to here as ‘pure’ FeOx, were prepared following established protocols using Fe(III) hydrolysis for Fh synthesis and Fe(II) oxidation for Lp synthesis at circumneutral pH (7–8).²³ Base hydrolysis of a concentrated Fe(III) solution is a widely used method to synthesize Fh-type precipitates representative of poorly ordered FeOx in natural soils and sediments. Such precipitates form in anthropogenic landscapes by neutralization of acidic, Fe(III)-rich drainage from acid sulfate soils and mine tailings, where the FeOx play an important role in scavenging contaminants.²⁴⁻²⁷ Lepidocrocite from oxidation of a concentrated Fe(II) solution serves as analog for more crystalline natural FeOx formed around redox interfaces in natural soils and sediments where initial poorly ordered FeOx precipitates undergo further crystallization by reaction with Fe(II).^{13, 28-29}

In addition to pure FeOx, we prepared FeOx with coprecipitated P or Si at a molar (P, Si)/Fe ratio of 0.1. A P/Fe ratio of 0.1 is representative of the ratio between Fe(III) and associated P in oxic surface sediments.³⁰⁻³² Furthermore, previous studies have elucidated in detail the impact of P and Si coprecipitation on FeOx structure at similar oxyanion/Fe ratios.^{3, 7, 10} The FeOx prepared by Fe(III) hydrolysis and Fe(II) oxidation are referred to as FeOx_{hydr} and FeOx_{oxid}, respectively.

Details on FeOx synthesis are given in the Supporting Information and have been reported previously.⁶

Preparation of (FeOx-bearing) hydrogels and samplers. Several types of (bis)acrylamide-cross linked hydrogels were prepared. Using existing protocols, DET (diffusive equilibrium in thin films¹⁹) hydrogels were prepared for measurement of pore-water Fe(II)³³ and P³⁴ and DGT (diffusive gradients in thin films) binding hydrogels were prepared for pore-water sulfide.³⁵ Using an adapted protocol (see Supporting Information), FeOx-bearing thin film (FeOx-BT) hydrogels were prepared by adding small aliquots of FeOx in suspension during hydrogel casting. Diffusive samplers for pore-water analysis or FeOx deployment were prepared by placing DGT-DET or FeOx-BT hydrogels, covered with filter membrane, in rigid TeflonTM samplers (DGT Research, Lancaster) as detailed in the Supporting Information. The FeOx-BT samplers allow diffusive exchange with pore-water but prevent direct physical contact between the target FeOx and solid-phase sediment (i.e. contamination).

Deployment of hydrogel samplers. Samplers were deployed during the summer period (June - September) of 2016 at two contrasting sites in The Netherlands. Site 1 was a freshwater drainage ditch on the Utrecht University campus (52°5.27' N, 5°9.80' E) with clayey to sandy sediment rich in dissolved Fe(II) and P. Site 2 was a tidal mud flat with sulfidic silty to sandy sediment in the marine Oosterschelde estuary (salinity 28 PSU) in the southwest of The Netherlands (51°26.35' N, 4°9.87' E). Site 1 was used to investigate Fe(II)-catalyzed transformation of Fh and Site 2 was used to study Fh and Lp sulfidation. At each site, DGT-DET and FeOx-BT samplers were

introduced into the sediment. Pore-water samplers were left overnight, the FeOx were deployed for four weeks. Details on deployment and recovery of the samplers are given in the Supporting Information. We conducted an additional short-term FeOx sulfidation experiment under controlled conditions. For this laboratory-based experiment, a microcosm of surface sediment with 5 cm of overlying water was created using mixed surface sediment (top 10 cm) and unfiltered seawater from Site 2. These benthic microcosms were left to equilibrate at room temperature for a week while the overlying water was kept O₂-saturated by bubbling air through it. Subsequently, pore-water samplers and FeOx-BT samplers were inserted and left overnight.

In the field and laboratory experiments, the DGT-DET pore-water samplers were deployed in sets of two to measure ΣH₂S (DGT) and Fe (DET) or ΣH₂S (DGT) and P (DET). During the field experiments, we deployed duplicate samplers for each type of FeOx-BT hydrogel were deployed. In the laboratory sulfidation experiment, single samplers for each type of FeOx-BT hydrogel were used.

Processing of DET-DGT and FeOx-BT hydrogels. The pore-water DGT and DET hydrogels were analyzed as described previously (also see Supporting Information). Briefly, DET hydrogels that had equilibrated with pore-water were overlain with a DET hydrogel that had equilibrated with a solution containing reagent for spectrophotometric detection of either Fe(II) (FerrozineTM) or P (molybdate blue). Color formation by diffusive exchange between the sample and staining gels was quantified and converted into dissolved Fe(II) or P concentration.³³⁻³⁴ In DGT gels, sulfide is bound as black-colored Ag₂S which stains the initially white gel. From the intensity of this color change, the dissolved sulfide concentration can be calculated.³⁵

The FeO_x-BT hydrogels were removed from the samplers in an Ar-filled anoxic glovebox. One of the duplicate gels was immediately cut horizontally in four equal pieces of ~ 3.5 cm length and 1.5 cm width. These gel cuts were shortly stored in anoxic milliQ (with two water changes to remove any dissolved components) prior to further preparation for X-ray analysis. The second duplicate gel was stored intact in Ar-purged anoxic milliQ. A selection of these duplicate gels (Fh, P9-FeO_x_{hydr}, Si9-FeO_x_{hydr}) were cut as described above and used for micro-scale XRF (μ XRF) scanning.

Micro-scale XRF and synchrotron-based XAS analysis of hydrogel sections. Detailed descriptions of sample preparation and analysis for micro-XRF and XAS analysis are given in the Supporting Information. Briefly, for μ XRF analysis, selected duplicate wet gel cuts were placed on microscope slides and vertical line scans were collected at 30 μ m resolution for Fe, P and S. For synchrotron-based X-ray absorption spectroscopy (XAS), samples were freeze-dried, ground and mixed with cellulose and pressed into pellets that were used for Fe K-edge XAS analysis at beamline BM26A (DUBBLE) of the European Synchrotron Radiation Facility (ESRF) in Grenoble, France. Details on the beamline setup can be found in Nikitenko et al.³⁶ Spectra were recorded at room temperature in transmission mode from 6962 to 7862 eV, corresponding to a maximum k of 13 \AA^{-1} in the EXAFS region. For each sample, two spectra were collected and merged. Repeated measurement of an FeS-rich sample showed no sign of oxidation or beam damage during analysis. Data normalization was done using the Athena software package.³⁷ Iron speciation was obtained by linear combination fitting of components (representing Fe phases) extracted from the entire XAS dataset (24 spectra including references) by iterative transform factor analysis using the ITFA software package.³⁸ Analysis was performed in k space, using a

fitting window of $k = 2 - 11 \text{ \AA}^{-1}$ of the k^3 weighted EXAFS spectra. For sulfidation samples, the degree of Fe sulfidation (DOS)³⁹ was determined as the fraction of total Fe transformed into iron (mono)sulfide based on the results from the LCF.

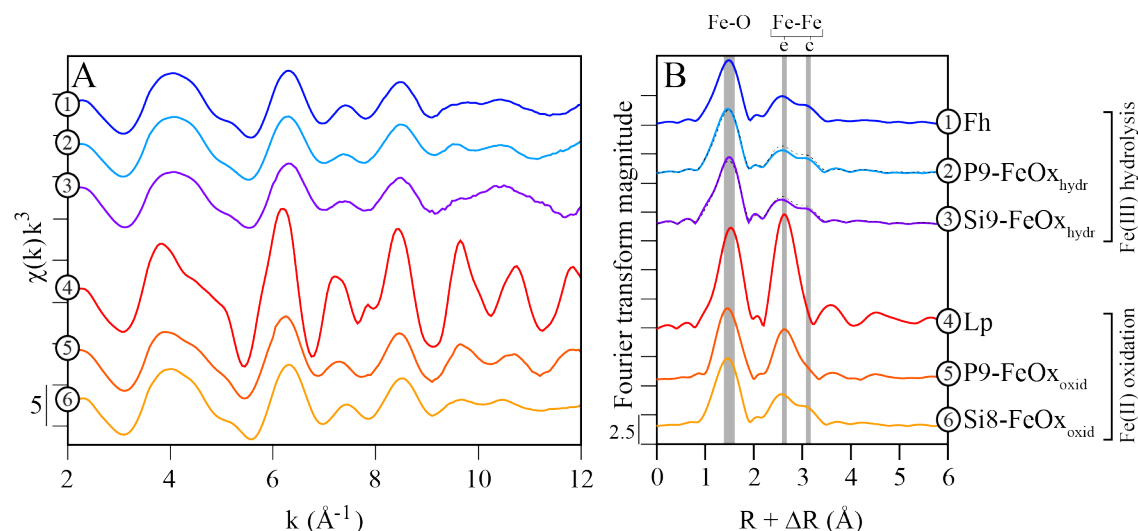


Figure 1. (A) Fe K-edge EXAFS spectra and (B) Fourier-transformed Fe K-edge EXAFS spectra of FeOx synthesized for this study. Vertical grey bands in (B) mark magnitude peaks arising from Fe - O bonds and edge-sharing (e) and corner-sharing (c) Fe octahedra. In (B), dashed black lines plotted with spectrum 2 and 3 show the Fh spectrum for reference.

RESULTS AND DISCUSSION

Structure of the initial FeOx. Hydrolysis of concentrated Fe(III) solutions formed Fh-type precipitates (Fig. 1). During Fe(III) hydrolysis in the presence of 10 mol% dissolved P or Si relative to Fe, the poorly ordered FeOx that formed contained 9 mol% P (P9-FeOx_{hydr}) or Si (Si9-FeOx_{hydr}). The FeOx_{hydr} with coprecipitated P and Si showed slightly less amplitude in the second and third peak in the Fourier-transformed (FT) spectra ($R + \Delta R = 2.4$ and 3.1 \AA) that represent edge- and corner-sharing Fe octahedra. This may reflect inhibition of Fe polymerization and formation of aggregates of Fe oligomers linked by P or Si. This impact of strongly-sorbing oxyanions such as phosphate, silicate and arsenate on FeOx formation from Fe(III) hydrolysis has been described in

detail in previous work.^{6, 9-10} The small decrease in peak amplitude under the current experimental conditions at P/Fe = 0.1 suggests that inhibition of Fe polymerization was minor under the current experimental conditions.

For Fe precipitates from Fe(II) oxidation, the impact of P or Si coprecipitation at an initial P/Fe ratio of 0.1 (10 mol% P relative to Fe) was more profound. In the absence of dissolved P or Si, pure Lp formed because freshly formed nano-scale FeOx undergoes rapid Fe(II)-catalyzed transformation.²³ The Lp showed a strongly developed second-shell Fe peak at ~2.6 Å (uncorrected for phase shift) in the FT spectrum arising from edge-sharing Fe(III) octahedra (Fig. 1B). The FT spectrum of P9-FeOx_{oxid} (9 mol% P relative to Fe) was markedly different with decreased amplitudes of the Fe-O and edge-sharing Fe peaks. The P9-FeOx_{oxid} spectrum represents a mixture of Fe phases, resulting from sequential formation of Fe precipitates by Fe(II) oxidation in the presence of dissolved P.⁷⁻⁸ Initially, highly disordered Fe(III)-P polymers and/or P-rich hydrous ferric oxide (P-HFO) form and rapidly deplete dissolved P. The Fe(III) that is subsequently formed under P-depleted conditions partly contributes to polymerization of Fe(III)-P into P-HFO and partly precipitates as Lp. Previous work indicates that under the current experimental conditions (P/Fe = 0.1), Fe precipitates most likely consist of a mixture of highly disordered P-HFO and Lp.⁸ The LCF results showed that P9-FeOx_{oxid} could be reproduced as a combination of P-HFO (70 %) and Lp (30%). Assuming all P was associated with the P-HFO fraction, this precipitate would have a molar P/Fe ratio of ~ 0.13. The fraction of P-HFO was likely overestimated in the LFC procedure and therefore the P/Fe of the P-HFO would have been underestimated.

Dissolved Si has a lower affinity for Fe(III) and does not become depleted but rather coprecipitates with FeOx precipitates throughout FeOx synthesis, thereby shifting the reaction product of Fe(II) oxidation from Lp to Si-HFO.^{3, 7} Our results showed a Fe K-edge EXAFS spectra

of Si8-FeOx_{oxid} (8 mol% Si relative to Fe) that was similar to that of pure Fh (Fig. 1), with a more developed corner-sharing Fe peak compared to P9-FeOx_{oxid} (Fig. 1B). The PCA-ITFA results indicated that Si8-FeOx_{oxid} consisted of 100% poorly ordered Si-HFO. The ITFA-LCF results are presented in detail in the Supporting Information.

Transformation of FeOx in freshwater sediment. The ITFA results for the entire Fe K-edge EXAFS dataset indicated that all sample spectra, including starting materials, samples from freshwater Site 1 and samples from the sulfidation experiment, could be reproduced by three independent components that represented Fh, Lp and FeS (see Supporting Information).

The pore-water in the freshwater drainage ditch was rich in Fe(II) (up to 600 $\mu\text{mol L}^{-1}$) and P (up to 566 $\mu\text{mol L}^{-1}$) (Figure 2A). Dissolved sulfide was not detected with the DGT sulfide-binding gel. Due to spectral similarity, all spectra of the hydrogel samples from Site 1 were reproduced in ITFA-LCF by one component — poorly ordered FeOx. To further explore potential minor changes in FeOx structure, we also performed LCF on these samples in Athena³⁷ using Fh and P9-FeOx_{hydr} or Si9-FeOx_{hydr} as references. Preliminary LCF had shown that candidate references such as lepidocrocite, goethite, hematite, magnetite, vivianite, siderite were all negligible. There were also no observable shifts in the Fe-K absorption edge (Fig. 2B), further refuting formation of reduced Fe phases in the hydrogels, even though chemical modelling in Visual Minteq 3.1 (settings: 500 $\mu\text{mol L}^{-1}$ Fe(II) and P, pH 6.8, I = 0.01) indicated that the pore-water was oversaturated (saturation index \approx 8) with respect to vivianite. The LCF results indicated a shift from Si9-FeOx_{hydr} and P9-FeOx_{hydr} to Fh during deployment of the amended FeOx_{hydr} (Fig. 2B), suggesting a minor increase in short-range structure. It is possible that ageing during storage of the FeOx suspensions contributed to this change.

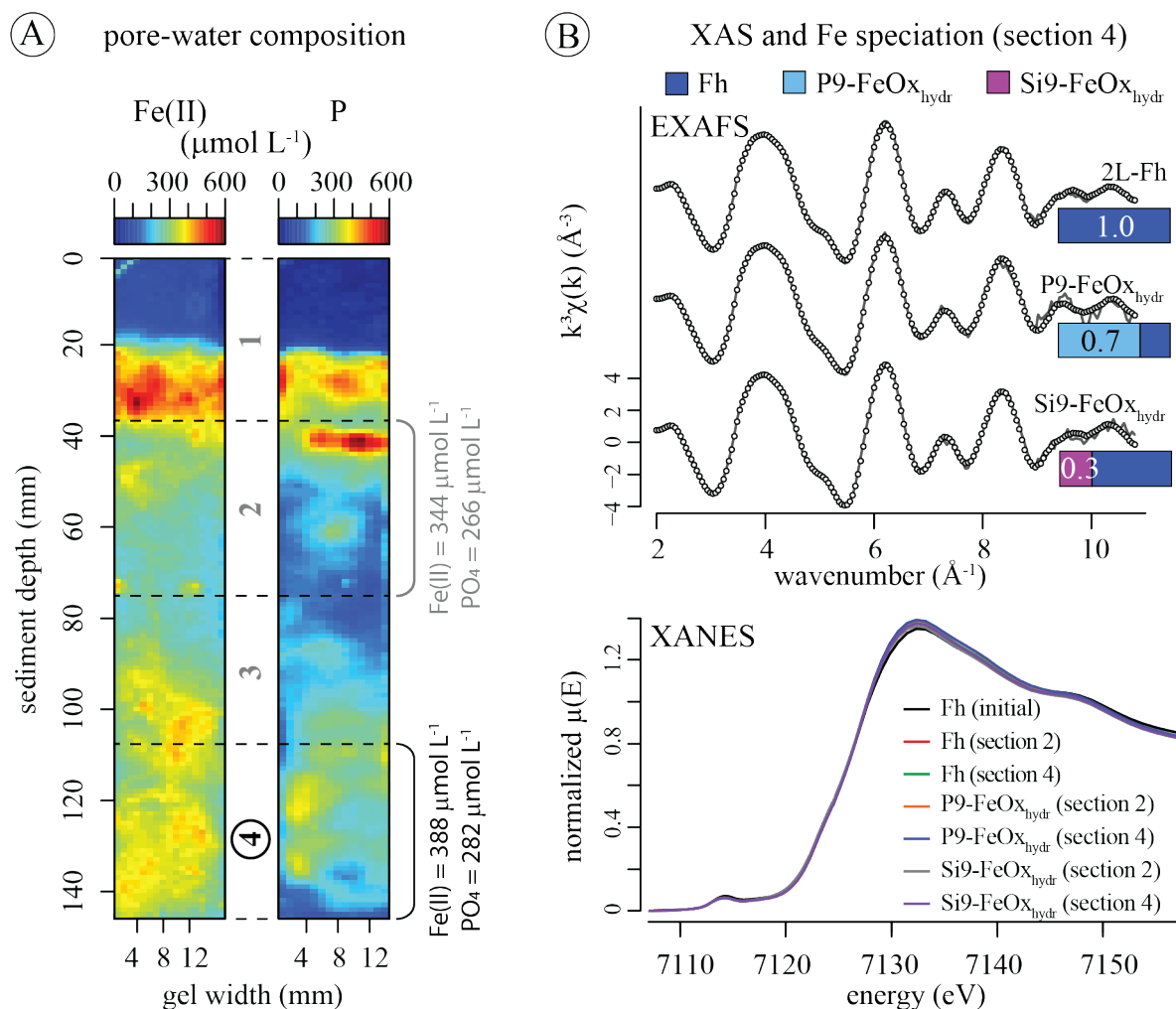


Figure 2. (A) Pore-water concentrations of Fe(II) and P in the freshwater sediment from Site 1. (B) Upper panel: data (lines) and fits (dots) for bulk Fe K-edge EXAFS spectra for selected FeOx-BT hydrogel sections after deployment (also see Fig. S6 and Table S1). Bar charts show LCF results. Numbers in the bars show fraction of starting FeOx remaining. Lower panel: XANES spectra of initial Fh and Fh, P9-FeOx_{hydr} and Si9-FeOx_{hydr} after deployment.

The limited extent of FeOx crystallization found here contrasts with rapid and extensive Fe(II)-catalyzed transformation of poorly ordered FeOx observed in laboratory studies.^{14-15, 40-42} For example, Hansel et al.¹⁵ observed that in sediment columns with advective flow of artificial groundwater with 100 – 300 $\mu\text{mol L}^{-1}$ Fe(II), 70% of Fh was transformed into Lp and Gt within hours. Transformation will likely be more rapid under advective flow compared to our diffusion-

limited experiment with passive samplers in natural sediment, but significant FeOx transformation was expected during a four-week deployment in Fe(II)-rich sediment. Our observations are best explained by the difference in solution chemistry in the sediments relative to aforementioned laboratory studies. The sediments we investigated were characterized by a high dissolved P concentration ($> 250 \mu\text{mol L}^{-1}$) and are also likely to contain dissolved organic carbon. By binding to FeOx surfaces, P and other strongly-sorbing inorganic and organic species can passivate the particle surface and decrease the rate of FeOx transformation.^{4, 6, 43-44} However, we note that adsorbed species (P and organic carbon) were not determined in the current study.

Previous work showed that the apparently minor structural changes to Fh by coprecipitation of P or Si notably enhance reductive and acidic dissolution rates (i.e. destabilized the FeOx), likely due to altered macroscopic properties such as decreased particle aggregation.^{2, 6} In the current study, FeOx surface passivation by adsorption of P (and potentially other abundant pore-water species) overwhelmed any potential effects on altered FeOx properties from P or Si coprecipitation. Inhibition of FeOx transformation acts to preserve poorly ordered FeOx that have relatively high sorption capacity compared to more crystalline counterparts such as Lp and Gt.

In-situ transformation of FeOx in sulfidic tidal flat sediment. The surface sediment from the tidal flat in the southwest of The Netherlands contained relatively low concentrations ($< 20 \mu\text{mol L}^{-1}$) of dissolved Fe(II) and P compared to the freshwater ditch (see Supporting Information) and showed signs of abundant burrows. During recovery of the samplers, we observed that sediment around the samplers had been eroded to different degrees, resulting in small-scale variability in the sediment depth interval covered by the sampler. Visual inspection of the FeOx-

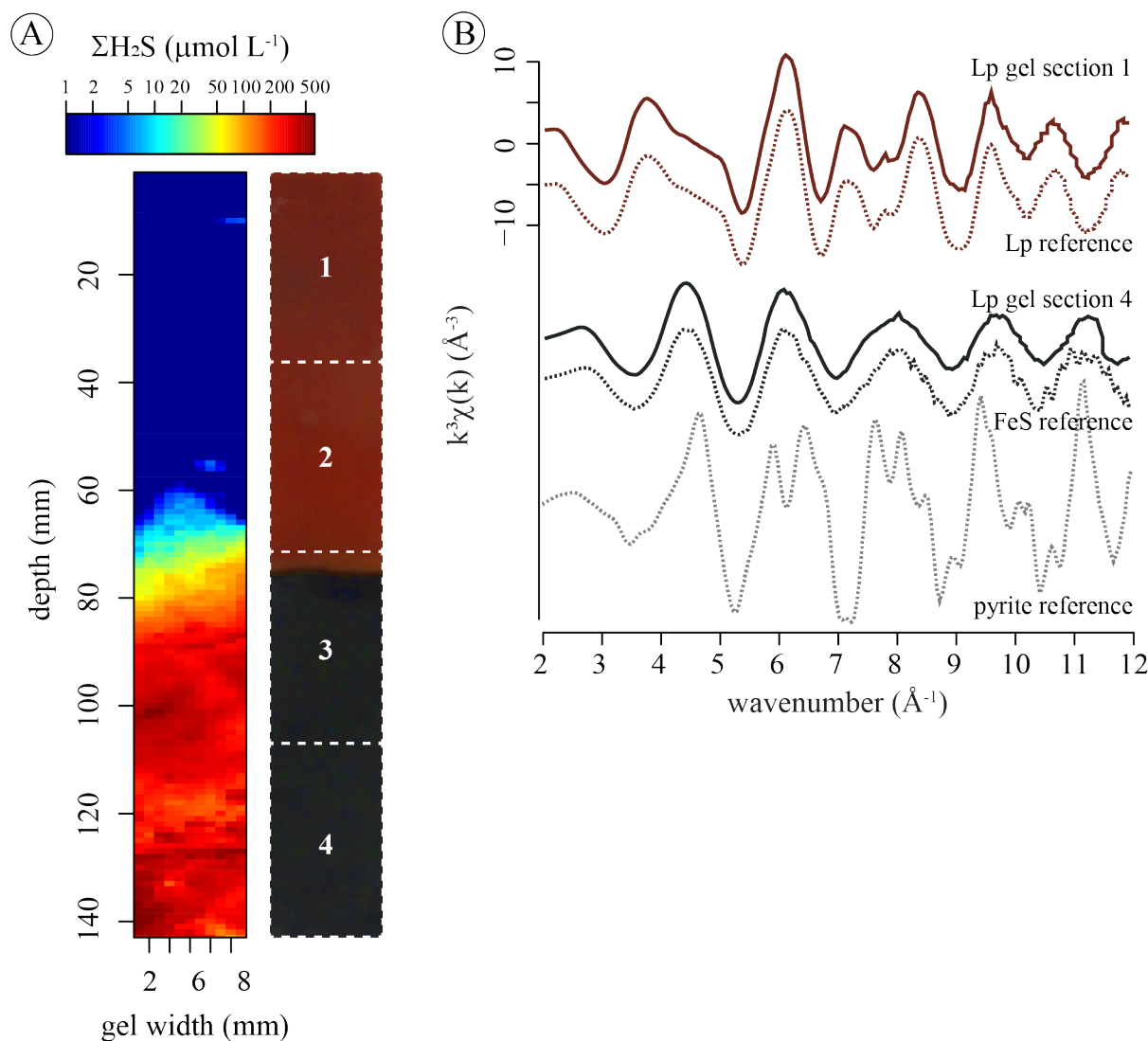


Figure 3. (A) Dissolved $\Sigma\text{H}_2\text{S}$ pore-water map determined by DGT hydrogel (left) and image of Lp hydrogel (right) after 4 weeks of deployment in sediment with high dissolved $\Sigma\text{H}_2\text{S}$ at depth. Dashed lines and white numbers indicate sections into which the gel was cut for bulk Fe speciation. (B) Fe K-edge EXAFS spectra of Lp gel section 1 and 4 plotted with the Lp, FeS and pyrite (FeS_2) reference spectra.

BT hydrogels after recovery showed strong variability in the depth of the sulfidation front, between 2 to 10 cm sediment depth (see Supporting Information). Visual and spectroscopic evidence of FeOx transformation due to sulfidation was observed in all FeOx-BT hydrogels deployed at Site 2. However, the high spatial variability in chemical conditions in the sediment, likely due to tidal

effects and bioturbation, did not allow for systematic analysis of the effect of FeOx structure on degree of sulfidation. Therefore, in this section we focus on individual samplers. The sharp transition from anoxic to sulfidic conditions with dissolved $\Sigma\text{H}_2\text{S}$ concentrations up to $\sim 500 \mu\text{mol L}^{-1}$ as recorded by DGT sampling (which is an average calculated from integrated sulfide accumulation in the DGT gel over the deployment period³⁵) (Fig. 3A, left image) was reflected in a Lp-loaded hydrogel as a sharp color change (Fig. 3A, right image). Synchrotron-based Fe speciation in the top and bottom slices of the Lp hydrogel showed complete transformation of Lp to iron monosulfide (FeS) during deployment in the sulfidic sediment (Fig. 3B). The Fe K-edge XAS results did not show pyrite (FeS_2) formation in the analyzed gel sections from Site 2 after four weeks of deployment. These results highlight the potential for extensive alteration of gel-embedded FeOx. Furthermore, the results are in line with other field studies^{39, 45} that show relatively slow transformation of FeS to pyrite under natural sulfidic conditions compared to laboratory experiments.⁴⁶

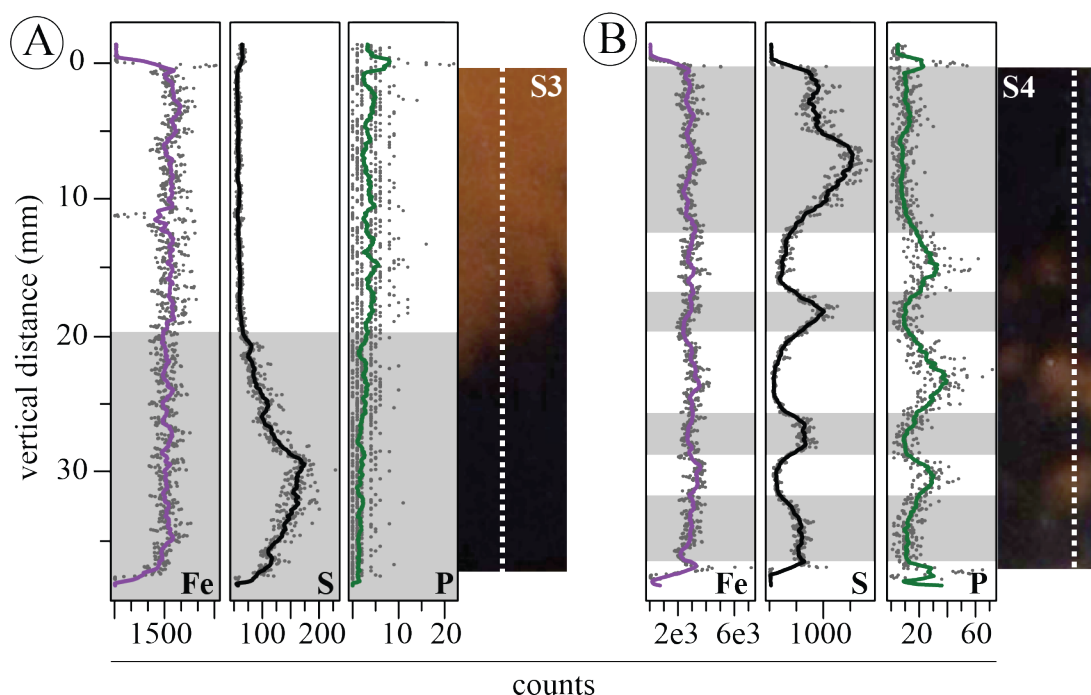


Figure 4. Line scans of iron (Fe), sulfur (S) and phosphorus (P) in sections of a P9-FeOx_{hydr} hydrogel deployed under sulfidic field conditions at Site 2. Images of analyzed gel sections to the right, line scan indicated as dashed white line. (A) Scan across the sulfidation front; (B) Scan across patchy region with sulfidized (dark) and non-sulfidized (light) areas.

To investigate the impact of reaction with dissolved sulfide on FeOx transformation and nutrient dynamics, high-resolution element scans were performed with μ XRF on a P9-FeOx_{hydr}-loaded hydrogel. We selected intervals that captured the sulfidation front (S3, 7–10.5 cm) and a mottled sulfidation pattern (S4, 10.5–14 cm) (Fig. 4). The μ XRF line scan across the sulfidation front showed an increase in solid-phase S content and a concurrent decrease in P (Fig. 4A). A line scan across a patchy area in the bottom gel section (with better vacuum and higher counts) showed an even more clear anti-correlation between S and P (Fig. 4B), reflecting release of P by Fe sulfidation.⁴⁷ In contrast to S and P, Fe counts were stable throughout the gel sections, irrespective of sulfidation. Similar stable Fe counts were observed across sulfidation fronts in hydrogels with Fh and Si9-FeOx_{hydr} (data not shown). Leaching of Fe from FeOx-BT hydrogels was apparently

limited during reaction with dissolved $\Sigma\text{H}_2\text{S}$, in line with sedimentary evidence for efficient Fe retention under sulfidic conditions.⁴⁸

Laboratory-based FeOx sulfidation experiment with natural sediment. Under stagnant conditions at room temperature (21 °C), dissolved $\Sigma\text{H}_2\text{S}$ accumulated with concentrations in excess of 2 mmol L⁻¹ within 5 cm sediment depth (SI section S2.6). Dissolved P showed an increase below ~ 1 cm sediment depth and reached concentrations at depth up to ~ 100 $\mu\text{mol L}^{-1}$. The DET for dissolved Fe(II) was compromised but indicated a similar small peak (< 50 $\mu\text{mol L}^{-1}$) Fe(II) in the non-sulfidic surface sediment (0–2 cm sediment depth) as was observed in the sediment core (see SI for details). We analyzed Fe speciation in mid-depth, ($\Sigma\text{H}_2\text{S}$ 0.8–1.8 mmol L⁻¹) and bottom ($\Sigma\text{H}_2\text{S} \geq 2$ mmol L⁻¹) sections of FeOx-BT hydrogels after overnight deployment. The sample spectra and LCF results can be found in SI section S2.7. The results showed that the degree of sulfidation (DOS, here calculated as $\text{FeS-Fe}/\text{Fe}_{\text{tot}}$) for FeOx from Fe(II) oxidation was consistently higher in bottom sections that were exposed to higher $\Sigma\text{H}_2\text{S}$ concentrations (Fig. 5). For FeOx from Fe(III) hydrolysis, such a correlation between DOS and $\Sigma\text{H}_2\text{S}$ concentration was not observed.

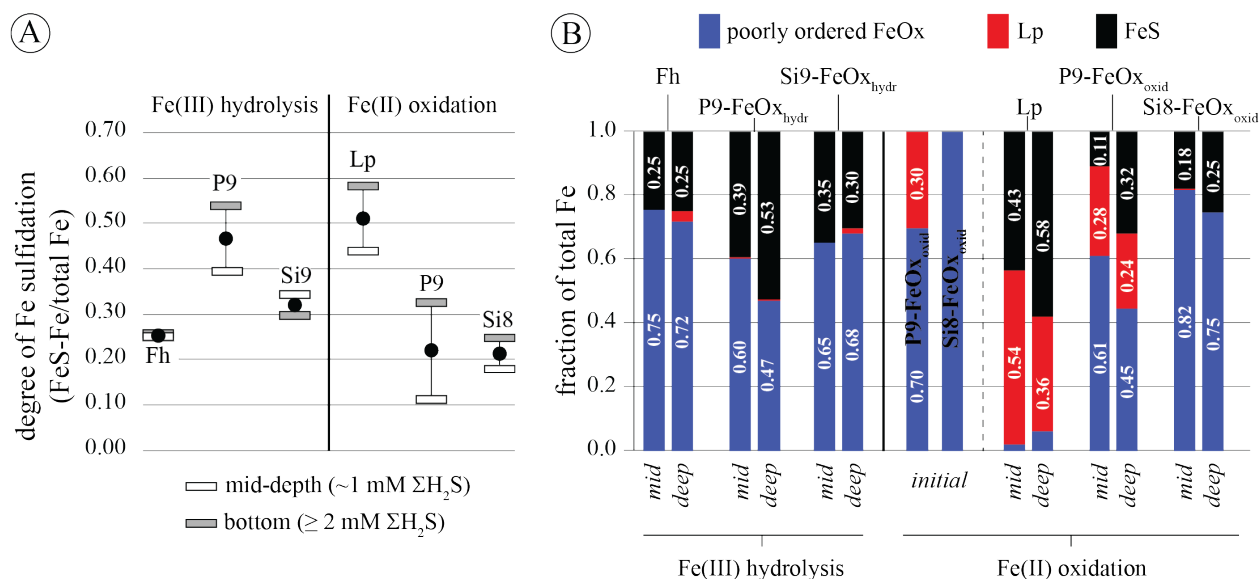


Figure 5. (A) Degree of Fe sulfidation (i.e. FeS as fraction of total Fe) in FeOx-BT hydrogel sections, grouped by FeOx synthesis method (Fe(III) hydrolysis versus Fe(II) oxidation). White and grey bars represent DOS in mid-depth and bottom hydrogel sections, respectively; black dots indicate the mean DOS value ($n = 2$). (B) ITFA-LCF results for the FeOx-BT hydrogels sections. For reference, Fe speciation in the P9-FeOx_{oxid} and Si8-FeOx_{oxid} starting materials is shown. *Mid* and *deep* refer to mid-depth (3.5–7 cm) and bottom (10.5–14 cm) sections of the hydrogels, respectively.

When comparing the pure starting materials Fh and Lp, the DOS was higher for Lp (0.43–0.58) than for Fh (0.25). This may be attributed to relatively rapid $\Sigma\text{H}_2\text{S}$ oxidation by Fh, leading to local depletion of $\Sigma\text{H}_2\text{S}$ and limited FeS formation. The reactivity of Lp towards sulfide is about an order of magnitude slower compared to Fh^{49–50}, limiting $\Sigma\text{H}_2\text{S}$ depletion in and around the Lp-bearing hydrogel. Furthermore, laboratory studies on FeOx sulfidation pathways suggest that Fh-type FeOx can facilitate relatively rapid surface-to-bulk electron transport compared to more crystalline FeOx such as Lp. As a result, Fe(II) formed at the Fh surface by reaction with dissolved $\Sigma\text{H}_2\text{S}$ is re-oxidized to Fe(III) while structural non-sulfidized Fe(II) (“excess Fe(II)”) forms, which prevents FeS formation at the Fh surface. This excess Fe(II) has been determined chemically but

has so far escaped spectroscopic (Mössbauer) characterization.^{18, 51} In our study, ITFA-LCF did not indicate the presence of a Fe(II) phase other than FeS, but the Fe K-edge XANES data may provide some tentative evidence for formation of non-sulfidized Fe(II) (see Supporting Information). Overall, diffusion-limited $\Sigma\text{H}_2\text{S}$ supply potentially combined with formation of non-sulfidized excess Fe(II) at the expense of FeS formation may explain the relatively low DOS for poorly ordered Fh compared to Lp.

The DOS of P9-FeOx_{hydr} and (to a lesser extent) Si9-FeOx_{hydr} was elevated relative to pure Fh (Fig. 5A). We propose that coprecipitation of P or Si lowers the rate of $\Sigma\text{H}_2\text{S}$ oxidation at the Fh surface (at P/Fe = 0.1, approximately half of reactive surface sites on Fh-type FeOx are occupied^{6, 52}), thereby decreasing diffusion limitations to sulfidation as described above. Surface passivation by P has been suggested to inhibit reaction of dissolved $\Sigma\text{H}_2\text{S}$ with FeOx surfaces.⁴⁴ Silicate sorption had a more limited effect on DOS than P, likely because Si has a lower affinity for FeOx surfaces.^{7, 53-54}

Coprecipitation of P or Si during Fe(II) oxidation resulted in a shift in reaction products from Lp to highly disordered HFO (Fig. 1). Consistent with this shift to highly reactive FeOx that deplete dissolved $\Sigma\text{H}_2\text{S}$, the DOS of P9-FeOx_{oxid} and Si8-FeOx_{oxid} markedly decreased (0.1–0.3; Fig. 5B) relative to Lp. For Si8-FeOx_{oxid}, the value and small range of DOS were similar to Fh, supporting the XAS results that indicated that Si8-FeOx_{oxid} consisted of Si-HFO with a structure similar to Fh (Fig. 1, Fig. S3). Furthermore, the DOS of Si8-FeOx_{oxid} indicated that occupation of FeOx surface sites by Si did not interfere with sulfidation under the experimental conditions ($\Sigma\text{H}_2\text{S}$ 1–2 mmol L⁻¹) and Si/Fe of ~ 0.1. Moreover, the commonalities between Fh and Si8-FeOx_{oxid} suggest that FeOx preparation by Fe(III) hydrolysis produces a poorly ordered FeOx that is representative of natural Fh-like material that commonly forms by Fe(II) oxidation in the presence of dissolved Si.²⁹

For P9-FeOx_{oxid}, a 70:30 mixture of highly disordered P-HFO and Lp (Fig. 1, Fig. S3), the overall DOS showed relatively large variability (Fig. 5B) and differences existed between DOS of the individual Fe components (Fig. 5A). Approximately 5 and 20% of Lp was sulfidized in the mid-depth and bottom sections, respectively, compared to 10 and 40% for P-HFO. This suggests that reaction with highly reactive P-HFO can deplete ΣH₂S and inhibit sulfidation of co-occurring Lp.

Implications. Our experimental approach with gel-based diffusive samplers was able to show that structural alterations in FeOx caused by coprecipitation of oxyanions such as PO₄ and SiO₄ can alter the environmental fate of FeOx. In the case of FeOx sulfidation, poorly ordered, highly reactive FeOx can rapidly deplete dissolved sulfide and thereby limit sulfidation and the associated release of nutrients and contaminants. Diffusion-limited resupply of dissolved sulfide may be more pronounced for gel-embedded FeOx but can also occur in natural sediments when local sulfide gradients develop due to fast reaction of FeOx with sulfide. As such, differences in the rate and extent of sulfidation of FeOx might be particularly relevant in dynamic coastal systems with strong (diurnal) variations in redox conditions. At the same time, environmental conditions, such as high pore-water P concentrations and subsequent FeOx surface passivation, can be the dominating factor controlling FeOx transformations and overwhelm the effect of structural differences. Furthermore, our results show different behavior of amended FeOx from Fe(III) hydrolysis and Fe(II) oxidation, suggesting that laboratory-based findings from Fe(III)-derived Fh-type precipitates may not always be transferable to natural settings, where poorly ordered FeOx commonly form by Fe(II) oxidation. We show that deploying FeOx in natural sediments using gel-based diffusive samplers can elucidate the processes controlling environmental FeOx cycling

341 under natural conditions and help bridge the gap between laboratory experiments and the
342 environmental fate of FeOx.

ASSOCIATED CONTENT

Supporting Information Available.

Detailed descriptions of FeOx synthesis; Preparation of hydrogels and samplers; Deployment and recovery of samplers; Analysis of DGT and DET gels; Sample preparation and analysis for micro-scale XRF and XAS; PCA-ITFA results for entire dataset; ITFA-LCF results for P9-FeOx_{oxid} and Si8-FeOx_{oxid}; Fe K-edge spectra and LCF results for FeOx deployed in freshwater Site 1; Pore-water sulfide and phosphate for marine sediment in the laboratory microcosm experiment; Results of ITFA analysis of the Fe K-edge EXAFS spectra for samples from the laboratory sulfidation experiment; XANES results from the laboratory sulfidation experiment. This material is available free of charge via the Internet at <http://pubs.acs.org>.

AUTHOR INFORMATION

Corresponding Author

* Phone: +31-222-369455

Author Contributions

The manuscript was written through contributions of all authors. All authors have given approval to the final version of the manuscript.

Funding Sources

This work was funded by a grant from the Dutch Research Council, NWO Veni grant 863.14.014 to P. Kraal. The work was further supported by NWO DUBBLE grant 195.068.1039 for ESRF beamline BM26A.

ACKNOWLEDGMENTS

Case M. van Genuchten acknowledges NWO Veni grant 14400. Wytze K. Lenstra acknowledges support from NWO-Vici grant (865.13.005). We gratefully thank the technical support and advice of Dipanjan Banerjee at the DUBBLE beamline during Fe K-edge EXAFS data collection. Anton Tramper, José Mogollón, Nikki Dijkstra, Simon Müller and Silvia Hidalgo-Martinez assisted with field work in the Oosterschelde estuary. Simon Müller and Karel As kindly assisted with data collection at ESRF. We gratefully acknowledge the insightful and constructive feedback from three anonymous reviewers that greatly increased the quality of this article.

ABBREVIATIONS

DET, diffusive equilibrium in thin films; DGT, diffusive gradients in thin films; EXAFS, extended X-ray fine structure; FeOx-BT, FeOx-bearing thin film; Fh, (2-line) ferrihydrite; Lp, lepidocrocite; Mag, magnetite; XANES, X-ray Absorption Near-Edge Spectroscopy; XAS, X-ray absorption spectroscopy

REFERENCES

1. van Genuchten, C. M.; Gadgil, A. J.; Peña, J. Fe(III) Nucleation in the Presence of Bivalent Cations and Oxyanions Leads to Subnanoscale 7 Å Polymers. *Environ. Sci. Technol.* **2014**, *48* (20), 11828-11836.
2. Kaegi, R.; Voegelin, A.; Folini, D.; Hug, S. J. Effect of phosphate, silicate, and Ca on the morphology, structure and elemental composition of Fe(III)-precipitates formed in aerated Fe(II) and As(III) containing water. *Geochim. Cosmochim. Acta* **2010**, *74* (20), 5798-5816.

3. Voegelin, A.; Kaegi, R.; Frommer, J.; Vantelon, D.; Hug, S. J. Effect of phosphate, silicate, and Ca on Fe(III)-precipitates formed in aerated Fe(II)- and As(III)-containing water studied by X-ray absorption spectroscopy. *Geochim. Cosmochim. Acta* **2010**, *74* (1), 164-186.
4. Paige, C. R.; Snodgrass, W. J.; Nicholson, R. V.; Scharer, J. M.; He, Q. H. The effect of phosphate on the transformation of ferrihydrite into crystalline products in alkaline media. *Water, Air, & Soil Pollution* **1997**, *97* (3), 397-412.
5. Cornell, R. M.; Giovanoli, R.; Schindler, P. W. Effect of silicate species on the transformation of ferrihydrite into goethite and hematite in alkaline media. *Clays & Clay Minerals* **1987**, *35* (1), 21-28.
6. Kraal, P.; van Genuchten, C. M.; Behrends, T.; Rose, A. L. Sorption of phosphate and silicate alters dissolution kinetics of poorly crystalline iron (oxyhydr)oxide. *Chemosphere* **2019**, *234*, 690-701.
7. van Genuchten, C. M.; Peña, J.; Amrose, S. E.; Gadgil, A. J. Structure of Fe(III) precipitates generated by the electrolytic dissolution of Fe(0) in the presence of groundwater ions. *Geochim. Cosmochim. Acta* **2014**, *127*, 285-304.
8. Voegelin, A.; Senn, A.-C.; Kaegi, R.; Hug, S. J.; Mangold, S. Dynamic Fe-precipitate formation induced by Fe(II) oxidation in aerated phosphate-containing water. *Geochim. Cosmochim. Acta* **2013**, *117*, 216-231.
9. Waychunas, G. A.; Rea, B. A.; Fuller, C. C.; Davis, J. A. Surface chemistry of ferrihydrite: Part 1. EXAFS studies of the geometry of coprecipitated and adsorbed arsenate. *Geochim. Cosmochim. Acta* **1993**, *57* (10), 2251-2269.
10. Rose, J.; Manceau, A.; Bottero, J.-Y.; Masion, A.; Garcia, F. Nucleation and growth mechanisms of Fe oxyhydroxide in the presence of PO₄ ions. 1. Fe K-Edge EXAFS study. *Langmuir* **1996**, *12* (26), 6701-6707.
11. Rozan, T. F.; Taillefert, M.; Trouwborst, R. E.; Glazer, B. T.; Ma, S.; Herszage, J.; Valdes, L. M.; Price, K. S.; Iii, G. W. L. Iron-sulfur-phosphorus cycling in the sediments of a shallow coastal bay: Implications for sediment nutrient release and benthic macroalgal blooms. *Limnol. Oceanogr.* **2002**, *47* (5), 1346-1354.
12. Krom, M. D.; Berner, R. A. Adsorption of phosphate in anoxic marine sediments. *Limnol. Oceanogr.* **1980**, *25* (5), 797-806.
13. Cornell, R. M.; Schwertmann, U., *The Iron Oxides: Structure, Properties, Reactions, Occurrences and Uses*. Wiley: 2003; p 664.
14. Hansel, C. M.; Benner, S. G.; Fendorf, S. Competing Fe(II)-induced mineralization pathways of ferrihydrite. *Environ. Sci. Technol.* **2005**, *39* (18), 7147-7153.
15. Hansel, C. M.; Benner, S. G.; Neiss, J.; Dohnalkova, A.; Kukkadapu, R. K.; Fendorf, S. Secondary mineralization pathways induced by dissimilatory iron reduction of ferrihydrite under advective flow. *Geochim. Cosmochim. Acta* **2003**, *67* (16), 2977-2992.
16. Burton, E. D.; Johnston, S. G.; Bush, R. T. Microbial sulfidogenesis in ferrihydrite-rich environments: Effects on iron mineralogy and arsenic mobility. *Geochim. Cosmochim. Acta* **2011**, *75* (11), 3072-3087.
17. Yang, L.; Steefel, C. I.; Marcus, M. A.; Bargar, J. R. Kinetics of Fe(II)-Catalyzed Transformation of 6-line Ferrihydrite under Anaerobic Flow Conditions. *Environ. Sci. Technol.* **2010**, *44* (14), 5469-5475.

18. Peiffer, S.; Behrends, T.; Hellige, K.; Larese-Casanova, P.; Wan, M.; Pollok, K. Pyrite formation and mineral transformation pathways upon sulfidation of ferric hydroxides depend on mineral type and sulfide concentration. *Chem. Geol.* **2015**, *400*, 44-55.
19. Davison, W.; Zhang, H. In situ speciation measurements of trace components in natural waters using thin-film gels. *Nature* **1994**, *367* (6463), 546-548.
20. Davison, W.; Grime, G. W.; Morgan, J. A. W.; Clarke, K. Distribution of dissolved iron in sediment pore waters at submillimetre resolution. *Nature* **1991**, *352* (6333), 323-325.
21. Michel, F. M.; Ehm, L.; Antao, S. M.; Lee, P. L.; Chupas, P. J.; Liu, G.; Strongin, D. R.; Schoonen, M. A. A.; Phillips, B. L.; Parise, J. B. The structure of ferrihydrite, a nanocrystalline material. *Science* **2007**, *316* (5832), 1726-1729.
22. Hiemstra, T. Surface and mineral structure of ferrihydrite. *Geochim. Cosmochim. Acta* **2013**, *105* (0), 316-325.
23. Schwertmann, U.; Cornell, S., *Iron oxides in the laboratory: Preparation and characterization*. Wiley: 1991; p 188.
24. Bigham, J. M.; Schwertmann, U.; Traina, S. J.; Winland, R. L.; Wolf, M. Schwertmannite and the chemical modeling of iron in acid sulfate waters. *Geochim. Cosmochim. Acta* **1996**, *60* (12), 2111-2121.
25. Ferris, F. G.; Tazaki, K.; Fyfe, W. S. Iron oxides in acid mine drainage environments and their association with bacteria. *Chem. Geol.* **1989**, *74* (3), 321-330.
26. Yu, J.-Y.; Heo, B.; Choi, I.-K.; Cho, J.-P.; Chang, H.-W. Apparent solubilities of schwertmannite and ferrihydrite in natural stream waters polluted by mine drainage. *Geochim. Cosmochim. Acta* **1999**, *63* (19), 3407-3416.
27. Carlson, L.; Bigham, J. M.; Schwertmann, U.; Kyek, A.; Wagner, F. Scavenging of As from Acid Mine Drainage by Schwertmannite and Ferrihydrite: A Comparison with Synthetic Analogues. *Environ. Sci. Technol.* **2002**, *36* (8), 1712-1719.
28. Jambor, J. L.; Dutrizac, J. E. Occurrence and constitution of natural and synthetic ferrihydrite, a widespread iron oxyhydroxide. *Chemical Reviews* **1998**, *98* (7), 2549-2586.
29. Schwertmann, U., Occurrence and Formation of Iron Oxides in Various Pedoenvironments. In *Iron in Soils and Clay Minerals*, Stucki, J. W.; Goodman, B. A.; Schwertmann, U., Eds. Springer Netherlands: Dordrecht, 1988; pp 267-308.
30. Jensen, H. S.; Kristensen, P.; Jeppesen, E.; Skytthe, A. Iron-phosphorus ratio in surface sediment as an indicator of phosphate release from aerobic sediments in shallow lakes. *Hydrobiologia* **1992**, *235*, 731-743.
31. Kraal, P.; Dijkstra, N.; Behrends, T.; Slomp, C. P. Phosphorus burial in sediments of the sulfidic deep Black Sea: key roles for adsorption by calcium carbonate and apatite authigenesis. *Geochim. Cosmochim. Acta* **2017**, *204*, 140-158.
32. Hyacinthe, C.; Van Cappellen, P. An authigenic iron phosphate phase in estuarine sediments: composition, formation and chemical reactivity. *Mar. Chem.* **2004**, *91* (1-4), 227-251.
33. Bennett, W. W.; Teasdale, P. R.; Welsh, D. T.; Panther, J. G.; Jolley, D. F. Optimization of colorimetric DET technique for the in situ, two-dimensional measurement of iron(II) distributions in sediment porewaters. *Talanta* **2012**, *88* (0), 490-495.
34. Pagès, A.; Teasdale, P. R.; Robertson, D.; Bennett, W. W.; Schäfer, J.; Welsh, D. T. Representative measurement of two-dimensional reactive phosphate distributions and co-distributed iron(II) and sulfide in seagrass sediment porewaters. *Chemosphere* **2011**, *85* (8), 1256-1261.

35. Robertson, D.; Teasdale, P. R.; Welsh, D. T. A novel gel-based technique for the high resolution, two-dimensional determination of iron (II) and sulfide in sediment. *Limnol. Oceanogr. Meth.* **2008**, *6*, 502-512.
36. Nikitenko, S.; Beale, A. M.; van der Eerden, A. M. J.; Jacques, S. D. M.; Leynaud, O.; O'Brien, M. G.; Detollenaere, D.; Kaptein, R.; Weckhuysen, B. M.; Bras, W. Implementation of a combined SAXS/WAXS/QEXAFS set-up for time-resolved in situ experiments. *J. Synchrotron Radiat.* **2008**, *15* (6), 632-640.
37. Ravel, B.; Newville, M. Athena, Artemis, Hephaestus: data analysis for X-ray absorption spectroscopy using IFEFFIT. *J. Synchrotron Radiat.* **2005**, *12*, 537-541.
38. Rossberg, A.; Reich, T.; Bernhard, G. Complexation of uranium(VI) with protocatechuic acid—application of iterative transformation factor analysis to EXAFS spectroscopy. *Anal. Bioanal. Chem.* **2003**, *376* (5), 631-638.
39. Boesen, C.; Postma, D. Pyrite formation in anoxic environments of the Baltic. *Am. J. Sci.* **1988**, *288* (6), 575-603.
40. Jones, A. M.; Collins, R. N.; Waite, T. D. Redox characterization of the Fe(II)-catalyzed transformation of ferrihydrite to goethite. *Geochim. Cosmochim. Acta* **2017**, *218*, 257-272.
41. Pedersen, H. D.; Postma, D.; Jakobsen, R.; Larsen, O. Fast transformation of iron oxyhydroxides by the catalytic action of aqueous Fe(II). *Geochim. Cosmochim. Acta* **2005**, *69* (16), 3967-3977.
42. Yee, N.; Shaw, S.; Benning, L. G.; Nguyen, T. H. The rate of ferrihydrite transformation to goethite via the Fe(II) pathway. *Am. Mineral.* **2006**, *91* (1), 92-96.
43. Schoepfer, V. A.; Burton, E. D.; Johnston, S. G.; Kraal, P. Phosphate loading alters schwertmannite transformation rates and pathways during microbial reduction. *Sci. Total Environ.* **2019**, *657*, 770-780.
44. Biber, M. V.; dos Santos Afonso, M.; Stumm, W. The coordination chemistry of weathering: IV. Inhibition of the dissolution of oxide minerals. *Geochim. Cosmochim. Acta* **1994**, *58* (9), 1999-2010.
45. Kraal, P.; Burton, E. D.; Bush, R. T. Iron monosulfide accumulation and pyrite formation in eutrophic estuarine sediments. *Geochim. Cosmochim. Acta* **2013**, *122* (0), 75-88.
46. Rickard, D. Kinetics of pyrite formation by the H₂S oxidation of iron (II) monosulfide in aqueous solutions between 25 and 125°C: The rate equation. *Geochim. Cosmochim. Acta* **1997**, *61* (1), 115-134.
47. Slomp, C. P.; Van Raaphorst, W. Phosphate adsorption in oxidized marine sediments. *Chem. Geol.* **1993**, *107* (3-4), 477-480.
48. Scholz, F.; McManus, J.; Mix, A. C.; Hensen, C.; Schneider, R. R. The impact of ocean deoxygenation on iron release from continental margin sediments. *Nature Geosci.* **2014**, *7* (6), 433-437.
49. Poulton, S. W.; Krom, M. D.; Raiswell, R. A revised scheme for the reactivity of iron (oxyhydr)oxide minerals towards dissolved sulfide. *Geochim. Cosmochim. Acta* **2004**, *68* (18), 3703-3715.
50. Canfield, D. E.; Raiswell, R.; Bottrell, S. H. The reactivity of sedimentary iron minerals toward sulfide. *Am. J. Sci.* **1992**, *292* (9), 659-683.
51. Hellige, K.; Pollok, K.; Larese-Casanova, P.; Behrends, T.; Peiffer, S. Pathways of ferrous iron mineral formation upon sulfidation of lepidocrocite surfaces. *Geochim. Cosmochim. Acta* **2012**, *81*, 69-81.

52. Dzombak, D. A.; Morel, F. M. M., *Surface complexation modeling: Hydrous ferric oxide*. Wiley: New York, 1990; p 416.
53. Xu, N.; Christodoulatos, C.; Braida, W. Modeling the competitive effect of phosphate, sulfate, silicate, and tungstate anions on the adsorption of molybdate onto goethite. *Chemosphere* **2006**, *64* (8), 1325-1333.
54. Hiemstra, T. Ferrihydrite interaction with silicate and competing oxyanions: Geometry and Hydrogen bonding of surface species. *Geochim. Cosmochim. Acta* **2018**, *238*, 453-476.

Supporting Information

Coprecipitation of phosphate and silicate affects environmental iron
(oxyhydr)oxide transformations: a gel-based diffusive sampler approach

Peter Kraal, Case M. van Genuchten, Wytze K. Lenstra, Thilo Behrends

15 pages

11 figures

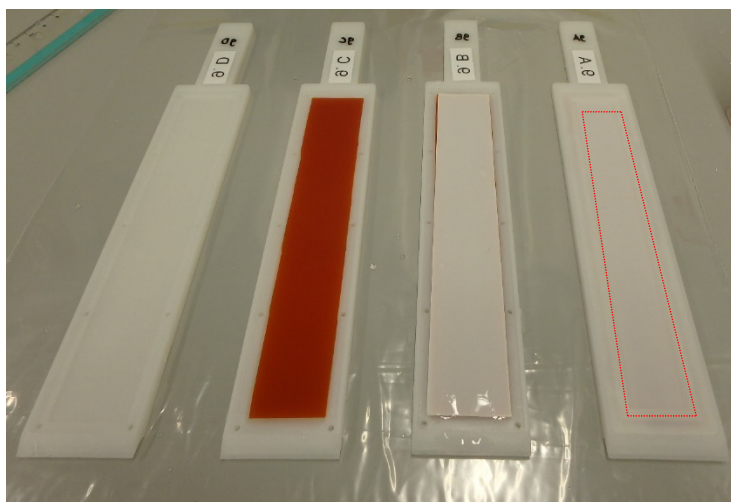
2 tables

Synthesis of FeOx. Pure two-line ferrihydrite (Fh; $\text{Fe}_{10}\text{O}_{14}(\text{OH})_2 \cdot n\text{H}_2\text{O}$ with $n < 1$)¹⁻² was synthesized by rapid base hydrolysis of a concentrated $\text{Fe(III)(NO}_3)_3$ solution with CO_2 -free (boiled and N_2 -purged) 1 M NaOH, following the protocol of Schwertmann and Cornell.³ Lepidocrocite (Lp; $\gamma\text{-FeOOH}$) was synthesized by oxidation of a concentrated Fe(II)Cl_2 solution by bubbling with regular air using a diffuser connected to a compressed air outlet, setting the flow rate to $\sim 100 \text{ mL min}^{-1}$.³ Fh and Lp synthesis was also performed in the presence of dissolved phosphate (PO_4 as $\text{Na}_2\text{HPO}_4 \cdot 7\text{H}_2\text{O}$) or silicate (SiO_4 as $\text{Na}_2\text{SiO}_3 \cdot 9\text{H}_2\text{O}$) with a molar (P, Si)/Fe ratios of 1/10. These amended protocols, detailed elsewhere,⁴ were applied to synthesize ferric iron precipitates (iron (oxyhydr)oxides, referred to here as FeOx) with co-precipitated P or Si. For Fe(III) hydrolysis with coprecipitated P or Si, the NaOH solution was amended with $\sim 30 \text{ mmol L}^{-1}$ P or Si and added at a rate of $\sim 20 \text{ mL min}^{-1}$ with a total volume that resulted in a final molar (P, Si)/Fe ratio of 0.1. For FeOx by oxidation of Fe(II)Cl_2 , a 1 M NaOH solution amended with $\sim 50 \text{ mmol L}^{-1}$ P or Si was added continuously during oxidation of the Fe(II) solution using a Metrohm Titrando 809 device to buffer pH at 6.8, also resulting in a final (P, Si)/Fe ratio of 0.1. Because of the high pH of the Si-amended NaOH stock solution and relatively low Si/Fe ratio during synthesis, it is unlikely that significant Si polymerization occurred in solution or at the FeOx surface.⁵⁻⁶ Each FeOx precipitate was poured into Spectra/Por 4 regenerated cellulose dialysis tubing (SpectrumLabs) and dialyzed in 10 L of ultrapure water with regular water changes until the electric conductivity was $< 10 \mu\text{S cm}^{-1}$ (4 days). The FeOx were stored in suspension in the dark at 4°C for up to two weeks prior to further processing. The final concentration in all suspensions was $25 - 30 \text{ mg FeOx mL}^{-1}$. The composition of the FeOx was determined by dissolution of freeze-dried sub-samples in 6 mol L^{-1} HCl and spectrophotometric measurement of Fe, P and Si.

Preparation of (FeOx-BT) hydrogels and samplers. The preparation of diffusive and binding hydrogels is described in detail in previous studies.⁷⁻⁹ For this study, FeOx-bearing thin film (FeOx-BT) hydrogels were prepared by mixing adding 3 mL FeOx suspension per 10 mL of acrylamide solution, stirring vigorously and adding ammonium persulfate and N,N,N',N'-tetramethylethylenediamine. This suspension was used to cast 0.8 mm-thick Fe-MDT hydrogels. The total Fe content in the fresh hydrated gels was $\sim 60 \mu\text{mol cm}^{-3}$, which is in the range of FeOx-Fe contents in oxic surface sediments. Adding FeOx suspension rather than dried FeOx optimized the homogeneity of FeOx distribution after casting the gel. In previous work¹⁰⁻¹¹, Fh was precipitated directly within the pores of hydrogels. Here, for consistency we preferred to produce stock FeOx suspensions that were used for a range of laboratory and field experiments (also reported elsewhere⁴), including deployment of FeOx-BT hydrogels for the current study. All hydrogels were cleaned with repeated milli-Q baths (at least 5 water changes) and cut into $2.5 \times 16 \text{ cm}$ hydrogel strips.

For pore-water sampling, DGT-DET samplers were prepared by placing a DGT hydrogel strip in a rigid TeflonTM sampler (DGT Research), which was then overlain by a DET hydrogel strip and a strip of $0.45 \mu\text{m}$ nitrocellulose filter membrane. The gels and filter were kept in place by front window (Fig. S1). For FeOx deployment, instead of a DGT-DET gel stack, a FeOx-BT hydrogel strip was covered by filter membrane.

45



46

47 **Fig. S1.** Preparation of FeOx-BT hydrogel sampler in four steps. From left to right: empty
 48 Teflon sampler (DGT Research); loading of FeOx-BT hydrogel; filter membrane cover;
 49 securing FeOx-BT gel and filter in place with front window ($\sim 1.5 \times 12$ cm open area for
 50 diffusive exchange, indicated by red dashed line) for solute exchange between gel and pore-
 51 water.

52

53 **Deployment and recovery of hydrogel samplers.** DGT-DET and FeOx-BT samplers
 54 were pushed vertically into the sediment, for pore-water samplers ~ 1 cm of the sampling
 55 window remaining above the sediment-water interface. Pore-water DGT-DET samplers for
 56 Fe(II), P and $\Sigma\text{H}_2\text{S}$ were deployed overnight at the beginning of the FeOx deployment period.
 57 The selected FeOx-BT samplers were deployed in the field for four weeks, in duplicate (i.e.
 58 two samplers for each FeOx). At Site 1, samplers were placed in the ditch sediment under \sim
 59 50 cm of water. At Site 2, samplers were placed in tidal flat sediment under shallow water (\sim
 60 20 cm) during low tide, assuming the area never fell completely dry (which would cause the
 61 gels to dry out). Upon recovery, the DGT-DET samplers were processed and analyzed on-site.
 62 The FeOx-BT samplers were placed in ~ 5 L of Ar-purged anoxic water (milliQ for
 63 freshwater Site 1, 0.7 M sodium chloride for marine Site 2) and immediately transported to an
 64 Ar-filled anoxic glovebox in the Utrecht University laboratory, within 4 h after retrieval.
 65 Pore-water sampling with DGT-DET proved challenging at Site 2 because of the low pore-
 66 water Fe(II) and P concentrations. Therefore, we also took a 20 cm-long sediment core with
 67 overlying water that was sectioned at 1-cm vertical resolution in a N_2 -purged glovebag within
 68 2 h after recovery. The 1-cm sediment slices were transferred into polyethylene centrifuge
 69 tubes, centrifuged (4000 rpm, 15 min) and the $0.45 \mu\text{m}$ -filtered supernatant was used for
 70 analysis of dissolved Fe(II), $\Sigma\text{H}_2\text{S}$, P and Si using standard spectrophotometric methods:
 71 molybdate blue (P, Si), methylene blue ($\Sigma\text{H}_2\text{S}$) and 1,10-phenanthroline.¹² The results of this
 72 classical pore-water sampling are shown in Fig. S2.

73

74

75

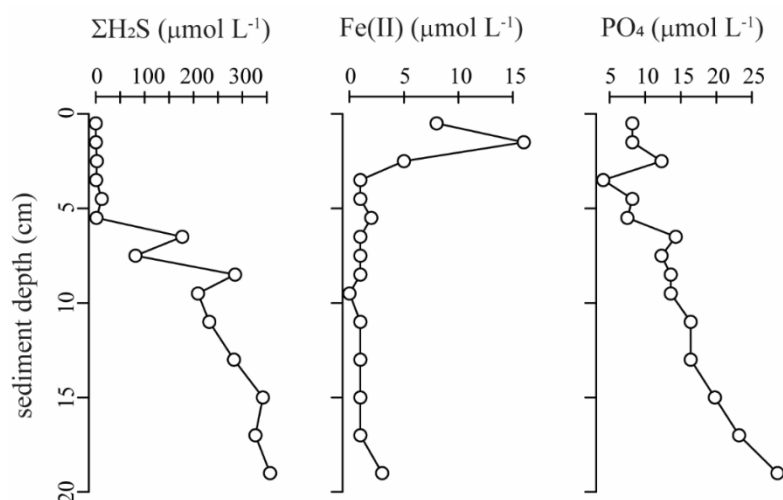


Fig. S2. Concentration profiles for dissolved $\Sigma\text{H}_2\text{S}$, Fe(II) and P determined by spectrophotometric analysis of filtered pore-water from Site 2.

Upon recovery, inspection of the FeOx-BT hydrogels after deployment at Site 2 showed that there were large differences in the depth of the sulfidation front (S3):

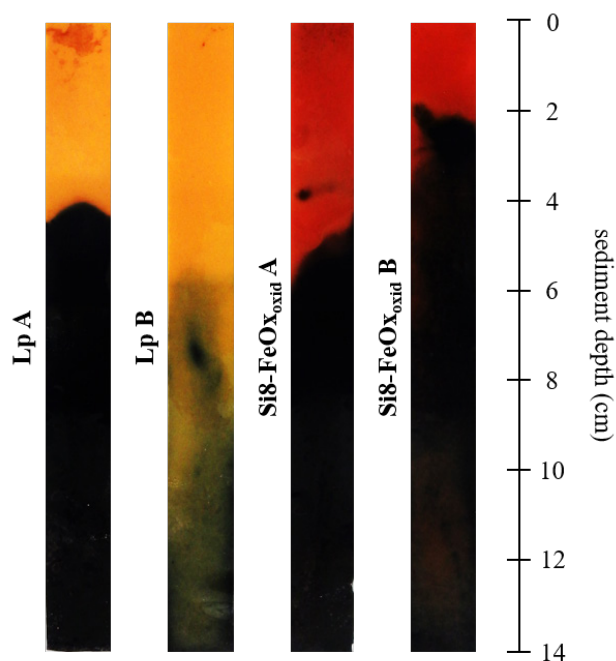


Fig. S3. Images of duplicate Lp and $\text{Si8-FeOx}_{\text{oxi}}\text{d}$ gels, showing the large variability in the depth and intensity of the sulfidation front. Image color corrections: saturation (+ 40 %) and contrast (+ 40 %) were increased to enhance visibility of sulfidized gel sections.

Processing and analysis of DGT and DET gels. Duplicate DGT-DET samplers were used for simultaneous Fe(II), P and $\Sigma\text{H}_2\text{S}$ measurement. The DET hydrogel from the first probe (sample gel) was removed and immediately covered by a DET hydrogel that had equilibrated (2 h) with a FerrozineTM solution (staining gel). Upon diffusive exchange, the FerrozineTM in the staining gel forms a magenta complex with Fe(II) in the sample gel.¹³ A similar process was applied to the DET sample gel from the second probe to measure P using a DET staining gel equilibrated in a solution with molybdate and ascorbic acid which forms a reduced blue complex with P.⁹ After 15 min, 300 dpi color images of the gel stacks were collected (Canon CanoScan LiDE 210). The color images were converted to monochrome grayscale and the grayscale intensity (GSI) maps were converted into two-dimensional Fe(II) and P distributions, using Fe(II) and P versus GSI calibration curves with a range of 0–1500 and 0–750 $\mu\text{mol L}^{-1}$ (see below). In the time frame between recovering DGT-DET gels and scanning, diffusional relaxation is limited and fidelity to the sampled distribution of the pore-water Fe(II) and P is retained as much as possible.⁷

The two dissolved sulfide DGT gels from the duplicate DGT-DET samplers were placed overnight in 0.01 mol L⁻¹ hydroxylamine to remove staining by FeOx. The gels were then scanned as described for the DET gels. In case of DGT gels, sulfide is bound by reaction with AgI that is mixed into the DGT hydrogel. This reaction results in a color change from yellowish white (AgI hydrogel) to black (Ag₂S stains). Sulfide concentrations were calculated from grayscale intensity using a previously published calibration.⁷

For calibration of the DET method for Fe(II) and P, duplicate series of calibration solutions were prepared. For each analyte, one set of duplicate calibration solutions was prepared in 0.01 M NaCl as freshwater representative and another set of duplicates was prepared in 0.7 M NaCl as seawater representative. For Fe(II), following Bennet et al.¹³, a set of ten calibration solutions was prepared using ammonium Fe(II) sulfate covering a 0 - 1500 $\mu\text{mol L}^{-1}$ Fe(II) range in 0.01 mol L⁻¹ hydroxylamine hydrochloride. For P, following Pagès et al.¹⁴, a set of ten calibration solutions was prepared using sodium dihydrogen phosphate covering a 0 - 750 $\mu\text{mol L}^{-1}$ P range in 5% (v/v) hydrochloric acid. Two circular pieces (1.5 cm diameter) of DET hydrogel were placed in each solution and allowed to equilibrate for at least two hours. At the same time, corresponding gel pieces were placed for at least two hours in the Fe(II) staining reagent (0.01 mol L⁻¹ Ferrozine in 0.1 mol L⁻¹ acetic acid/sodium acetate) or the P staining reagent (0.045 mol L⁻¹ ammonium heptamolybdate, 0.5 mmol L⁻¹ potassium antimonyl tartrate, 2.5 mol L⁻¹ sulfuric acid, 0.03 mol L⁻¹ ascorbic acid). Subsequently, a gel piece from a Fe(II) or P calibration solution was overlain with a gel piece from the appropriate staining reagent. The gel stacks were scanned after 15 minutes and the color intensity was converted to grey scale intensity. The results showed that the upper detection limit was $\sim 1000 \mu\text{mol L}^{-1}$ for Fe(II) and $> 750 \mu\text{mol L}^{-1}$ for P (Fig. S4). The practical lower detection limit for both analytes was $\sim 10 \mu\text{mol L}^{-1}$.¹³ As established previously¹³, the effect of ionic strength was minor ($< 10\%$ difference between 0.01 M NaCl and 0.1 M NaCl) and we only report the results from the 0.7 M NaCl calibration solutions.

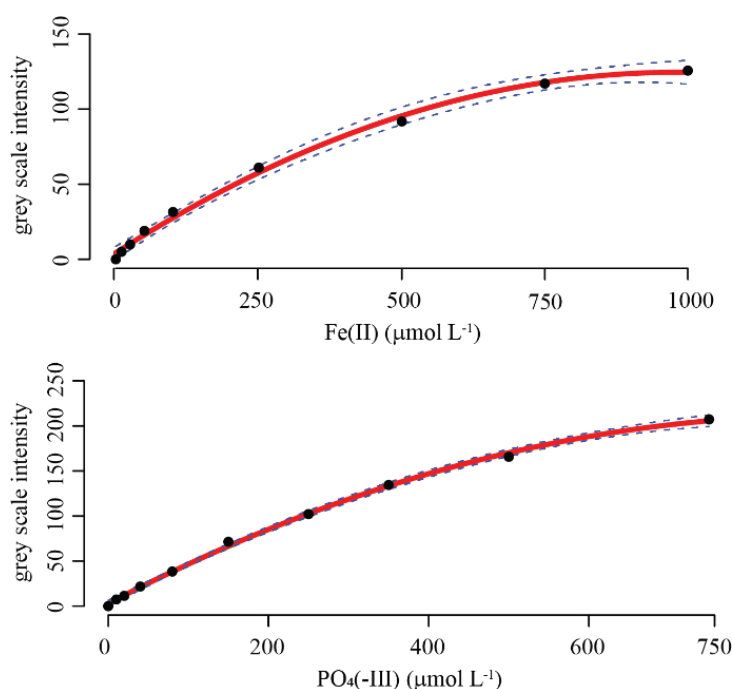


Fig. S4. Calibration curves for dissolved Fe(II) (top) and PO₄³⁻ (bottom) in 0.7 M NaCl, showing grey scale intensity plotted against analyte concentration. Black dots show data points, red line is non-linear regression and blue dashed lines represent the 95% confidence interval based on duplicate series.

Sample preparation and analysis for micro-scale XRF and synchrotron based XAS. For μ XRF analysis, selected gel cuts were placed on an acid-washed glass microscope slide and tightly covered with cling film. The slide with gel cut was then immediately placed in the sample chamber of an Orbis Micro X-ray Fluorescence Analyzer (EDAX) and brought under vacuum. The Orbis EDAX was operated at 30 keV with 30 \times 30 μ m spatial resolution and a dwell time of 800 ms. After reaching sufficient vacuum, vertical (one-dimensional) line scans from the top to the bottom of hydrogel sections (\sim 3.5 cm) were collected. Because the hydrogel sections dried out under vacuum and started deforming after about 30 min, these analyses were carried out within 20 min of placing gel sections in the vacuum chamber of the Orbis EDAX.

The hydrogel sections for synchrotron-based X-ray absorption spectroscopy (XAS) were freeze-dried under an N₂ atmosphere and powdered in the anoxic glovebox with an agate mortar and pestle. An amount of Fe-doped hydrogel containing the Fe equivalent of \sim 1 absorption length¹⁵ was mixed with microcrystalline cellulose (Sigma-Aldrich) and pressed into a 7 mm-diameter pellet (thickness \sim 1 mm) with a hand press (Pike Technologies). The pellets were transported in Ar-purged, gas-tight aluminum laminate bags to the European Synchrotron Radiation Facility (ESRF) in Grenoble, France. Iron K-edge XAS spectra were collected at beamline BM26A (DUBBLE). Details on the beamline setup can be found in Nikitenko et al.¹⁶ Samples were loaded into the sample holder in an anoxic N₂-purged glovebag at the beamline and data was collected as described in detail previously.⁴ In short, samples were sealed between two layers of Kapton tape and spectra were recorded at room

temperature in transmission mode from 6962 to 7862 eV, corresponding to a maximum k of 13 \AA^{-1} in the EXAFS region. For each sample, two spectra were collected and merged. Repeated measurement of FeS-rich samples showed no sign of oxidation or beam damage during analysis. Data normalization and manipulation were performed using the Athena software package.¹⁷ Fourier transformation was performed using a Rkbg value of 1.0, a k -range of $2\text{--}11 \text{ \AA}^{-1}$ and a Hanning window function. Using the ITFA software package¹⁸, principal components analysis (PCA) was performed to assess the required number independent components to reproduce the major variance of the entire EXAFS data set that consisted of 24 spectra: FeOx starting materials, samples from the freshwater Site 1 hydrogels and hydrogel samples from the laboratory-based sulfidation experiment. Also included was an existing reference spectrum of synthetic mackinawite (FeS).¹⁹ The components were extracted from the dataset by iterative transformation factor analysis (ITFA), and linear combination fitting (LCF) was used to determine the relative contribution of the independent components to each sample spectrum (i.e. Fe speciation). PCA-ITFA-LCF analysis was performed in k space, using a fitting window of $k = 2 - 11 \text{ \AA}^{-1}$ of the k^3 weighted EXAFS spectra. The ITFA analysis requires that the relative concentration of each independent component is constrained in at least one sample. The starting materials Fh and Lp and the reference spectrum of FeS were used as endmembers: the relative concentration of the component with the strongest loading was set to 1 and that of all other components to 0.

PCA-ITFA analysis of the Fe K-edge EXAFS dataset. Principal component analysis (PCA) and iterative transformation factor analysis (ITFA) on all normalized EXAFS spectra (FeOx starting materials, freshwater samples, samples from laboratory-based sulfidation experiment) were performed with the ITFA software.¹⁸ PCA showed that the major variance in the EXAFS dataset could be reproduced using three components (Fig. S5).

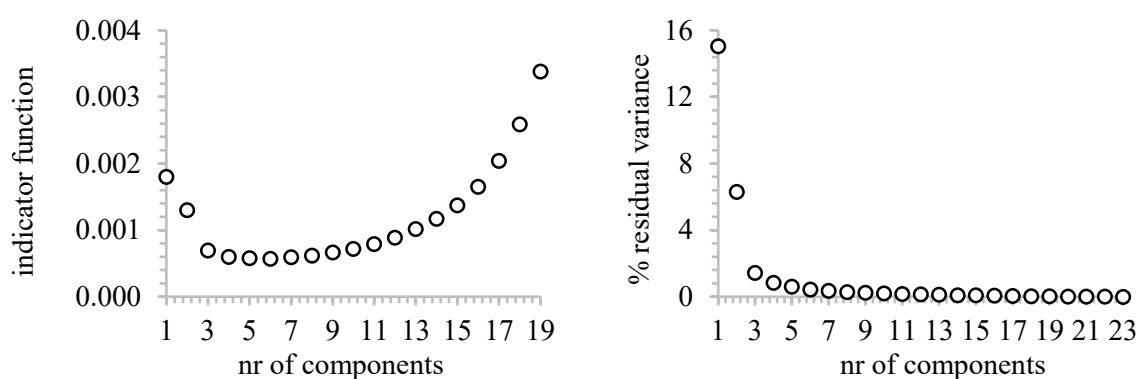


Fig. S5. ITFA indicator function plotted against number of independent components (left) and % residual variance as function of number of components (right). Minimum error and low residual variance indicate that optimal results are obtained with using three components to reproduce the sample spectra. Note: the indicator function rapidly increased for number of components > 19 and therefore errors using 20–23 components are omitted in the left panel as it would mask the decrease in error for low numbers of components.

The components showed spectral properties similar to the spectra of the Fe reference standards Fh, Lp and FeS (Fig. S6).

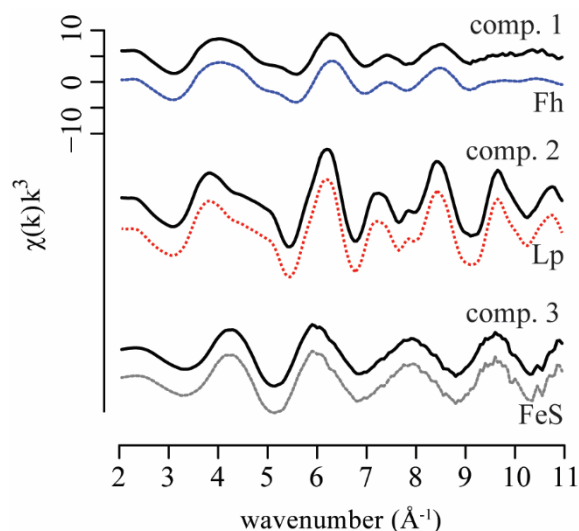


Fig. S6. Spectra of the three ITFA components and the three reference materials (Fh, Lp and FeS) that these components represent.

Fe speciation from PCA-ITFA in starting materials from Fe(II) oxidation. Linear combination fitting (LCF) using the spectra of the PCA-ITFA-derived independent components indicated that P9-FeO_x was a mixture of highly disordered P-HFO (70 %) and Lp (30 %). The spectrum of Si8-FeO_x indicated poorly ordered Si-HFO as the only phase (Fig. S7). It is important to note that LCF was performed with the independent components representing Fh and Lp (Fig. S6). The spectrum of Si8-FeO_x is similar to that of Fh, but the spectrum of P9-FeO_x has a very poorly developed corner-sharing Fe peak (Fig. 1). This indicates that the structure of the P-rich Fe precipitate is different from Fh, which does show a corner-sharing Fe peak. Indeed, previous work indicates that Fe precipitate formed by Fe(II) oxidation in the presence of dissolved P consists of highly disordered Fe(III)-P polymers and/or P-rich hydrous ferric oxide.²⁰⁻²¹ This material is less ordered than Fh, and therefore an exaggerated relative concentration of Fh is needed in the LCF to reproduce the relatively high degree of disorder in the P9-FeO_x spectrum. As poorly ordered Fe is overestimated, the relative concentration of Lp in the sample is likely underestimated. Fitting the P9-FeO_x spectrum with reference spectra for Lp and for highly disordered Fe(III)-P polymers (obtained during previous work²¹) did not improve the quality-of-fit. This suggests that the P-rich Fe precipitate on P9-FeO_x consisted of P-HFO rather than Fe(III)-polymers, which is to be expected under the experimental conditions where the excess of Fe(III) over P (P/Fe = 0.1) was large enough to drive polymerization of Fe(III)-P to P-HFO during synthesis.²⁰

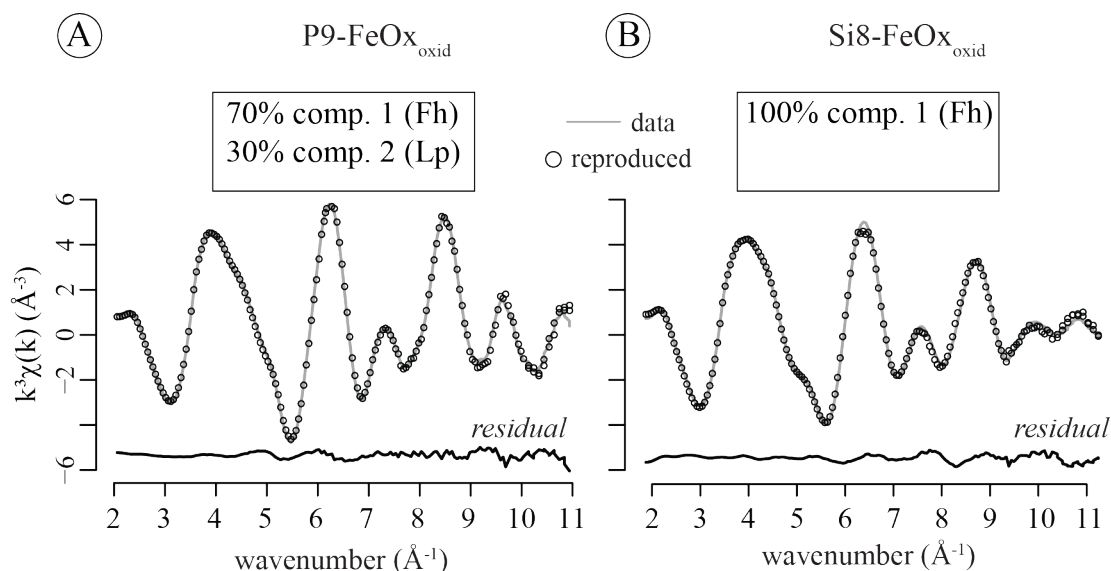


Figure S7. Data (grey lines), reproduced spectra (black dots) and residuals (data - fit) for the k^3 -weighted Fe K-edge EXAFS data (fitting in k space, 2 - 11 \AA^{-1}) of the FeOx starting materials synthesized by Fe(II) oxidation with coprecipitation of P ($\text{P9-FeOx}_{\text{oxid}}$, panel A) or Si ($\text{Si8-FeOx}_{\text{oxid}}$, panel B).

LCF results for the Fe K-edge EXAFS spectra from Site 1 (freshwater) samples.

Results of ITFA-LCF indicated that all spectra of samples from freshwater Site 1 were reproduced as 100% poorly ordered FeOx . To look in more detail into Fe speciation in these samples, we performed additional LCF in Athena, where sample spectra were reproduced as linear combinations of spectra from reference materials rather than independent components (Fig. S8). The results indicated that slight alteration of the starting material occurred in the cases of $\text{P9-FeOx}_{\text{hydr}}$ and $\text{Si9-FeOx}_{\text{hydr}}$, even though all Fe remained present as poorly ordered FeOx (Table S1).

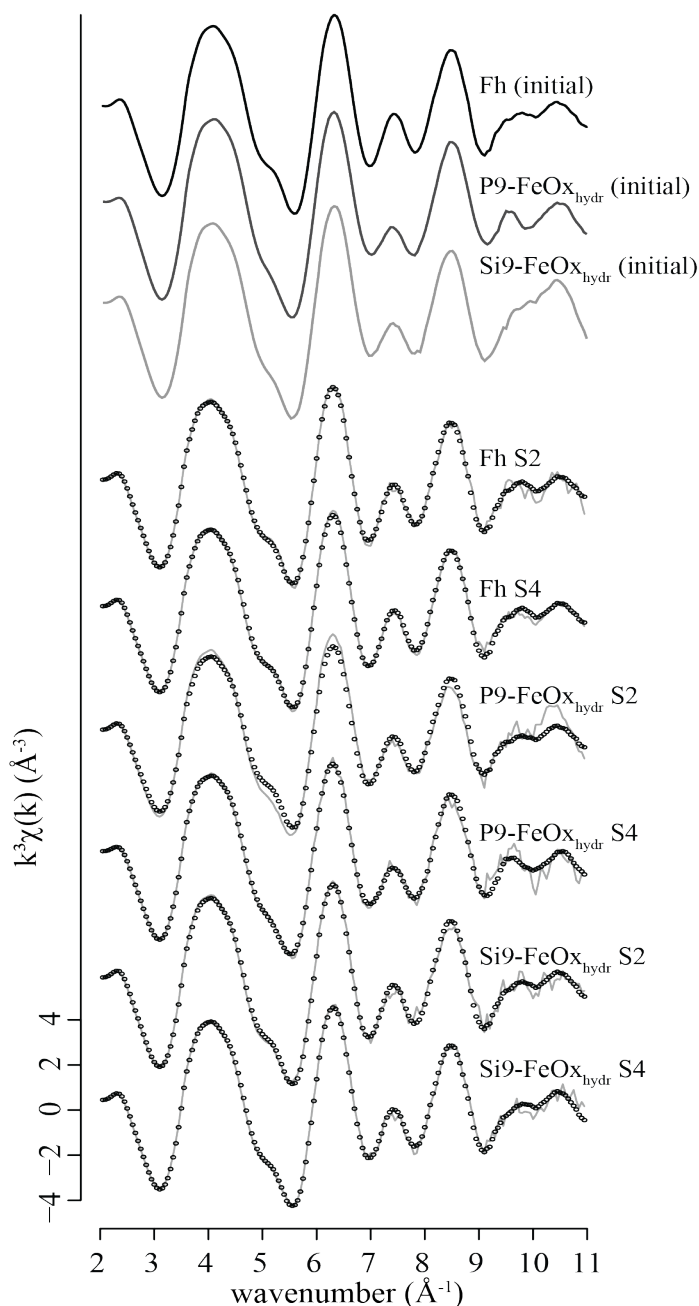


Figure S8. Data (grey line) and LCF fits (dots) (fitting k^3 -weighted spectra in k space, 2 - 11 \AA^{-1}) for the FeOx after four weeks of deployment in the freshwater sediment of Site 1. S2 is section 2 of gel (3.5 - 7 cm) and S4 is section 4 of gel (10.5 - 14 cm). For each spectrum, the two spectra of samples from duplicate probes were merged. For comparison, spectra of starting materials are included at the top.

Table S1. Results of linear combination fitting of the Fe K-edge EXAFS data (fitting in χ space, k 2 - 11 \AA^{-1}) from the freshwater drainage ditch samples. Fractions were recalculated to sum 1. S2 and S4 refer to the mid-depth (S2) and deepest (S4) sections of the deployed Fe-MDT hydrogels. The standard deviation, representing the variability between samples from duplicate probes, is between parentheses. Preliminary LCF with a suite of (mixed) ferric and ferrous Fe phases such as lepidocrocite, goethite, hematite, magnetite, vivianite and siderite

241 showed that the contribution of these reference spectra to the sample spectra was negligible
 242 (accounting for < 8% of total Fe in any combination of references).

	Fh S2	Fh S4	P9-FeOx _{hydr} S2	P9-FeOx _{hydr} S4	Si9-FeOx _{hydr} S2*	Si9-FeOx _{hydr} S4
P9-FeOx _{hydr}	-	-	0.00 (0.00)	0.73 (0.40)	-	-
Si9-FeOx _{hydr}	-	-	-	-	0.82	0.29 (0.04)
Fh	1.00 (0.01)	1.00 (0.01)	1.00 (0.00)	0.27 (0.37)	0.18	0.71 (0.05)
R factor**	0.028	0.015	0.039	0.039	0.019	0.020

243

244 * No duplicate available; no standard deviation

245 ** R factor is an estimate of the fitting error calculated as:

246 $\frac{\sum((data - fit)^2)}{\sum((data)^2)}$ over the fitting range $k = 2 - 11 \text{ \AA}^{-1}$. Highest R factor of
 247 duplicates is reported.

248

249

250

Dissolved sulfide and phosphate concentrations in marine sediment in the micro-cosm laboratory experiment. Pore-water profiles for DGT-DET showed accumulation of dissolved H_2S and P I the surface sediment. The DET gel for Fe(II) showed dark smears that precluded appropriate processing of the gel to obtain a Fe(II) map. Visual observation of areas of the gel that were not compromised, compared to the magenta color of the Fe(II) calibration series, indicated a transient Fe(II) peak of small magnitude ($< 50 \mu\text{mol L}^{-1}$) between $\sim 2\text{--}4 \text{ cm}$ depth where H_2S was absent.

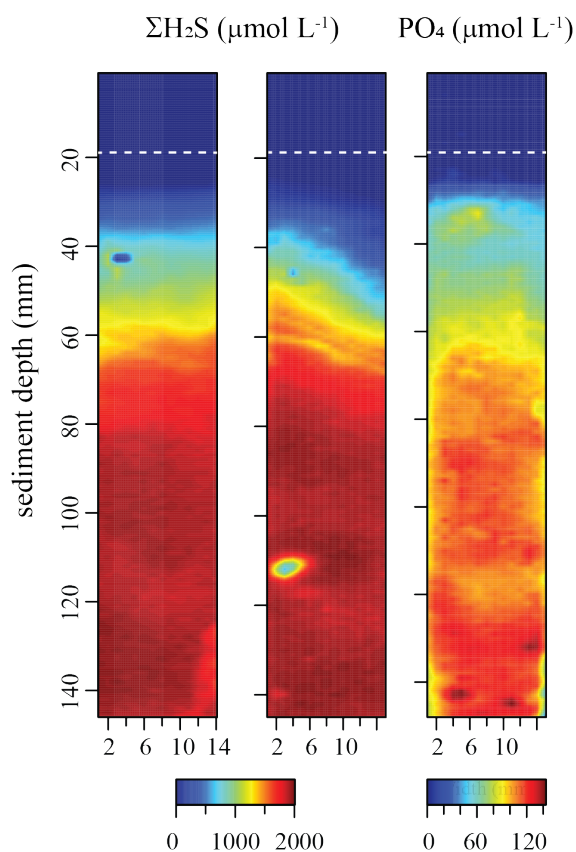


Fig. S9. Dissolved $\Sigma\text{H}_2\text{S}$ concentrations in the laboratory-based experiment with marine sediment from Site 2. Dashed white line indicates approximate position of sediment-water interface. The $\Sigma\text{H}_2\text{S}$ concentrations were determined by DGT hydrogel. The passive samplers with DGT gels were deployed overnight, due to the high concentrations of dissolved $\Sigma\text{H}_2\text{S}$ the sulfide binding agent likely was saturated. As such, the measurements in the lower parts of the gels represent underestimates; below $\sim 4 \text{ cm}$ sediment depth the dissolved $\Sigma\text{H}_2\text{S}$ concentration was at least 2 mmol L^{-1} .

Fe K-edge EXAFS spectra and ITFA-LCF results for samples from sulfidation experiment. The sample spectra and ITFA-LCF results are shown in Fig. S10, the relative contribution of the three components are listed in Table S2.

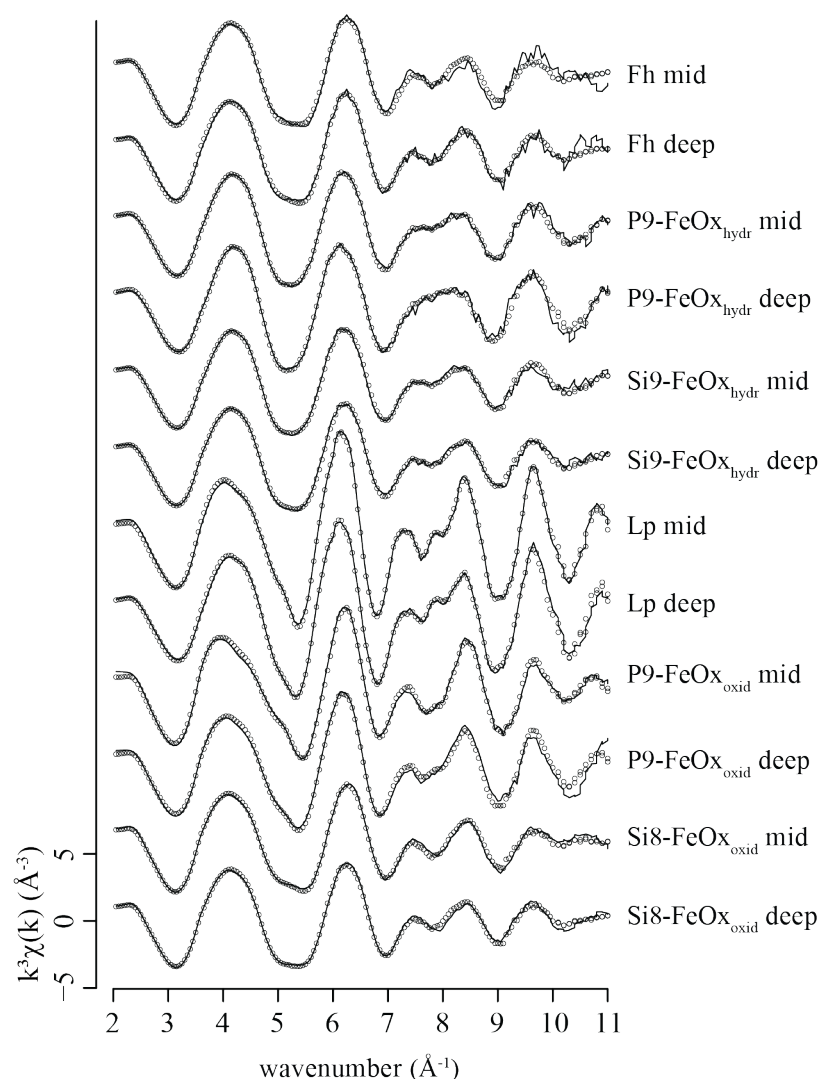


Fig. S10. Fe K-edge spectra of the samples from the microcosm sulfidation experiment (lines) and reproduction of the sample spectra using the three ITFA-derived components (dots). For each FeOx, gel sections 2 (“mid”, 3.5–7 cm) and 4 (“deep”, 10.5–14 cm) were analyzed.

Table S2. Weights of different ITFA-derived components in the reproduced spectrum. The Fe phase that each component represents is in parentheses. The weights represent the relative contributions of the different Fe phases (poorly ordered (P- or Si-bearing) FeOx, Lp and FeS).

		comp. 1	comp. 2	comp. 3
Sample		(Fh)	(Lp)	(FeS)
1	Fh mid	0.75	0.00	0.25
2	Fh deep	0.72	0.03	0.25
3	P9-FeOx _{hydr} mid	0.60	0.00	0.39
4	P9-FeOx _{hydr} deep	0.47	0.00	0.53
5	Si9-FeOx _{hydr} mid	0.65	0.00	0.35
6	Si9-FeOx _{hydr} deep	0.68	0.02	0.30
7	Lp mid	0.02	0.54	0.43
8	Lp deep	0.06	0.36	0.58
9	P9-FeOx _{oxid} mid	0.61	0.28	0.11
10	P9-FeOx _{oxid} deep	0.45	0.24	0.32
11	Si8-FeOx _{oxid} mid	0.82	0.01	0.18
12	Si8-FeOx _{oxid} deep	0.75	0.00	0.25

Fe K-edge XANES results for samples from sulfidation experiment. The DOS was determined from ITFA-LCF that produced a component representing FeS as the only ferrous Fe phase (Fig. S4). As such, analysis of the Fe K-edge EXAFS data did provide evidence for a significant presence of non-sulfidized, excess Fe (see section 3.3.2). To further ascertain whether the XAS data provided any indication of excess Fe(II), we also analyzed the Fe K-edge XANES data. The positions of the Fe absorption edges (AEs) of (sulfur-bound) ferrous Fe phases lie at lower energies than the AE of ferric Fe phases (Fig. S11A). We used biotite, a ferrous Fe-bearing aluminosilicate, and siderite, Fe(II)CO₃, as representatives of fully reduced, non-sulfidized Fe. In addition, iron monosulfide and pyrite are used to represent ferrous Fe sulfide. Adapting an approach to use AE shifts for Fe speciation²², we plot the energies at which the normalized AE of the different references is 0.5 ($E_{AE=0.5}$) (Fig. S11B). From this, we can discern that Fe(III) sulfidation results in a larger energy shift than Fe(III) reduction. We then tested for the presence of non-sulfidized excess Fe(II) by plotting $E_{AE=0.5}$ against DOS for samples from the lab-based sulfidation experiment. A large contribution of excess Fe(II) in addition to sulfidized Fe(II) would amplify the negative energy shift of the absorption edge, and consequently data points would plot below the sulfidation trendline. This is not observed: rather, data points plot close to or, particularly at lower DOS, slightly above the sulfidation trendline and align with the reduction trendline. Bearing in mind that caution has to be applied when interpreting such a small dataset, this could indicate that part of the energy shift for samples with relatively low DOS is in fact due to Fe(III) reduction rather than Fe(III) sulfidation (the former causes a smaller negative energy shift). The deviation from the sulfidation trendline is not observed for the samples with the highest DOS, which lends support to the idea that excess Fe(II) forms at the expense of FeS and is not observed when FeS formation is efficient (i.e. high DOS). We interpret these findings as tentative evidence for the presence of non-sulfidized Fe(II) in the samples from the sulfidation experiment. This would extend the model for different sulfidation pathways for FeOx based on batch laboratory experiments to natural sediments.

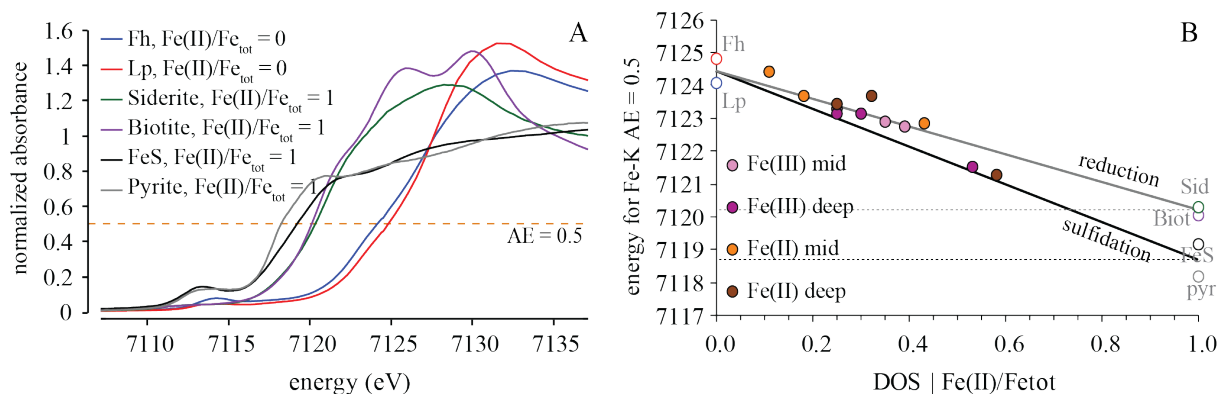


Fig. S11. (A) XANES spectra for selected ferric and ferrous references collected for this study and from a reference library collected at beamline 2-3 of the Stanford Synchrotron Radiation Lightsource (SSRL), kindly provided by Benjamin Kocar. The dashed orange line indicates normalized absorption edge (AE) = 0.5. (B) Energy at which AE = 0.5 ($E_{AE=0.5}$) plotted against degree of sulfidation (DOS) or fraction of reduced Fe for ferric and ferrous references (white fill, color corresponding to colors in panel A) and for samples from the sulfidation experiment (filled circles). The solid black and grey lines represent the trend in $E_{AE=0.5}$ associated with progressive sulfidation or reduction of ferric FeOx, respectively. The horizontal dashed black and grey lines indicate the AE = 0.5 energies for spectra of fully sulfidized or reduced Fe, respectively. Mid and deep refer to mid-depth and deep gel sections analyzed in this study (see also Fig. 5).

References used

- Michel, F. M.; Ehm, L.; Antao, S. M.; Lee, P. L.; Chupas, P. J.; Liu, G.; Strongin, D. R.; Schoonen, M. A. A.; Phillips, B. L.; Parise, J. B. The structure of ferrihydrite, a nanocrystalline material. *Science* **2007**, *316* (5832), 1726-1729.
- Hiemstra, T. Surface and mineral structure of ferrihydrite. *Geochim. Cosmochim. Acta* **2013**, *105* (0), 316-325.
- Schwertmann, U.; Cornell, S., *Iron oxides in the laboratory: Preparation and characterization*. Wiley: 1991; p 188.
- Kraal, P.; van Genuchten, C. M.; Behrends, T.; Rose, A. L. Sorption of phosphate and silicate alters dissolution kinetics of poorly crystalline iron (oxyhydr)oxide. *Chemosphere* **2019**, *234*, 690-701.
- Davis, C. C.; Chen, H.-W.; Edwards, M. Modeling Silica Sorption to Iron Hydroxide. *Environ. Sci. Technol.* **2002**, *36* (4), 582-587.
- Swedlund, P. J.; Webster, J. G. Adsorption and polymerisation of silicic acid on ferrihydrite, and its effect on arsenic adsorption. *Water Res.* **1999**, *33* (16), 3413-3422.
- Robertson, D.; Teasdale, P. R.; Welsh, D. T. A novel gel-based technique for the high resolution, two-dimensional determination of iron (II) and sulfide in sediment. *Limnol. Oceanogr. Meth.* **2008**, *6*, 502-512.
- Davison, W.; Zhang, H. In situ speciation measurements of trace components in natural waters using thin-film gels. *Nature* **1994**, *367* (6463), 546-548.
- Pagès, A.; Teasdale, P. R.; Robertson, D.; Bennett, W. W.; Schäfer, J.; Welsh, D. T. Representative measurement of two-dimensional reactive phosphate distributions and co-distributed iron(II) and sulfide in seagrass sediment porewaters. *Chemosphere* **2011**, *85* (8), 1256-1261.
- Santner, J.; Prohaska, T.; Luo, J.; Zhang, H. Ferrihydrite containing gel for chemical imaging of labile phosphate species in sediments and soils using diffusive gradients in thin films. *Anal. Chem.* **2010**, *82* (18), 7668-7674.

11. Stockdale, A.; Davison, W.; Zhang, H. High-resolution two-dimensional quantitative analysis of phosphorus, vanadium and arsenic, and qualitative analysis of sulfide, in a freshwater sediment. *Environmental Chemistry* **2008**, *5* (2), 143-149.
12. APHA, *Standard methods for the examination of water and wastewater*. 21 ed.; American Public Health Association - American Water Works Association - Water Environment Federation: 2005.
13. Bennett, W. W.; Teasdale, P. R.; Welsh, D. T.; Panther, J. G.; Jolley, D. F. Optimization of colorimetric DET technique for the in situ, two-dimensional measurement of iron(II) distributions in sediment porewaters. *Talanta* **2012**, *88*, 490-495.
14. Pages, A.; Teasdale, P. R.; Robertson, D.; Bennett, W. W.; Schafer, J.; Welsh, D. T. Representative measurement of two-dimensional reactive phosphate distributions and co-distributed iron(II) and sulfide in seagrass sediment porewaters. *Chemosphere* **2011**, *85* (8), 1256-61.
15. Kelly, S. D.; Hesterberg, D.; Ravel, B., Analysis of soils and minerals using X-ray absorption spectroscopy. In *Methods of Soil Analysis Part 5—Mineralogical Methods*, Ulery, A. L.; Drees, L. R., Eds. Soil Science Society of America: Madison, Wisconsin, 2008; Vol. SSSA Book Series.
16. Nikitenko, S.; Beale, A. M.; van der Eerden, A. M. J.; Jacques, S. D. M.; Leynaud, O.; O'Brien, M. G.; Detollenaere, D.; Kaptein, R.; Weckhuysen, B. M.; Bras, W. Implementation of a combined SAXS/WAXS/QEXAFS set-up for time-resolved in situ experiments. *J. Synchrotron Radiat.* **2008**, *15* (6), 632-640.
17. Ravel, B.; Newville, M. Athena, Artemis, Hephaestus: data analysis for X-ray absorption spectroscopy using IFEFFIT. *J. Synchrotron Radiat.* **2005**, *12*, 537-541.
18. Rossberg, A.; Reich, T.; Bernhard, G. Complexation of uranium(VI) with protocatechuic acid—application of iterative transformation factor analysis to EXAFS spectroscopy. *Anal. Bioanal. Chem.* **2003**, *376* (5), 631-638.
19. Kraal, P.; Burton, E. D.; Rose, A. L.; Kocar, B. D.; Lockhart, R. S.; Grice, K.; Bush, R. T.; Tan, E.; Webb, S. M. Sedimentary iron–phosphorus cycling under contrasting redox conditions in a eutrophic estuary. *Chem. Geol.* **2015**, *392* (0), 19-31.
20. Voegelin, A.; Senn, A.-C.; Kaegi, R.; Hug, S. J.; Mangold, S. Dynamic Fe-precipitate formation induced by Fe(II) oxidation in aerated phosphate-containing water. *Geochim. Cosmochim. Acta* **2013**, *117*, 216-231.
21. van Genuchten, C. M.; Peña, J.; Amrose, S. E.; Gadgil, A. J. Structure of Fe(III) precipitates generated by the electrolytic dissolution of Fe(0) in the presence of groundwater ions. *Geochim. Cosmochim. Acta* **2014**, *127*, 285-304.
22. Shoenfelt, E. M.; Sun, J.; Winckler, G.; Kaplan, M. R.; Borunda, A. L.; Farrell, K. R.; Moreno, P. I.; Gaiero, D. M.; Recasens, C.; Sambrotto, R. N.; Bostick, B. C. High particulate iron(II) content in glacially sourced dusts enhances productivity of a model diatom. *Science Advances* **2017**, *3* (6), e1700314.