



Phylogeographic study using autonomous reef monitoring structures indicates fast range expansion of the invasive bryozoan *Juxtacribrilina mutabilis*

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Abstract This is a phylogeographical study of *Juxtacribrilina mutabilis*, a recently described bryozoan from Japan with sightings in Sweden, Norway, and Maine (US), to test how fast and far it has expanded across Europe in recent years. *J. mutabilis* settles easily on ship hulls, making it a useful model organism for studying long-distance invasion pathways. The study was conducted using Autonomous Reef Monitoring Structures (ARMS) to monitor the entire European coastline and Svalbard for *J. mutabilis* using DNA metabarcoding. During the time between its first sighting in Europe 2008 and the last ARMS retrieval 2020, the data shows how *J. mutabilis* has gained a pan-European distribution, being genetically identified in 14 new distinct locations. Presence/absence of barcodes were confirmed by image-based identification in 74% of the samples. Fourteen haplotypes never reported before were discovered in a 264 bp region of the cytochrome oxidase I gene. Two haplotypes (HP1 and HP3) occurred frequently and geographically widely dispersed, indicating

intercontinental connectivity. Two locations, Koster and Getxo showed particularly high genetic diversity with similar haplotype networks suggesting continuous gene flow across oceanographically unlinked regions. Given the recent description of *J. mutabilis* and the relatively few historical encounters, the genetic diversity described here suggests an unusually fast range expansion within the last two decades. Such global spreading events of fouling organisms may become more common in the future as a result of more frequent and interconnected ship traffic.

Keywords *Cribrilina mutabilis* · Metabarcoding · ARMS · Northeast passage · Suez channel · Non-indigenous species

Introduction

Juxtacribrilina mutabilis (Ito et al., 2015), previously *Cribrilina mutabilis*, is an invasive cheilostome bryozoan. It was first described at lake Akkeshi (Japan) and has also been found in Kristineberg (Sweden), Bergen (Norway), and in Cosco Bay, Maine (US) (Ito et al., 2015). Before that it is thought to have been observed under the name *Membraniporella aragoi* (Audouin, 1826) in the Russian Northwest Pacific by Audouin (1826) and Kubanin (1975, 1977, 1997) (Dick et al., 2020). The north-eastern passage is a possible vector of invasion from Asia to Scandinavia, from where it most

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likely has spread to Maine. Its main reported impact is the fouling of eelgrass, which is its most common settling substrate, but it has also been found on other *Zostera* species, several algae, plastic panels and strips and ship hulls (Dick et al., 2020).

The ability to settle on ship hulls in conjunction with previous demonstrated spreading over large distances makes *J. mutabilis* a good model for studying invasion pathways and expansion rates associated with new shipping routes (Gunnarsson, 2021). In addition to new shipping routes, also the intensity of ship traffic has multiplied over the past decades (UNCTAD, 2018). Hence, we can expect to see new patterns of spreading, not only with regard to the number of species but also in the rate of expansion. Estimates of expansion rate can have important applications in conservation management. Quantitative measurements of invasion speed are difficult to produce as they are likely dependent on underlying evolutionary processes (Philips, 2015). Yet, understanding the scale at which species can invade regions will be important for defining windows of opportunity for counteraction. With climate change shown to cause profound marine species distribution shifts (Perry et al., 2005) we can expect even more unwarranted invasions in the near future, especially alongside ship traffic that connects previously incompatible ecosystems.

In this study we investigate the phylogeographic structure of *J. mutabilis* to test how far and how fast the species has expanded in European waters in the past decade. More specifically, we describe the species current distribution and genetic diversity in Europe, try to identify signatures of recent translocation of the species across the Pacific and Atlantic Ocean, and investigate its true place of origin. To this end we utilized the genetic monitoring data collected the European ARMS MBON program (www.arms-mbon.eu/), which deploys settling plates, so-called Autonomous Reef Monitoring Structures (ARMS) in ports, marinas, and nature reserves along the European coastline and in the polar regions. These plates are used to collect eDNA metabarcoding data (Ficetola et al., 2008) as well as photographic images of the settlement plates and allow for analysis of genetic diversity across benthic habitats in European waters (Obst et al., 2020).

Methods

Analysis of ARMS samples

Sixty-six ARMS samples for which all data are openly accessible were analyzed across 18 locations in Europe (Table 1). Samples were deployed, retrieved, photographed, and processed by individual partners in the ARMS-MBON network (www.arms-mbon.eu), while DNA extraction, PCR amplification and sequencing was done centrally following the Molecular Standard Operating Procedures (MSOP) published by the ARMS-MBON program (Obst et al., 2020), using the mlCOIintF and jgHCO2198 primers. To confirm any genetic match, the plate photographs of all 66 ARMS were scanned for presence/absence of *J. mutabilis*. In some cases, *J. mutabilis* was not identified on the images but on closer inspection there were instances where ancestrulas (initial stage of bryozoan colony development) were detected on the images. Ancestrulas could not be determined to species level, yet could potentially belong to *J. mutabilis*. The number of colonies per sample was not taken into account, as deployment time varied between the samples (Fig. 1).

Bioinformatic analysis was performed with a custom pipeline written in the R programming language (R-Core-Team, 2019) available in the Supplementary material. Raw sequencing files in fastq format (ENA accession numbers are given in Table 1) were processed with Dada2 (Callahan et al., 2016). Primer location and orientation was assessed, while primer removal was performed with Cutadapt (Martin, 2011). Quality profiles for all samples and markers were assessed with FastQC. Low-quality read ends (quality score < 2) were trimmed off. Furthermore, the error model was trained and evaluated, resulting in a good match between observed and expected errors. Thereafter de-replication and inference of Amplicon Sequence Variants (ASV) were performed, merging forward and reverse reads. Finally, chimeras and singletons were removed using “pseudo-pooling.” Remaining sequences were trimmed to a minimum length to exclude short reads (< 300 bp). After filtering, the number of output reads per sample varied widely from none to tens of thousands. Samples with less than 5,000 reads were considered to be of “poor quality,” but remained in the analysis.

Table 1 Genetic and photographic matches of *Juxtacribrilina mutabilis*

Location name (Country)	ARMS ID	Geographic coordinates	Depth (m)	Sample period	ENA accession numbers	Quality	Genetic match	Photomatch
Koster (SE)	VH1	58.8752, 11.1932	24	Apr 2018– May 2019	ERR4018466, ERR4018454, ERR4018455	+	+	+
Koster (SE)	VH2	58.8763, 11.1120	22	Apr 2018– May 2019	ERR4018451, ERR4018452, ERR4018453	+	+	+
Koster (SE)	VH3	58.8600, 11.0805	25	Apr 2018 – May 2019	ERR4018469, ERR4018479, ERR7127610	+	+	+
Koster (SE)	VH1	58.8752, 11.1932	24	May 2019– Aug 2020	ERR4914103, ERR4914104, ERR4914105	+	+	+
Koster (SE)	VH2	58.8763, 11.1120	22	May 2019– Aug 2020	ERR4914109, ERR4914110, ERR4914111	+	+	+
Koster (SE)	VH3	58.8600, 11.0805	25	May 2019– Aug 2020	ERR4914106, ERR4914107, ERR4914108	+	+	+
Lysekil (SE)	Preemraff1	58.3533, 11.4339	3	Apr 2020– Aug 2020	ERR4914154, ERR4914155, ERR4914156	+	+	+
Lysekil (SE)	Preemraff2	58.3540, 11.4339	3	Apr 2020– Aug 2020	ERR4914157, ERR4914158, ERR4914159	+	+	+
Marstrand (SE)	Marstrand1	58.9144, 11.5942	4–6	Feb 2020– May 2020	ERR4914136, ERR4914137, ERR4914138	+	+	–
Marstrand (SE)	Marstrand2	57.9035, 11.5816	4–6	Feb 2020– May 2020	ERR4914139, ERR4914140, ERR4914141	+	+	–
Marstrand (SE)	Marstrand3	57.8893, 11.5857	1–2	Feb 2020– May 2020	ERR4914142, ERR4914143, ERR4914144	+	+	+
Gothenburg (SE)	Gbg1	57.6648, 11.7147	5	Feb 2020– May 2020	ERR4914115, ERR4914116, ERR4914117	+	+	+
Gothenburg (SE)	Gbg2	57.6646, 11.7329	5	Feb 2020– May 2020	ERR4914118, ERR4914119, E RR4914120	+	+	+
Gothenburg (SE)	Gbg3	57.6805, 11.7406	5	Feb 2020– May 2020	ERR4914121, ERR4914122, ERR4914123	+	+	+
Gothenburg (SE)	Gbg4	57.6808, 11.7283	5	Feb 2020– May 2020	ERR4914124, ERR4914125, E RR4914126	+	+	+
Gothenburg (SE)	Hjuvik1	57.7932, 11.7114	2	Jan 2020– May 2020	ERR4914133, ERR4914134, E RR4914135	+	–	+
Gothenburg (SE)	Bjorko1	57.7180, 11.6800	1–2	Jan 2020– May 2020	ERR4914112, ERR4914113, E RR4914114	+	–	–
Varberg (SE)	Varberg1	57.1134, 12.2300	2	Feb 2020–Jun 2020	ERR4914145, ERR4914146, E RR4914147	+	–	–
Varberg (SE)	Varberg2	57.1126, 12.2303	2	Feb 2020–Jun 2020	ERR4914148, ERR4914149, E RR4914150	+	–	+
Varberg (SE)	Varberg3	57.1107, 12.2440	2	Feb 2020–Jun 2020	ERR4914151, ERR4914152, E RR4914153	+	–	–
Helsingborg (SE)	Helsingborg1	56.0263, 12.6957	2	Mar 2020–Jun 2020	ERR4914127, ERR4914128, E RR4914129	+	+	–
Helsingborg (SE)	Helsingborg2	56.0181, 12.7005	2	Mar 2020–Jun 2020	ERR4914130, ERR4914131, E RR4914132	+	–	–
Laesoe (DK)	Laesoe1	57.2569, 11.1419	10	Aug 2019– Aug 2020	ERR7127583, ERR7127586, E RR7127589	+	+	+
Laesoe (DK)	Laesoe2	57.2569, 11.1420	10	Aug 2019– Aug 2020	ERR7127592, ERR7127606, E RR7127595	+	+	+
Laesoe (DK)	Laesoe3	57.2569, 11.1420	10	Aug 2019– Aug 2020	ERR7127598, ERR7127601, E RR7127604	+	+	+
Limfjord (DK)	Yellow8	56.9013, 9.0573	4	Jun 2019–Oct 2020	ERR4018494, ERR4018495, E RR4018496	+	–	–
Limfjord (DK)	Red2	56.9016, 9.0552	3–4	Jun 2019–Oct 2020	ERR4018497, ERR4018498, E RR4018499	+	–	–
Limfjord (DK)	Green33	56.8999, 9.0566	4	Jun 2019–Oct 2020	ERR4018450, ERR4018451, E RR4018452	+	–	–

Table 1 (continued)

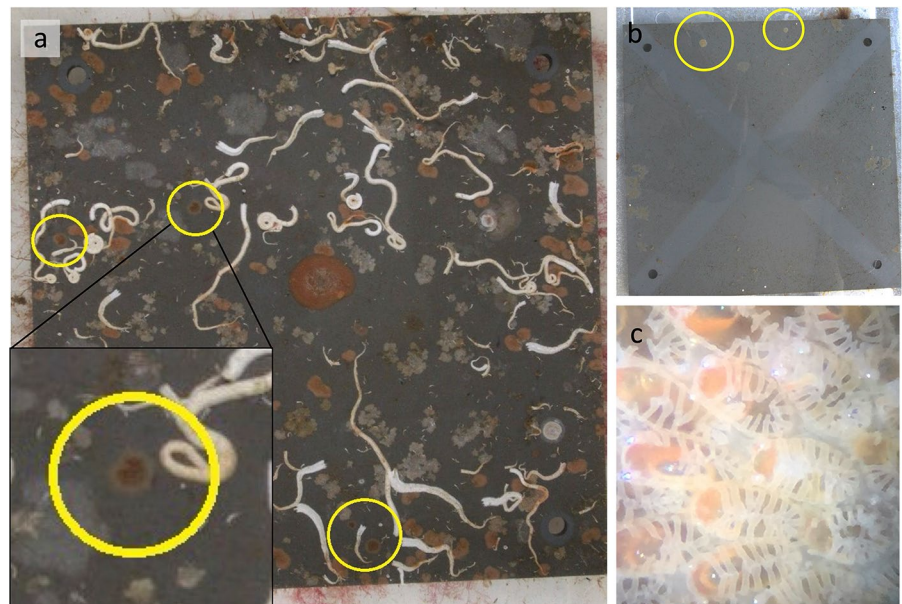
Location name (Country)	ARMS ID	Geographic coordinates	Depth (m)	Sample period	ENA accession numbers	Quality	Genetic match	Photomatch
Crete (GR)	1HERP	35.3432, 25.1366	4	Jun 2019–Oct 2019	ERR7125540, ERR7125534, ERR7125537	-	-	-
Crete (GR)	1HERP	35.3432, 25.1366	4	Jun 2019–Feb 2020	ERR7125549, ERR7125543, ERR7125546	+	+	-
Crete (GR)	1HERP	35.3432, 25.1366	4	Jun 2019 Jun 2020	ERR7125558, ERR7125552, ERR7125555	+	+	-
Crete (GR)	2UBPC	35.3466, 25.2788	21	Sep 2019–Dec 2020	ERR7125576, ERR7125570, ERR7125573	+	+	+
Svalbard (NO)	S1	78.2130, 15.2336	14	Jul 2018–Aug 2019	ERR7127659, ERR7127636, ERR7127662	+	+	-
Svalbard (NO)	S2	78.2130, 15.2336	14	Jul 2018–Aug 2019	ERR7127648, ERR7127661, ERR7127662	+	-	-
Svalbard (NO)	S1	78.2130, 15.2336	14	Aug 2019–Aug 2020	ERR7127642, ERR7127645, ERR7127660	+	-	-
Svalbard (NO)	S2	78.2130, 15.2336	14	Aug 2019–Aug 2020	ERR7127633, ERR7127664, ERR7127651	+	-	-
Toralla (ES)	TorallaA	42.2361, -8.7861	12	Jun 2019–Oct 2019	ERR4018716, ERR4018717, ERR7127691	+	+	+
Toralla (ES)	TorallaB	42.2333, -8.7833	12	Jun 2019–Oct 2019	ERR4018718, ERR4018719, ERR7127714	-	-	+
Toralla (ES)	TorallaC	42.2340, -8.7830	12	Jun 2019–Oct 2019	ERR4018720, ERR4018721	-	+	+
Getxo (ES)	G1	43.3385, -3.0148	4–10	Jun 2019–Oct 2019	ERR7127706, ERR7127704, ERR7127705	+	+	-
Getxo (ES)	G2	43.3385, -3.0149	4–10	Jun 2019–Oct 2019	ERR7127708, ERR7127676, ERR7127707	+	-	-
Getxo (ES)	G3	43.3384, -3.0148	4–10	Jun 2019–Oct 2019	ERR7127711, ERR7127709, ERR7127710, ERR7127685, ERR7127688	+	+	-
Angbat (FI)	Spikarna	59.8109, 23.2062	3	Jul 2018–Nov 2018	ERR4018732, ERR7125524, ERR7125525	+	-	-
Angbat (FI)	Kummelkobben	59.8645, 23.2652	3–4	Jun 2020–Nov 2020	ERR7125513, ERR7125510, ERR7125516	-	+	+
Angbat (FI)	Angbat	59.8415, 23.2489	4	Jul 2018–Nov 2018	ERR4018733, ERR7125522, ERR7125523	+	-	-
Angbat (FI)	Angbat	59.8415, 23.2489	4	Jun 2020–Nov 2020	ERR7125504, ERR7125501, ERR7125507	-	+	+
Plymouth (UK)	MBA1	50.3674, -4.1552	1	Jul 2018–Oct 2018	ERR7125591, ERR7125592, ERR7125593, ERR7125594, ERR7125595, ERR7125596	+	+	-
Plymouth (UK)	MBA2	50.3501, -4.1592	10	Jul 2018–Oct 2018	ERR4018650, ERR7125589, ERR7125590	+	+	-
Plymouth (UK)	MBA1A	50.3673, -4.1554	2	Jun 2019–Sep 2019	ERR4018669, ERR4018670, ERR4018671	+	-	-
Plymouth (UK)	MBA1B	50.3668, -4.1561	2	Jun 2019–Sep 2019	ERR4018672, ERR4018673, ERR4018674	+	-	-
Plymouth (UK)	MBA1C	50.3674, -4.1544	2	Jun 2019–Sep 2019	ERR4018675, ERR4018676, ERR4018677	+	-	+
Gulf of Piran (SI)	Loc1	45.5188, 13.5673	9	Aug 2018–Nov 2018	ERR7127578, ERR4018549, ERR7127580	+	-	-
Gulf of Piran (SI)	Loc1	45.5188, 13.5673	9	Aug 2018–Nov 2018	ERR7127579, ERR7127581, ERR7127651	+	-	-
Roscoff (FR)	BasBloS1	48.7287, -3.9589	12	Jul 2018–Oct 2018	ERR4018615, ERR4018616, ERR7125532, ERR7125631, ERR7125634, ERR7125633	+	-	-

Table 1 (continued)

Location name (Country)	ARMS ID	Geographic coordinates	Depth (m)	Sample period	ENA accession numbers	Quality	Genetic match	Photomatch
Roscoff (FR)	MarBloR1	48.7175, -3.9665	1	Jul 2018–Oct 2018	ERR7125640, ERR7125638, ERR7125641, ERR7125649, ERR7125643, ERR7125642	+	+	+
Roscoff (FR)	MarBloR1	48.7175, -3.9665	1	Jun 2019–Oct 2019	ERR4018635, ERR4018636, ERR4018637	+	-	-
Roscoff (FR)	MarBloR2	48.7175, -3.9665	1	Jun 2019–Oct 2019	ERR4018638, ERR4018639, ERR4018640	+	-	-
Roscoff (FR)	MarBloR3	48.7175, -3.9665	1	Jun 2019–Oct 2019	ERR4018641, ERR4018642, ERR4018642	+	-	-
Eilat (IL)	NR_1	29.5, 34.9	10	Oct 2018–Apr 2020	ERR7127629, ERR7127630, ERR7127631	-	-	-
Eilat (IL)	NR_2	29.5, 34.9	10	Oct 2018–May 2020	ERR7127632, ERR7127633, ERR7127634	-	-	-
Eilat (IL)	Katza_1	29.52, 34.93	10	Oct 2018–Jun 2020	ERR7127623, ERR7127624, ERR7127625	+	-	-
Eilat (IL)	Katza_2	29.52, 34.93	10	Oct 2018–Aug 2020	ERR7127626, ERR7127627, ERR7127628	+	-	-
Knokke-Heist (BE)	ZBE1	51.3645, 3.2070	7	Sep 2019–Mar 2020	ERR4914228, ERR7125498, ERR7125478	+	+	+
Knokke-Heist (BE)	JJC86	51.6534, 2.7918	38	May 2020–Oct 2020	ERR7125493, ERR7125484, ERR7125481	+	+	-
Knokke-Heist (BE)	JJC81	51.6534, 2.7918	38	May 2020–Oct 2020	ERR7125493, ERR7125484, ERR7125481	-	+	-
Knokke-Heist (BE)	ZFPin	51.6526, 2.7937	37	May 2020–Oct 2020	ERR7125490, ERR7125487	-	+	-

Sample period is the deployment and retrieval of ARMS. ENA accession numbers are used by ARMS-MBON for recording sequencing runs, one number per sample and fraction size. Positive quality means that the sample generate at least 5,000 genetic reads after processing. Positive genetic match means that at least one of the reads are interpreted as *J. mutabilis*. Positive photo match means that *J. mutabilis* was observed at least once on photos of that ARMS

Fig. 1 Examples of photographs of ARMS containing *Juxtacribrilina mutabilis*. Colonies have been highlighted with yellow circles. **a** ARMS plate from Koster deployed for 1 year, **b** ARMS plate from Gothenburg deployed for four months, **c** microscopic image of *J. mutabilis*



Taxonomic assignment of COI ASVs was done against the BOLD reference library (Ratnasingham & Herbert, 2007) using BOLDigger (Buchner & Leese, 2020). Matches of *J. mutabilis* of 97% or higher were singled out from the data and tracked throughout the various locations. The 97% threshold was chosen empirically, as most matches were >99% or <90%. Each ASV was checked for alternative matches manually in the BOLD database to rule out any incorrect taxonomic assignments. ASVs were screened for possible pseudogene copies using MACSE v2.05 (Ranwez et al., 2018) and the ‘enrichAlignment’ function with genetic code 5, allowing no in-frame stop codons, frameshifts, and a maximum of three deletions on the amino acid level. The reference alignment consists of curated COI reference sequences mined from BOLD, representing the sequence diversity of deposited bryozoan COI sequences—and is provided by the developers of MACSE (Delsuc & Ranwez, 2020). All ASVs could be aligned to the reference alignment, and hence no ASVs were considered pseudogenes. All ASVs matching *J. mutabilis* were submitted to the ASV portal of the Swedish Biodiversity Data Infrastructure (asv-portal.biodiversitydata.se/) and are also made available to GBIF (Martaeng et al., 2022).

Haplotype network analysis and haplotype map

In order to analyze frequency and relationships between the discovered genetic sequences, a haplotype network was calculated combining the public *J. mutabilis* sequences (downloaded from BOLD) with all unique COI sequences found in the ARMS data set. The method was the same as in Ito et al. (2015), a Tight Span Walker network created using PopArt (Leigh & Bryant, 2015), downloaded at <http://popart.otago.ac.nz/downloads.shtml> (last accessed December 2021). The original study used a region of 555 bp length (some 631 bp long), while the ARMS data only covered a 313 bp region. The overlap between the public sequences and the ARMS sequences was 264 bp. Hence, all sequences were aligned and trimmed in R and a haplotype network was calculated for the overlapping 264 bp region. In order to compare the haplotype composition between the two locations with highest diversity, two additional networks were calculated for the two locations with the highest genetic diversity, Getxo and Koster using the 264 bp

alignment. Finally, haplotype diversity was plotted on a map using QGIS (QGIS Development Team, 2022).

Statistical analysis

Statistical analysis was made for the whole dataset as well as for Getxo (ES) and Koster (SE) individually due to their notable haplotype diversities. First, measurements of DNA polymorphism; haplotype diversity *H_d* and nucleotide diversity Π (Nei, 1987), were calculated. Next, the following statistical tests were calculated: Tajima’s D-test (Tajima, 1989), Fu’s *F_s* test (Fu & Li, 1993), Ramos-Onsins, and Rozas’s *R₂* test and raggedness index (*rg*) (Ramos-Onsins & Rozas, 2002). All statistical calculations were done in DnaSP v6.12.03 (Librado & Rozas, 2009), utilizing coalescent simulations of 1000 re-sampling replicates. The null hypothesis for the tests was that the population is stable, while the alternative hypothesis is that the population is (rapidly) expanding.

Results

Of the 66 ARMS samples investigated, 9 were of poor genetic quality (delivering <5,000 sequence reads) but could still be positive for *J. mutabilis* (Table 1). The species was identified genetically and photographically in 23 samples, while we found only genetic signals in 13 samples, and only photographic evidence in 4 samples, resulting in overall 40 positive samples. The remaining 26 samples had no matches at all. Photographic presence/absence and genetic presence/absence was in concurrence in 49 of the 66 samples (74%). Overall, *J. mutabilis* was identified genetically in 14 new locations showing a pan-European distribution from Svalbard in the North to the Eastern Mediterranean in the South, and from the Atlantic coast in the West to the Gulf of Finland in the East (Fig. 2).

Overall, we retrieved 21 unique ASVs with a length of 313 bp from all ARMS samples. When aligned with the public sequences, the resulting 264 bp fragment returned 16 haplotypes (Fig. 2). In this data set, three haplotypes were previously known (HP1, HP2, and HP3), but only two of the sequences were recovered in the ARMS data (HP1, HP3). One (HP2) was not found in our data set. Hence, we included 17 globally known haplotypes in our

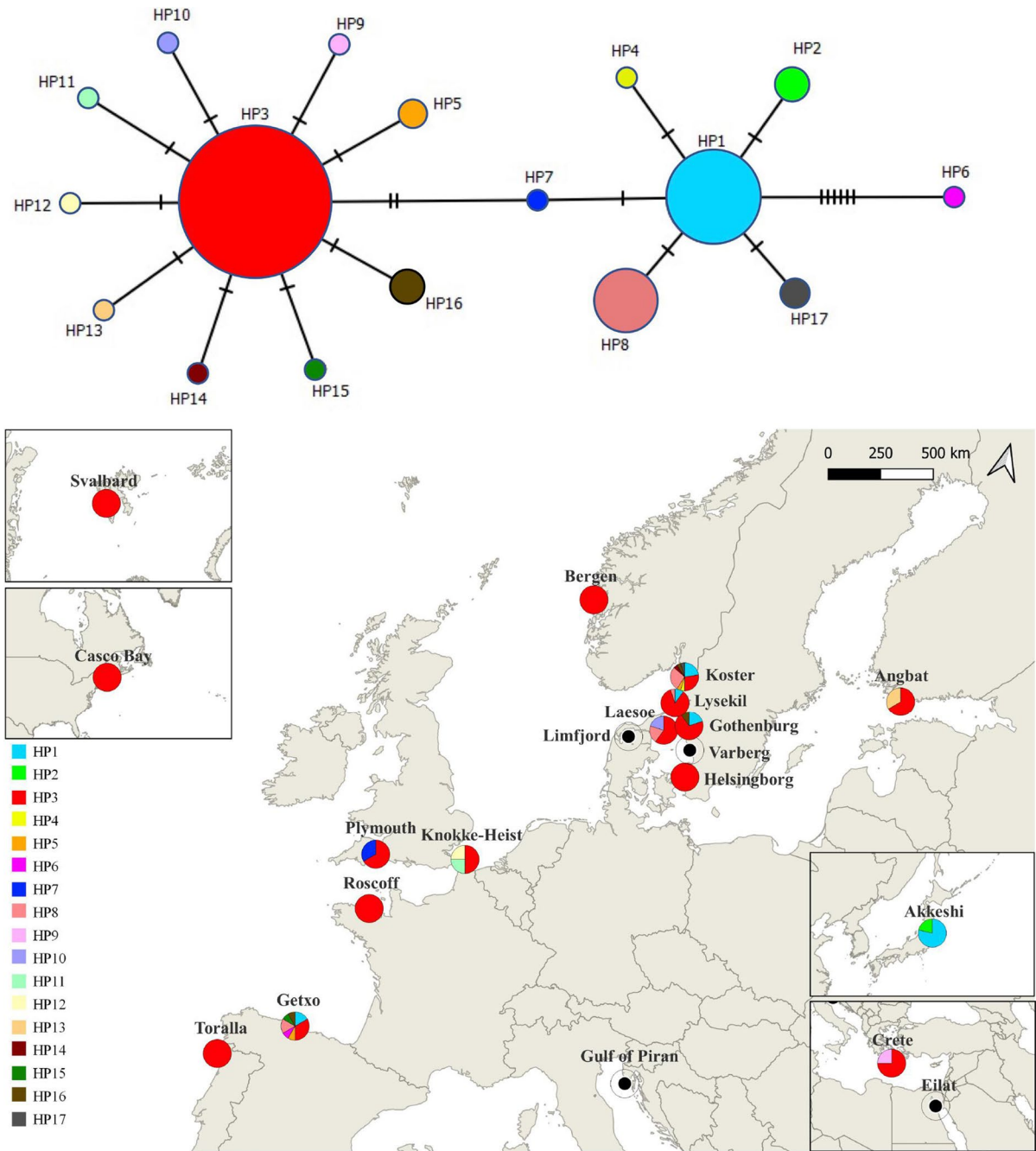


Fig. 2 (Top) Tight Span Walker haplotype network for *Juxtacribrilina mutabilis* using all 264 bp haplotypes found in our European eDNA metabarcoding dataset and GenBank. The sizes of circles are scaled to the number of samples. (Bottom) Haplotype map created in QGIS visualizing the locations of all known 264 bp haplotypes of *J. mutabilis* found in European

ARMS and public databases. For clarity, certain adjacent locations have been merged: Gothenburg (Gothenburg, Marstrand) and Lysekil (Lysekil, Kristineberg). A location with only a black dot indicates one or more ARMS samples without a genetic match for *J. mutabilis*

analysis of which 14 were newly discovered by this study (HP4–HP17). The complete haplotype network (Fig. 2) further expands on the network produced by Dick et al. (2020) with emphasis on European coastal water. The genetic centroids of the network (HP1 and HP3) are also the most frequent and geographically widespread haplotypes (Fig. 2; Table 2). HP1 is present at six locations while HP3 is present at every location except Akkeshi (JP).

Two locations, Getxo (ES) and Koster (SE) have notably high haplotype diversities: 7 and 8 haplotypes respectively (Table 2). The individual haplotype networks calculated for these locations (Fig. 3) are very similar, despite their geographical distance. Both networks share the same core structure with a few variant haplotypes in the periphery. The statistical analysis of the DNA polymorphism (Table 3) showed high haplotype diversity but low nucleotide diversity for all three sets (entire geographic range, Getxo, and Koster). All tests were, however, statistically insignificant in relation to various thresholds commonly associated with these tests and hence failed to reject the null hypotheses that the population is stable.

Discussion

Biogeography and range expansion

In this study we mapped distribution and genetic diversity of the recently described invasive species from Japan, *Juxtacribrilina mutabilis*, across Europe using standardized genetic sampling methods. Our results show considerable genetic diversity of *J. mutabilis* across its entire European distribution range. The high haplotype diversity in conjunction with low nucleotide diversity observed at Koster, Getxo and in all samples combined (Table 3) suggest a rapid recent expansion, i.e., a “population bottleneck followed by rapid population growth and accumulation of mutations,” according to Grant and Bowen (1998). But the statistical tests (e.g., Tajima’s D) failed to reject the null hypothesis that the population(s) are stable. A reason of the weak performance of the statistical tests may be the relatively short fragment size and the small sample sizes, especially for the locations Koster and Getxo.

The similar genetic signatures across distant locations described in this study (Figs. 2, 3; Tables 2, 3)

Table 2 Locations for each COI haplotype found in the dataset and in the original tissue samples

HP	Akk	Ang	Cas	Cre	Get	Got	Hel	Kno	Kos	Kri	Ku	Lae	Lys	Ma	Ply	Ros	Sva	Tor
1	X				X	X			X				X	X				
2	X																	
3		X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X
4									X									
5					X				X									
6					X													
7															X			
8					X				X			X	X					
9				X														
10												X						
11								X										
12								X										
13		X																
14									X									
15					X													
16									X									
17					X				X									

Only locations with at least one haplotype are shown. Names are abbreviated: Akkeshi (Akk), Angbat (Ang), Casco Bay (Cas), Crete (Cre), Getxo (Get), Gothenburg (Got), Helsingborg (Hel), Knokke-Heist (Kno), Koster (Kos), Kristineberg (Kri), Kummelkobben (Ku), Laesoe (Lae), Lysekil (Lys), Marstrand (Ma), Plymouth (Ply), Roscoff (Ros), Svalbard (Sva) and Toralla (Tor). Also visualized in a haplotype map, Fig. 2

Fig. 3 **a** Tight Span Walker haplotype network for *Juxtacribrilina mutabilis* using only 264 bp sequences found in ARMS from Koster (Sweden). **b** The respective haplotype network for ARMS from Getxo (Spain). In both networks, the sizes of circles are scaled to the number of samples, and haplotypes have been color coded to match Fig. 2

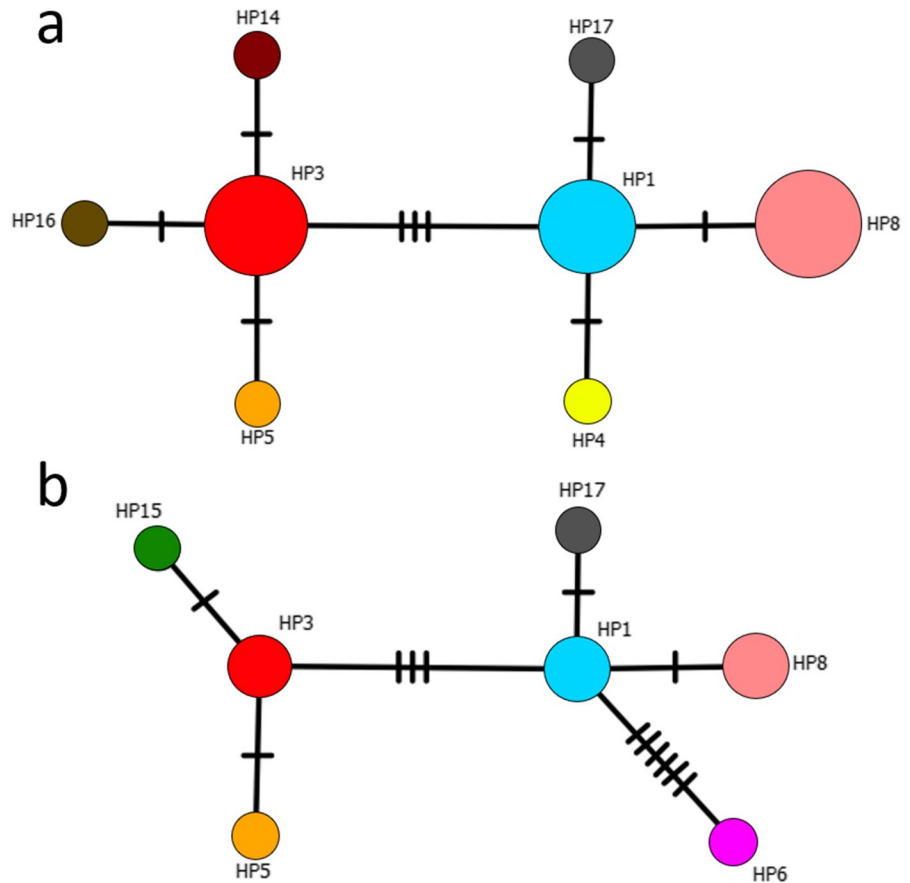


Table 3 Number of haplotypes (Nh), haplotype diversity (Hd), nucleotide diversity (Π), Tajima's D, Fu's FS, Ramos-Onsins and Rozas R2 test and raggedness index (Rg) for all samples, from Getxo and Koster

Group	Nh	Hd	Π	D	FS	R2	Rg
All	17	0.7427	0.01420	-0.0822	0.08036	0.1445	0.1434
Getxo	7	0.8049	0.01732	-0.0972	0.2480	0.2038	0.1952
Koster	8	0.7423	0.01218	-0.0259	0.3972	0.2007	0.2008

are indicative of recent and intense translocations of the species. HP3 is present at every location except Akkeshi, while the previous genetic distance of 6 bp per 555 bp between Japanese and European haplotypes shown by Dick et al. (2020) has been reduced to no distance at all. HP1 has been found in Akkeshi, Getxo, and at several Swedish locations. This intercontinental distribution of haplotypes HP1 and HP3 suggests swift migration between Japan, Europe, and North America. Furthermore, the similar haplotype networks in Getxo and Koster indicate continuous gene flow across oceanographically unlinked regions

in Europe, such as the Bay of Biscay and the Skagerrak. In both cases intense and regular ship traffic may provide a plausible explanation for the observed gene flow.

The species was first found in Europe in Bergen in 2008, but was not reported from Sweden until 2011 (Dick et al., 2020). Reasons for such non-observance could be taxonomic confusion with similar species such as *Callopora rylandi* or *Cribrilina annulata* caused by the polymorphism exhibited by *J. mutabilis* (Dick et al., 2021). Alternatively, *J. mutabilis* could have also been rare in Europe for a long time

and only recently became a native invasive. However, given the strong taxonomic expertise in the region we consider these scenarios unlikely and we see strong evidence that the species has invaded all regional seas of Europe (except the Black sea) within the 12 years between its first European sighting 2008 and the last retrieved ARMS data 2020. This rate of continental expansion outperforms several other marine invertebrate groups. *Carcinus maenas* for example showed its fastest regional expansion when it emigrated from San Francisco (California) to British Columbia over one decade (Carlton & Cohen, 2003). *Neogobius melanostomus*, another renowned invasive species, needed one decade to spread through the Great Lakes of North America, and approximately two decades to spread from the Baltic sea to the English Channel (Kornis et al., 2012). A third example is the invasive polychaete *Marenzelleria viridis* which was first sighted in Scotland in the late 1970s and which reached the Gulf of Bothnia and Gulf of Finland roughly two decades later (Maximov, 2011). All three species have a much shorter expansion range per decade than *J. mutabilis*.

In contrast to the recent expansion of *J. mutabilis* in European waters, the species may have been present in the Northwest Pacific for much longer, if the discoveries of Audouin (1826) and Kubanin (1975, 1977, 1997) are to be interpreted as *J. mutabilis*. Together, these historical records and the type locality in northern Japan suggest that the Northwest Pacific remains the most likely place of origin. However, in order to be certain about the origin of the species, genetic diversity of *J. mutabilis* should also be assessed in the Northwest Pacific.

In conclusion, it appears that *J. mutabilis* has in less than two decades expanded from a presumed native range around the coast of Japan to a globally distributed species showing the capacity to tolerate a great range of water temperatures and salinities and traveling across large distances with the help of ship traffic. The presented rate of expansion exceeds that of most other benthic invertebrates and may be indicative for the increased global connectivity due to intensified ship traffic and opening of new trading routes in the arctic. We therefore argue that biological monitoring programs as well as mitigation policies need to prepare for faster biological invasions.

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Author contributions RM conceived the study, carried out the research, drafted the manuscript, wrote much of the R code beyond the core functions, performed the main processing and analysis of the data and produced the figures and tables. MO contributed to the design and research, provided supervision and counseling as well as handling organization and communication. PK performed sampling, critically revised the manuscript and provided bryozoan expertise. See Acknowledgements.

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Declarations

Competing interests The authors have no relevant financial or non-financial interests to disclose.

Data availability The ARMS raw data (fastq.gz) analyzed in this study can be found in the European Nucleotide Archive (ebi.ac.uk/ena/) under ENA accession numbers provided in Table 1 (e.g., 'ERR4018466'). Amplicon Sequence Variants for *Juxtacribrilina mutabilis* generated in this study from the ARMS data is available in the Swedish Biodiversity Data Infrastructure (asv-portal.biodiversitydata.se/) under submission 'BOLD:ADV9830'. At time of writing there are no other published studies using this data, although the ARMS-MBON

metadata is likely part of several studies currently produced or reviewed.

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References

- Audouin, J. V., 1826. Explication sommaire des planches de polypes de l'Égypte et de la Syrie, publiées par Jules-Cesar Savigny. *Descr Égypte Hist Nat* 1: 225–244.
- Buchner, D. & F. Leese, 2020. BOLDigger—a Python package to identify and organise sequences with the Barcode of Life Data systems. *Metabarcoding and Metagenomics* 4: e53535.
- Callahan, B. J., P. J. McMurdie, M. J. Rosen, A. W. Han, A. J. A. Johnson & S. P. Holmes, 2016. DADA2: high-resolution sample inference from Illumina amplicon data. *Nature Methods* 13(7): 581–583. <https://doi.org/10.1038/nmeth.3869>.
- Carlton, J. T. & A. N. Cohen, 2003. Episodic global dispersal in shallow water marine organisms: the case history of the European shore crabs *Carcinus maenas* and *C. aestuarii*. *Journal of Biogeography* 30: 1809–1820. <https://doi.org/10.1111/j.1365-2699.2003.00962.x>.
- Delsuc, F. & V. Ranwez, 2020. Accurate alignment of (meta) barcoding datasets using MACSE. *Phylogenetics in the Genomic Era* 2.3: 1–31.
- Dick, M. H., A. Waeschenbach, T. J. Trott, T. Onishi, C. Beveridge, J. D. Bishop, M. Ito & A. N. Ostrovsky, 2020. Global distribution and variation of the invasive cheilostome bryozoan *Cribrilina mutabilis*. *Zoological Science* 37(3): 217–231. <https://doi.org/10.2108/zs190142>.
- Dick, M. H., A. V. Grischenk, D. P. Gordon & A. N. Ostrovsky, 2021. The *Cribrilina annulata* problem and new species of Juxtacribrilina (Bryozoa: Cheilostomata: Cribriliniidae) from the North Pacific. *Zootaxa* 5016(3): 333–364. <https://doi.org/10.11646/zootaxa.5016.3.2>.
- Ficetola, G. F., C. Miaud, F. Pompanon & P. Taberlet, 2008. Species detection using environmental DNA from water samples. *Biology Letters* 4(4): 423–425. <https://doi.org/10.1098/rsbl.2008.0118>.
- Fu, Y. X. & W. H. Li, 1993. Statistical tests of neutrality of mutations. *Genetics* 133(3): 693–709. <https://doi.org/10.1093/genetics/133.3.693>.
- Grant, W. A. S. & B. W. Bowen, 1998. Shallow population histories in deep evolutionary lineages of marine fishes: insights from sardines and anchovies and lessons for conservation. *Journal of Heredity* 89(5): 415–426. <https://doi.org/10.1093/jhered/89.5.415>.
- Gunnarsson, B., 2021. Recent ship traffic and developing shipping trends on the Northern Sea Route: Policy implications for future arctic shipping. *Marine Policy* 124: 104369. <https://doi.org/10.1016/j.marpol.2020.104369>.
- Ito, M., T. Onishi & M. H. Dick, 2015. *Cribrilina mutabilis* n. sp., an eelgrass-associated bryozoan (Gymnolaemata: Cheilostomata) with large variation in zooid morphology related to life history. *Zoological Science* 32(5): 485–497. <https://doi.org/10.2108/zs150079>.
- Kornis, M. S., N. Mercado-Silva & M. J. Vander Zanden, 2012. Twenty years of invasion: a review of round goby *Neogobius melanostomus* biology, spread and ecological implications. *Journal of Fish Biology* 80(2): 235–285. <https://doi.org/10.1111/j.1095-8649.2011.03157.x>.
- Kubanin, A. A., 1975. Bryozoa of the order Cheilostomata of the Peter the Great Gulf of the Japan Sea. In Zevina, G. B. (ed), *Fouling in the Japan and Okhotsk Seas Far Eastern Division of the Academy of Sciences of the USSR, Vladivostok*: 108–136. **(In Russian with English summary)**.
- Kubanin, A. A., 1977. Species composition of bryozoans in the fouling of ships with different floating regimes. *Biologiya Morya* 6: 64–69 **(English translation)**.
- Kubanin, A. A., 1997. Phylum Tentaculata, subphylum Bryozoa. In Kussakin, O. G., M. B. Ivanova & A. P. Tsurpalo (eds), *A Check-list of Animals, Plants and Fungi from the Intertidal Zone of Far Eastern Seas of Russia* Dalnauka Press, Vladivostok: 119–125. **(In Russian)**.
- Leigh, J. W. & D. Bryant, 2015. PopART: full-feature software for haplotype network construction. *Methods in Ecology and Evolution* 6(9): 1110–1116. <https://doi.org/10.1111/2041-210X.12410>.
- Librado, P. & J. Rozas, 2009. DnaSP v5: a software for comprehensive analysis of DNA polymorphism data. *Bioinformatics* 25(11): 1451–1452. <https://doi.org/10.1093/bioinformatics/btp187>.
- Martaeng, R., M. Obst & P. Kuklinski, 2022. *Juxtacribrilina mutabilis*. Version 1.2. University of Gothenburg. Occurrence dataset. [available on internet at: <https://doi.org/10.15468/y3upe9>].
- Martin, M., 2011. Cutadapt removes adapter sequences from high-throughput sequencing reads. *EMBnet Journal* 17(1): 10–12. <https://doi.org/10.14806/ej.17.1.200>.
- Maximov, A. A., 2011. Large-scale invasion of *Marenzelleria* spp. (Polychaeta; Spionidae) in the eastern Gulf of Finland Baltic Sea. *Russian Journal of Biological Invasion* 2: 11–19. <https://doi.org/10.1134/S2075111711010036>.
- Nei, M., 1987. *Molecular Evolutionary Genetics*, Columbia University Press, New York.
- Obst, M., K. Exter, A. L. Alcock, C. Arvanitidis, A. Axberg, M. Bustamante & C. Pavlodi, 2020. A Marine Biodiversity Observation Network for genetic monitoring of hard-bottom communities (ARMS-MBON). *Frontiers in Marine Science* 7: 1031. <https://doi.org/10.3389/fmars.2020.572680>.

- Perry, A. L., P. J. Low, J. R. Ellis & J. D. Reynolds, 2005. Climate change and distribution shifts in marine fishes. *Science* 308: 5730. <https://doi.org/10.1126/science.1111322>.
- Phillips, B. L., 2015. Evolutionary processes make invasion speed difficult to predict. *Biological Invasions* 17: 1949–1960. <https://doi.org/10.1007/s10530-015-0849-8>.
- QGIS Development Team (2022) QGIS Geographic Information System. Open Source Geospatial Foundation Project. [available on internet at: <http://qgis.osgeo.org>].
- Ramos-Onsins, S. E. & J. Rozas, 2002. Statistical properties of new neutrality tests against population growth. *Molecular Biology and Evolution* 19(12): 2092–2100. <https://doi.org/10.1093/oxfordjournals.molbev.a004034>.
- Ranwez, V., E. J. P. Douzery, C. Cambon, N. Chantret & F. Delsuc, 2018. MACSE v2: toolkit for the alignment of coding sequences accounting for frameshifts and stop codons. *Molecular Biology and Evolution* 35(10): 2582–2584. <https://doi.org/10.1093/molbev/msy159>.
- Ratnasingham, S. & P. D. N. Herbert, 2007. BOLD: the barcode of life data system. *Molecular Ecology Notes* 7.3: 355–364. <https://doi.org/10.1111/j.1471-8286.2007.01678.x>.
- Tajima, F., 1989. Statistical method for testing the neutral mutation hypothesis by DNA polymorphism. *Genetics* 123(3): 585–595. <https://doi.org/10.1093/genetics/123.3.585>.
- UNCTAD, 2018. 50 Years of Review of Maritime Transport, 1968–2018: Reflecting on the past, exploring the future. [available on internet at: https://unctad.org/system/files/official-document/dtl2018d1_en.pdf].

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