

# Iron enrichment stimulates toxic diatom production in high-nitrate, low-chlorophyll areas

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Oceanic high-nitrate, low-chlorophyll environments have been highlighted for potential large-scale iron fertilizations to help mitigate global climate change. Controversy surrounds these initiatives, both in the degree of carbon removal and magnitude of ecosystem impacts. Previous open ocean enrichment experiments have shown that iron additions stimulate growth of the toxigenic diatom genus *Pseudonitzschia*. Most *Pseudonitzschia* species in coastal waters produce the neurotoxin domoic acid (DA), with their blooms causing detrimental marine ecosystem impacts, but oceanic *Pseudonitzschia* species are considered nontoxic. Here we demonstrate that the sparse oceanic *Pseudonitzschia* community at the high-nitrate, low-chlorophyll Ocean Station PAPA (50° N, 145° W) produces approximately 200 pg DA L<sup>-1</sup> in response to iron addition, that DA alters phytoplankton community structure to benefit *Pseudonitzschia*, and that oceanic cell isolates are toxic. Given the negative effects of DA in coastal food webs, these findings raise serious concern over the net benefit and sustainability of large-scale iron fertilizations.

carbon sequestering | domoic acid | ocean iron fertilization | *Pseudonitzschia* | toxicity

Mesoscale iron fertilization experiments have been performed in all of the major high-nitrate, low-chlorophyll (HNLC) regions (1–3) (Table 1), and the biological and chemical consequences of the resultant diatom blooms have substantially improved our understanding of marine biogeochemical cycles, including their complexities and links to climate processes. The success of these scientific efforts has led to policy proposals and emerging commercial endeavors that enhance atmospheric carbon sequestration in the deep ocean using iron-induced diatom blooms with high sinking potential. Debate on the wisdom of these ecological manipulations has focused mainly on the magnitude and quantification of stimulated carbon export (1, 4) and the resultant effects on deepwater chemistry and biological communities (5). However, there has been little discussion whether the iron-enhanced diatom community composition itself may have unintended consequences to surface-dwelling organisms.

Mesoscale iron enrichment experiments have focused on studying the broader issue of carbon cycling, rather than assessing the potential ecological impacts of larger-scale and longer-term geo-engineering-designed fertilizations. Here we consider that the large diatom blooms generated by previous iron enrichment experiments have in most cases been dominated by diatoms belonging to the genus *Pseudonitzschia* (Table 1), including studies conducted in the equatorial Pacific [IronEx II (6, 7)] and Southern Ocean [SOIREE (8), EisenEx (9), and SoFeX (10, 11)], and may have the capacity to create unanticipated ecological consequences. *Pseudonitzschia* species have the capacity to produce the potent neurotoxin domoic acid (DA), sometimes generating massive toxic harmful algal blooms (HABs) in coastal waters (12–14). To our knowledge, there have been only two attempts to measure DA production during iron fertilization experiments: Assmy et al. (15) reported that there were no measureable quantities of DA during the EisenEx fertilization, and Marchetti et al. (16) found no DA production in laboratory cultures

of oceanic *Pseudonitzschia* isolates obtained after iron fertilization. However, the EisenEx assessment used stored, preserved phytoplankton samples, and the longer-term stability of DA in these samples is unknown. Furthermore, *Pseudonitzschia* spp. are notoriously variable with respect to toxin production in laboratory cultures, with supposedly “non-toxic” isolates suddenly revealing their capacity to produce DA and vice versa. Nevertheless, these negative results, combined with recognition that some coastal *Pseudonitzschia* produce low concentrations of DA (17), has led to a perception that oceanic *Pseudonitzschia* are nontoxic and that iron-generated blooms in HNLC regions will not negatively impact surface dwelling organisms of the pelagic ecosystem. This position would require reassessment if oceanic *Pseudonitzschia* indeed had the capacity to produce DA.

We present here results from in situ measurements and ship-board culture experiments demonstrating that oceanic *Pseudonitzschia* species produce DA and retain that capacity upon iron and copper amendment. The findings demonstrate that toxin production can occur with iron fertilization of HNLC waters, that the specific composition of commercial iron substrates is a critical parameter in the degree of toxin production, and that the total toxin production potentially could reach ecologically harmful levels during large-scale iron fertilization programs.

## Results

**Measurement of Cellular Domoic Acid in Subarctic Pacific Surface Waters.** Biological and physical data were collected at Ocean Station PAPA (OSP; 50° N, 145° W) in the eastern subarctic Pacific HNLC region during May and June 2006 in a project studying the regulation of phytoplankton community structure by low-level iron amendments. *Pseudonitzschia turgidula* and the smaller *P. cf granii* were present in low concentrations (<1,000 cells L<sup>-1</sup>) at the surface (5 m) and at the chlorophyll maximum (approximately 45–60 m). No other *Pseudonitzschia* species were observed during this study. We collected 1-L water samples at each depth and measured particulate DA (pDA) on fresh samples on board the ship using an indirect competitive ELISA (cELISA) method (18). Particulate DA was measured at 5 m depth over 6 consecutive days (Fig. 1), often at concentrations >30 pg L<sup>-1</sup>, which is in excess of the minimum analytical method detection limit (MDL; 25.6 pg L<sup>-1</sup> for 1-L seawater samples; 6.4 pg mL<sup>-1</sup> detectable concentration in the extract). The vertical profile of pDA concentrations (June 5, 2006; Fig. 2) indicates that DA is found in cells throughout the photic zone, but DA concentration does not necessarily co-vary with the

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**Table 1. Comparison of the present, deck-board Fe-enrichment experiments with previous open-ocean mesoscale Fe enrichment experiments**

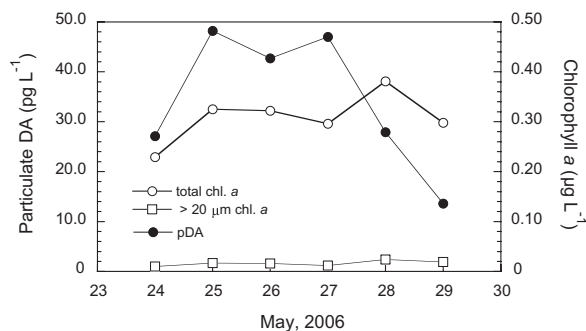
HNLC Region Project	Year	Dominant phytoplankton within bloom	Initial Chl a ( $\mu\text{g L}^{-1}$ )	Peak Chl a ( $\mu\text{g L}^{-1}$ )	Refs.
Equatorial Pacific IronEx-I	1933	<i>Synechococcus</i> red-fluorescing picoplankton autotrophic dinoflagellates	0.24	>0.65	37, 38
IronEx-II	1995	<i>Pseudonitzschia</i> spp. (Initially reported as <i>Nitzschia</i> )	0.15–0.20	<4.0	6, 7, 39
Northwest Pacific SEEDS-I	2001	<i>Chaetoceros debilis</i>	0.8–0.9	21.8	40, 41
SEEDS-II	2004	<i>Pseudonitzschia</i> spp., <i>Neodenticula semina</i>	0.80	2.48	42
Northeast Pacific SERIES	2002	<i>Pseudonitzschia</i> spp., <i>Neodenticula semina</i> , <i>Chaetoceros</i> spp.	0.27–0.43	6.30	43, 44
PAPA-SEEDS	2006	<i>Pseudonitzschia turgidula</i>	0.2–0.3	–	Present study
Southern Ocean SOIREE (Australian polar waters)	1999	<i>Fragilariopsis kerguelensis</i> , <i>Rhizosolenia</i> spp., <i>Pseudonitzschia</i> sp.	0.25	3.0	8, 45
EisenEx (Atlantic polar waters)	2000	<i>Pseudonitzschia lineola</i>	0.48–0.56	2.80	9, 15
SOFEX – North (Pacific subpolar waters)	2002	<i>Pseudonitzschia</i> spp.	0.26	2.60	10, 11
SOFEX – South (Pacific polar waters)	2002	<i>Fragilariopsis</i> spp., <i>Corethron</i> spp., <i>Chaetoceros</i> spp., <i>Rhizosolenia</i> spp.	0.20	4.0	10, 11
EIFEX (Atlantic subpolar waters)	2004	<i>Chaetoceros</i> spp., <i>Fragilariopsis kerguelensis</i> , <i>Pseudonitzschia</i> spp., large single-celled diatoms	0.54	2.85	46, 47
SAGE (Pacific subpolar Si-depleted coastal waters)	2004	–	~0.6–0.4	1.3 > 0.7	2, 48

The present, deck-board Fe-enrichment experiment was compared with previous open-ocean mesoscale Fe enrichment experiments conducted in HNLC regions of the Equatorial and North Pacific Oceans, and Southern Oceans with respect to initial and peak biomass [as chlorophyll a (Chl a)], and the dominant phytoplankton species or genera that increased in response to alleviation of in situ Fe-deficient conditions.

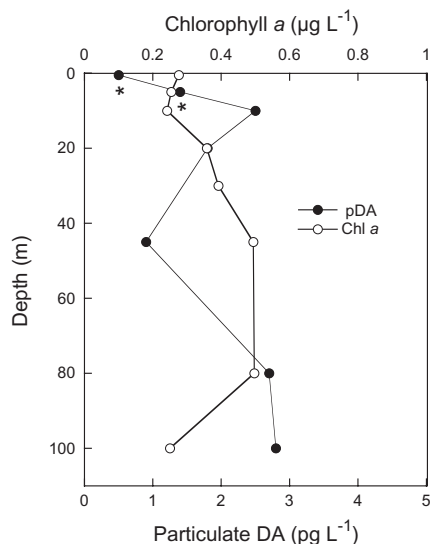
total phytoplankton community biomass (as measured by extracted chlorophyll *a*). Over the sampling period, DA concentrations periodically were substantially higher at the chlorophyll maximum (approximately 40 m): 59.0  $\text{pg L}^{-1}$  on May 25, 2006, and 36.3  $\text{pg L}^{-1}$  on May 27, 2006. Converting pDA concentrations to cellular DA (pDA normalized to cell abundance) by using microscopically determined total *Pseudonitzschia* cell abundances yields an estimated minimum cell toxin quota of approximately 15  $\text{fg DA cell}^{-1}$ , as only one species was observed, and assuming that all cells present were producing DA equally. These findings confirm that oceanic *Pseudonitzschia* species produce DA in situ and that DA accumulation by these cells is constitutive with their presence in HNLC waters.

**Effect of Iron and Copper Amendments on Toxin Production by *Pseudonitzschia* in Batch Cultures.** Previous studies have indicated that DA production and release into seawater by *Pseudonitzschia* species are linked to iron and copper nutrition of the cell (19, 20), so it is important to determine if these cells alter their toxicity or capacity to become toxic upon short-term alleviation of metal limitation. We tested this hypothesis using bottle-type growth experiments with OSP surface waters. *Pseudonitzschia* cell abundance and pDA concentrations increased substantially in the control bottles after 6 d (approximately 110  $\text{fg mL}^{-1}$ ) relative to in situ levels. The pDA concentrations were further enhanced by a small (1 nM) amendment of iron (approximately 160  $\text{fg mL}^{-1}$ ), whereas the combined addition of level Fe (1 nM) and Cu (10 nM) increased pDA by approximately 160% (approximately 290  $\text{fg mL}^{-1}$ ) relative to that in the unamended control bottles (Fig. 3). Cellular DA concentrations followed the same trend, with the control having approximately 0.1  $\text{fg DA cell}^{-1}$ , which doubled to approximately

0.2  $\text{fg DA cell}^{-1}$  with iron amendment and tripled to approximately 0.3  $\text{fg cell}^{-1}$  with the combined iron and copper amendment (Fig. 3). Two isolates of *P. turgidula* were successfully returned ashore and also produced DA in laboratory cultures, although at lower concentrations (0.0052  $\text{fg DA cell}^{-1}$ ) than those measured in the natural in situ population. These concentrations are below the analytical detection limits of the method employed by Marchetti et al. (16), who noted that DA was undetectable in their isolates.

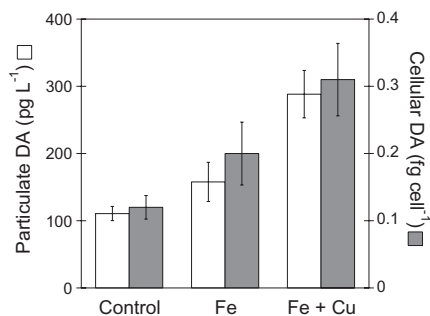


**Fig. 1.** Ambient concentrations of particulate DA and size-fractionated phytoplankton biomass as a function of time at OSP in May 2006. Discrete samples were collected using 10-L Niskin bottles; 1-L subsamples were filtered for pDA (●) and analyzed fresh at sea using a cELISA; 0.2-L and 0.4-L subsamples were filtered in parallel for phytoplankton biomass as chlorophyll *a* on Whatman GF/F filters (nominal pore size, 0.7  $\mu\text{m}$ ) for total chlorophyll *a* (○) and on Poretics 20  $\mu\text{m}$  pore size polycarbonate filters for the >20- $\mu\text{m}$  fraction (□), respectively, and analyzed fluorometrically (33). Values below the volume adjusted MDL of 25.6  $\text{pg L}^{-1}$  (for 1-L samples) are reported with an asterisk.



**Fig. 2.** A profile of particulate DA and phytoplankton biomass as a function of depth at OSP in June 2006. Discrete samples were collected using 10-L Niskin bottles; generally 7.3-L subsamples were filtered for pDA (●) and analyzed fresh at sea using a cELISA; 0.2-L subsamples were filtered for total phytoplankton biomass as chlorophyll a (○) on Whatman GF/F filters (nominal pore size, 0.7 µm) and analyzed fluorometrically (33). Values below the volume-adjusted MDL of 1.8 pg L<sup>-1</sup> for all samples (7.3 L), except the 45 m depth sample (50 L) (volume-adjusted MDL, 0.5 pg L<sup>-1</sup>), are reported with an asterisk.

**Domoic Acid Production in Continuous Cultures at Ocean Station PAPA.** The chemical speciation of iron regulates its availability to phytoplankton, and organically complexed species that dominate iron speciation in the HNLC regions are poorly available to large eukaryotic phytoplankton (21). Pulsed iron input, which occurs in aerosol deposition events or rapid iron spikes to batch cultures and mesoscale iron enrichment experiments, can yield colloidal iron precipitates (2, 3) that increase iron availability to eukaryotic phytoplankton (21). In the case of mesoscale Fe enrichments, this increase is short-lived, as the bulk of these labile colloidal precipitates aggregate and sink, whereas the remaining Fe colloids dissolve and equilibrate with less available strongly complexed Fe (III) species (3). To better evaluate the effects of elevated con-

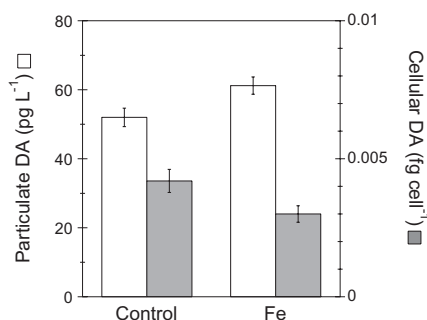


**Fig. 3.** Particulate and cellular DA concentrations in deck-board batch cultures containing *P. turgidula*. Basal growth media for the batch cultures was OSP seawater collected using trace metal-clean techniques. Treatments were: no metal amendments (control), amended with iron (1 nM Fe), or amended with iron and copper (1 nM Fe + 10 nM Cu). Particulate DA and *Pseudonitzschia* abundances were determined after 6 d of incubation when cells were in the stationary phase of growth. Open and solid bars are particulate DA (pDA) and cellular DA (pDA normalized to *Pseudonitzschia* cell abundance) concentrations, respectively. The volume adjusted MDL for analysis of pDA is 66 pg L<sup>-1</sup>.

centrations of ambient complexed Fe(III) species, we used ship-board continuous cultures comprising duplicate controls (filtered seawater) and iron amendments. Here, iron (1 nM) was pre-equilibrated (≥24 h) with the natural suite of organic ligands in filtered seawater before being supplied to cultures at a dilution rate of 0.5 d<sup>-1</sup> (2 d residence time). Under this growth scenario, the communities competing for the limiting nutrient are constantly replaced by being flushed from the culture vessels by new inflowing media, so that in the absence of significant grazing activity, the best-adapted (i.e., the fastest growing) species are retained in the highest numbers.

Adding iron doubled the steady-state abundance of *Pseudonitzschia* species after 9 d compared with controls (approximately 25,000 vs. 12,400 cells mL<sup>-1</sup>), consistent with trends in batch culture findings here and elsewhere (Table 1). In this case, the resulting *Pseudonitzschia* was not *P. turgidula*, but rather a smaller species identified by scanning electron microscopy as *P. cf granii*. Similar to *P. turgidula*, this small oceanic diatom was the only *Pseudonitzschia* species induced and produced DA in both control and iron-amended continuous cultures (Fig. 4). The cellular toxicity of the *P. cf granii* was 0.004 fg DA cell<sup>-1</sup> in the controls and changed little in the iron treatment (0.003 fg DA cell<sup>-1</sup>; Fig. 4). Even so, previous work has shown that much of the DA produced by cells during exponential growth is released to the dissolved phase (24) (in contrast to the increasing cellular toxicity observed when cell division slows in batch cultures). Dissolved DA was not detectable in our continuous cultures as a result of the high daily medium dilution rate (0.5 d<sup>-1</sup>).

**Ecological Consequences of Increased Domoic Acid Production and Release by *Pseudonitzschia* in the HNLC Eastern Subarctic Pacific.** The metabolic function of DA in *Pseudonitzschia* remains to be determined, but there is evidence that it is involved in the acquisition of iron and copper. Both actively growing and senescent cells release dissolved DA (17, 19, 20), with actively growing cells releasing the majority of the DA produced (24) (i.e., intracellular DA concentrations represent a small proportion of total DA production). Field measurements of dDA have reached >100 nM in Monterey Bay, California (22), and >400 nM in Sequim Bay, Washington (23). The release of DA can increase iron uptake rates by *Pseudonitzschia* and increased copper availability has been shown to facilitate iron uptake in this genus (19). It is reasonable then to ask whether production and release of DA might provide oceanic *Pseudonitz-*

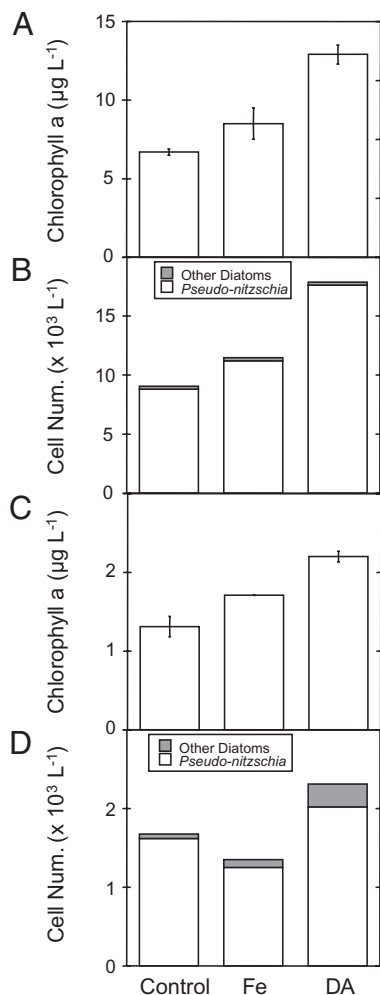


**Fig. 4.** Particulate and cellular DA concentrations in deck-board continuous cultures containing *P. cf granii*. Basal growth media for the continuous cultures was OSP seawater collected using trace metal-clean techniques. Treatments were: not amended with metals (control) or amended with iron (1 nM Fe). Culture dilution rates were maintained at 0.5 d<sup>-1</sup>. Particulate DA and *P. cf granii* abundances were determined after 9 d of incubation. Open and solid bars are particulate DA (pDA) and cellular DA (pDA normalized to *Pseudonitzschia* cell abundance) concentrations, respectively. The volume adjusted MDL for analysis of pDA is 37 pg L<sup>-1</sup>.

*schia* a competitive advantage over other diatoms in these iron-limited waters.

We tested the effect of dDA on diatom community response in continuous cultures at OSP in May 2007 using the same general methods described earlier. As in 2006, our experiments began before the spring bloom, and a significant increase in chlorophyll *a* biomass (10%) was observed in Fe-amended samples (Fig. 5A). However, addition of dissolved DA (25 nM) generated a twofold increase in total chlorophyll *a* biomass, apparently by increasing the relative availability of the ambient organically complexed dissolved iron. Diatoms in the control, iron, and DA treatments were dominated by *Pseudonitzschia* spp., with very minor contributions by *Thalassiosira* and *Chaetoceros* spp. (Fig. 5B).

We repeated this continuous culture experiment using surface waters from the central region of an anticyclonic mesoscale eddy in the Gulf of Alaska (GOA) in May 2007. These eddies advect excess nutrients and iron in their core waters from the continental margin as they propagate into the GOA and account for as much as 10% of the surface chlorophyll signal in the GOA.



**Fig. 5.** Effects of Fe and dissolved DA on the relative abundance of phytoplankton and *Pseudonitzschia* spp. Changes in phytoplankton biomass (A) (estimated by chlorophyll *a* concentrations) and phytoplankton cell abundance (B) (cells mL<sup>-1</sup>) of continuous cultures grown in OSP seawater in May 2007. Changes in phytoplankton biomass (C) (estimated by chlorophyll *a* concentrations) and phytoplankton cell abundance (D) (cells mL<sup>-1</sup>) of continuous cultures grown in GOA mesoscale eddy (May 2007) seawater under continuous dilution rates of 0.5 d<sup>-1</sup> for 6 d. Treatments were: not amended with metals (control), amended with iron (1 nM Fe), or amended with DA (25 nM).

Adding iron to these surface waters also enhanced chlorophyll *a* biomass by approximately 25% over that in the control continuous cultures (Fig. 5C). Diatom biomass followed this general trend, with *Pseudonitzschia* again dominating numerically over the *Chaetoceros* and *Tropodoneis* spp. present in surface waters (Fig. 5D). Addition of dDA further increased chlorophyll *a* biomass (approximately 40%) and *Pseudonitzschia* abundance (approximately 20%) over that in the controls (Fig. 5C and D), consistent with the findings at OSP.

## Discussion

The issue of potential toxin production during the massive ecosystem responses to mesoscale Fe enrichments of HNLC waters has attracted little attention, primarily because the limited evidence available indicated that oceanic *Pseudonitzschia* species lacked the ability to produce DA (16). The findings here show unequivocally that *Pseudonitzschia* species in the eastern subarctic Pacific do indeed produce DA as part of their normal physiology under ambient conditions, at least during spring. Particulate DA was measured over several days in surface waters and was distributed throughout most of the photic zone, despite *Pseudonitzschia* comprising an exceedingly small numerical component of the phytoplankton assemblage. Moreover, growth stimulation by iron additions selected for communities enriched by two *Pseudonitzschia* species (*P. turgidula* and *P. cf granii*) that not only retained the capacity to produce DA but also increased their cellular DA concentrations. The experimental outcomes here are consistent with findings from coastal waters and indicate that the physiological regulation of toxin production in *Pseudonitzschia* is similar between oceanic and coastal species (17). The central question now is not whether oceanic *Pseudonitzschia* are toxigenic but whether enhanced toxicity upon iron enrichment may reach levels sufficient to cause ecosystem damage.

The cellular DA concentrations generated during our deck-board experiments lie within the range of toxic blooms in coastal waters (17), but ascertaining whether these toxin levels may generate ecosystem impacts in oceanic waters is not straightforward. Krill (24) and planktivorous fish (25) feeding on *Pseudonitzschia* in coastal waters accumulate DA, which affects their feeding responses (26) and increases mortality in the food web (27), but we are aware of no equivalent studies for offshore waters. Although the total toxin concentrations generated in some of these preliminary experiments (approximately 0.2 µg DA L<sup>-1</sup>) may not be enough to generate acute toxicity at higher trophic levels, it is unclear whether higher concentrations may result under conditions of large-scale and persistent iron enrichments designed to obtain carbon credits. For example, toxin concentrations increased when iron amendments included enhanced copper availability, indicating that synergistic effects with other trace metals may occur. All mesoscale enrichment experiments to date have used expensive reagent-grade (i.e., pure) iron substrates (1, 3), but costs will necessitate that industrial-grade iron substrates (that may contain Cu or other trace metals) be used for proposed larger-scale enrichment experiments, so it is conceivable that bloom toxicity might be enhanced.

Our results indicate that the manner in which iron is added to HNLC surface waters will have an effect on toxin production by *Pseudonitzschia*. Batch cultures, which mimic aerosol deposition and mesoscale iron dispersment strategies, yielded similar results to continuous cultures, in which added iron equilibrated with natural ligands for ≥24 h before presentation to the phytoplankton community, an equilibration time far longer than that used in voltammetric methods employed to study iron speciation in seawater. Continuous culture experiments are a better mimic for the proposed enhanced nutrient (and metal) mixing of deep waters to the surface (28). In both cases, iron additions selected for a *Pseudonitzschia* species over other diatoms and resulted in increased concentrations of DA.

What can the findings here tell us about the potential for acute toxicity of *Pseudonitzschia* blooms resulting from iron fertilization? Commercial carbon credit goals plan to use Fe enrichment of approximately 100 km<sup>2</sup> of HNLC waters to sequester approximately 0.63 to 1.1 × 10<sup>7</sup> gC km<sup>-2</sup> in carbon offsets (29). The cellular carbon quota (in gC/cell) of actively growing *Pseudonitzschia* (8) is reasonably well established at approximately 5 × 10<sup>-14</sup> gC μm<sup>-3</sup> cell<sup>-1</sup>. The cell volume of *P. turgidula* in our cultures was 400 fL cell<sup>-1</sup> (30), corresponding to a cellular carbon quota of 2 × 10<sup>-11</sup> gC cell<sup>-1</sup>. Assuming a mixed layer depth of 80 m (measured here) and an estimated carbon export efficiency below 500 m of 5% (deep enough to ensure multidecadal sequestration of carbon) (11), sequestration of approximately 1 × 10<sup>7</sup> gC km<sup>-2</sup> would require blooms of approximately 6 × 10<sup>7</sup> cells L<sup>-1</sup>. These cell concentrations are comparable with large *Pseudonitzschia* blooms in coastal waters (12, 13) and presumably would be representative of those resulting from intensive iron fertilization. Although cellular toxin levels are highly variable, the reported cellular DA quota for *P. turgidula* in situ ranges between approximately 15 fg cell<sup>-1</sup> (measured here) to approximately 30 fg cell<sup>-1</sup> (31). Although these levels appear to be lower under rapid growth conditions, *Pseudonitzschia* cellular toxin loads generally increase as cell division slows in both batch and continuous cultures (32). Large-scale iron-induced blooms in these HNLC waters conceivably could yield toxin levels up to approximately 1 to 2 μg DAL<sup>-1</sup>, a level sufficient to close shellfish harvests and cause acute toxic effects on seabirds and marine mammals in near-shore waters (24–27).

Although there are clear uncertainties in this calculation, it is worth remembering that bloom toxicity in coastal waters tends to increase substantially as cells move into senescence, so actual cellular toxin quotas might be much greater than observed in these actively growing cultures. Even so, chronic sublethal exposure to DA causes substantial neurological impacts in mammals (27) and presumably other trophic levels shown to respond to DA. So if actual toxin levels generated during the blooms were closer to our lower estimated limits, persistent exposure may still impair the ecological health of the broader food web, including wild salmon and other fisheries dependent on HNLC environments.

Our findings show that oceanic *Pseudonitzschia* produce and accumulate DA and that they do not lose this capacity upon Fe enrichment. Although 12 major scientific field efforts have studied the role of iron in regulating ocean ecosystem and carbon cycling, they were not designed to evaluate potential negative impacts of oceanic iron fertilization as a carbon mitigation strategy. Although there remain uncertainties in extrapolating our results to large oceanic scales, the findings establish potential consequences for developing toxic phytoplankton blooms in pelagic ecosystems, which so far have not been adequately investigated. In particular, it will be important to assess the effect that these blooms may have on stressed fisheries that depend on HNLC regions. The current debate over the consequences of ocean iron fertilization as a carbon mitigation strategy must focus more closely on the impacts to organisms of the pelagic surface zone.

## Methods

**Water Collection.** Discrete seawater samples for in situ measures of phytoplankton biomass (as extracted chlorophyll *a*), phytoplankton species abundance, and DA concentrations were collected from 10-L Niskin bottles attached to an instrumented rosette. Seawater for use as the basal culture media of the deck-board, batch, and continuous cultures was collected from 5 m using a trace metal-clean, all-Teflon pumping system (19).

**Chlorophyll *a* Measurement and Cell Enumerations.** Chlorophyll *a* was determined using nonacidification (33) in vitro fluorometric analyses after parallel filtration onto Whatman GF/F filters (0.7-μm nominal pore size) and Poretics 20-μm pore size polycarbonate filters. Samples were extracted at sea in 90% acetone for approximately 24 h at -20 °C, and the fluorescence subsequently measured with a 10-AU fluorometer calibrated at the beginning of the cruise with pure chlorophyll *a* (Turner Designs). For most analyses,

fresh, nonpreserved samples were assayed for cell abundances. For the continuous culture experiments, formaldehyde-preserved samples were enumerated after the cruise. *Pseudonitzschia* cells were enumerated with a Palmer-Maloney counting chamber using a Zeiss Axiostar light microscope. Samples were settled for at least 24 h when required and counted at a total magnification of 200×.

**Domoic Acid Measurements.** Phytoplankton cells were collected from water samples collected using Niskin bottles, and culture media by filtering onto 47-mm-diameter nitrocellulose filters (HAWP04700 MF-Membrane filters; 0.45-μm pore size; Millipore). Filters were transferred to 15-mL Falcon tubes (BD Biosciences), and 4 mL of Nanopure water was added. Samples were vortexed, bath sonicated (model 5210; Branson) for 2 h to release cellular DA, and then vortexed again to mix before syringe filtration through a 25-mm-diameter (0.45-μm pore size) mixed cellulose ester filter (Millipore) to remove particulate matter. Filtrates were analyzed for DA content via an indirect cELISA.

An eight-point standard curve was prepared with DA standard (DACs-1C; National Research Council of Canada; standard range of 0.60–50,000 pg mL<sup>-1</sup>). Standards serially diluted in Nanopure water and samples were loaded in duplicate or triplicate into wells of a 96-well plate coated with BSA with carboxyl-linked DA antigen (18). Only Nanopure water was added to wells for determination of A<sub>max</sub>. Fifty microliters of anti-DA antibody [generated by injecting a rabbit with an imine-linked DA antigen (18)] was added to each well except blank wells, to which 50 μL of blocking buffer (3% nonfat dry milk in PBS solution) was added. The plate was sealed with plastic sealer, covered with foil, and gently mixed on an orbital shaker for 1 h at room temperature. The plate was washed four times with washing buffer (PBS-Tween), after which 100 μL of goat anti-rabbit antibody linked to horseradish peroxidase (Dako) was added to each well. After covering with plastic sealer and foil, the plates were gently rotated on an orbital mixer for 30 min at room temperature. The plate was washed four times with buffer, after which 100 μL of TMB (BioFX Laboratories) substrate was added to each well. After gentle rotation on an orbital mixer for 15 min at room temperature, 100 μL of stop solution (1N HCl) was added to each well, and after 2 min the absorbance was read at 450 nm on a Versamax plate reader (Molecular Devices). Sample concentrations were interpolated to the standard curve generated using Softmax Pro-4.8 software (Molecular Devices). The precision of reported DA concentrations, as estimated by the coefficient of variation (i.e., SD as a percentage of the mean) of duplicate analyses, averaged 33% and 12% for in situ and culture values, respectively. The DA values reported here have been corrected for the concentration factors determined for the assayed extract based on the filtered and extract sample volumes. The minimum analytical MDL of the cELISA (6.4 pg mL<sup>-1</sup>) was estimated as the product of the SD of six replicate analyses of a 5 pg mL<sup>-1</sup> DA standard (primary standard from National Research Council of Canada), with an SD of 1.896. For a df of 5, Student *t* value was 3.365, the MDL is 1.896 × 3.365 = 6.4 pg mL<sup>-1</sup>, and 19.0 pg mL<sup>-1</sup> is the limit of quantification.

***Pseudonitzschia* Isolation and Batch Culture Methods.** Representative phytoplankton fractions (1–2 mL) from a surface net tow (20-μm mesh size) were placed in *f/10*-enriched seawater medium [*f/2* (Sigma) diluted 5× with sterile-filtered ambient seawater collected at 5 m] and monitored microscopically for growth. Cell isolates [*P. turgidula*; identified after the cruise via scanning and transmission EM (34)] from the net tow were pooled to establish a culture. This was maintained for 5 to 6 d in *f/10* medium. Before enrichment experiments, *f/10* culture media was replaced with ambient seawater by concentrating cells in a 20-μm mesh concentration chamber and allowing cells to grow for 3 to 4 d. Batch culture experiments were initiated in deck-board Plexiglas incubators that were maintained at ambient seawater temperature and 50% surface irradiance via constant flow with surface seawater and by covering incubators with neutral density plastic film (Cinemills). Cells were enumerated daily using Palmer-Maloney counting chambers. Cultures reached stationary growth on d 6, at which time 140 to 200 mL was filtered and extracted in 2 mL Nanopure water before particulate DA analysis.

**Deck-Board Incubations and Culture Studies.** An all-Teflon sampling system was used to collect surface waters for incubation from the 50% light depth at OSP. The system comprised a plastic towfish with Kevlar-encased Teflon perfluoroalkoxyalkane (PFA) tubing, positioned 8 m outboard of the vessel using the ship's crane to avoid contamination associated with the vessel, and a double diaphragm pump that provided a continuous clean water supply to a shipboard-fabricated, positive-pressure clean room, where it was dispensed under a Class 100 HEPA bench. To reduce variance from potential small-scale surface patchiness during sample collection, a 50-L acid-cleaned

polyethylene carboy was rinsed thoroughly and filled with whole water for dispersal into the batch and continuous culture vessels.

**Continuous Culture Apparatus and Operation.** The Ecostat (35, 36) was comprised of fourteen 2-L custom-designed cylindrical polycarbonate culture vessels, fitted with inflow and outflow ports, supported in cradles that mechanically oscillated to keep the cultures well mixed. The bottles, cradles, and rocking assembly were contained within a transparent, enclosed Plexiglas recirculating water bath (approximately 240 L) maintained at ambient sea surface temperature. The culture media sources were filtered (<0.2  $\mu\text{m}$ ) ambient seawater with or without iron amendments stored in 20 L polycarbonate reservoir carboys in the dark and pumped continuously by a multichannel peristaltic pump with high-purity platinum-cured silicone pump tubing (1/16-inch outer diameter; Ismatec) to duplicate culture vessels per treatment through 15 m long 1/16" OD Teflon PFA high-purity tubing (Upchurch). Outflow from each culture vessel was carried by 1/8-inch outer diameter Teflon PFA HP tubing to a central point outside of the water bath

where flow rates were measured and samples collected daily. Continuous flow to the culture vessels was set to a dilution rate of 0.5  $\text{d}^{-1}$ , a rate that represents approximately 50% of the ambient in situ growth rate based on previous visits to the area.

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