

# The quantifying, mapping, and risk analysis of human-related stressors in the high seas

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## Abstract

**Objectives:** Marine biodiversity and ecosystem services in the high seas are threatened by numerous stress factors caused by human activities, including global shipping, high-sea fishing, marine plastic pollution, and anthropogenic climate change. Socioeconomic factors are one of the criteria for the establishment of area-based management tools in the high seas for marine biodiversity conservation beyond national jurisdiction. The aim of the work is to propose a spatiotemporal approach to identify risks from marine human activities and recommendations for high seas governance. **Methods:** Data related to human activities from 2014 to 2022 were used to calculate the distribution and changes of human-related stressors, and the risk to marine biodiversity in the high seas caused by human activities. **Results:** The North Atlantic, Philippine Sea, Arabian Sea, Bay of Bengal, and East Central Atlantic show high and increasing intensities of human-related stressors, and are therefore particularly at need for the protection and conservation of marine biodiversity. Risks from human activities vary within the marine areas that are prioritized for biodiversity protection. The study recommends that the designation of high seas protected areas should take into account the types of risks to which the different marine areas are exposed, and that the high seas protected areas should be established gradually. At the same time, appropriate management measures should be formulated according to the intensity of human activities in the different marine areas. **Conclusions:** Quantifying and classifying the risk from human-related stressors could help identify solution for the protection and conservation and facilitate the marine spatial planning, establishment area based management tools, including marine protected areas in the high seas.

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## Keywords

Risk assessment, marine conservation, marine plastic pollution, climate change, marine protection

## Introduction

The high seas are an area beyond national jurisdictions, which covers almost two-thirds of the global ocean. The high seas play essential roles such as sequestering carbon dioxide and provide important habitats for marine ecosystems,<sup>1</sup> and oceanic highways for marine species migration.<sup>2,3</sup> Since the 1950s, human activities have expanded to the high seas, owing to technological innovations and an increasing demand for deep-sea resources.<sup>4,5</sup> More than 80% of the world's trade flow is carried across the high seas,<sup>6</sup> accounting for trillions of dollars in trade.<sup>7</sup> The economic value of high seas fisheries is approximately \$7.6 billion per year.<sup>8</sup> However, the expanding of long-line and deep-sea fisheries,<sup>9</sup> international shipping,<sup>10</sup> marine plastic pollution,<sup>11,12</sup> and ocean acidification caused by anthropogenic climate change,<sup>13</sup> have led to threats to marine biodiversity.<sup>14</sup>

The conservation and management of high seas is a collective responsibility of all countries.<sup>15</sup> Since 2012, recognizing the need for high-sea biodiversity conservation,<sup>14</sup> the United Nations adopted an agreement for the “conservation and sustainable use of marine biodiversity of the areas beyond national jurisdiction” (hereinafter the BBNJ Agreement) in 2023.<sup>16</sup> The development of area-based management tools (ABMTs), including marine protected areas (MPAs) to protect and conserve at least 30% of the oceans by 2030 become a historic agreement to which nations committed in the “Kunming-Montreal Global Biodiversity Framework” of the Convention on Biological Diversity in 2022.<sup>17</sup> Economic and social factors are one of the indicative criteria for the identification of ABMTs,<sup>16</sup> and the marine systematic conservation planning.<sup>18,19</sup> The effective delineation of ABMTs in the high seas requires quantification of the distribution and risks of human activities.

Studies quantifying and mapping threats from human activities at a global scale have focused on different areas, including lands,<sup>20–25</sup> coastal areas,<sup>26</sup> and the oceans,<sup>13,27–29</sup> using different scales of measurement and various approaches for normalization. In studies that focused on marine and coastal areas, Halpern et al. calculated and mapped human activities impacts on the oceans based on fisheries, climate change, and land-based activities in 2008 and 2015.<sup>28,30</sup> O'Hara et al. intersected the spatial distribution of 21,267 marine species with 13 anthropogenic stressors to evaluate the impact risk of human activities in 2023.<sup>29</sup> Allan et al. evaluated the footprint of human activities in coastal areas based on 10 major terrestrial stressors and 10 marine stressors and analyzed the factors affecting human activity at the national scale in 2023.<sup>26</sup> Elliott et al. quantified human footprints based on multisectoral and trans-boundary impacts using the Drivers Activities Pressures State Impacts Responses model in 2020.<sup>27</sup> Other studies on the impact of human activities have focused on sea areas including the Mediterranean Sea,<sup>31</sup> Canadian<sup>32</sup> and Icelandic<sup>33</sup> coastal waters, and the North Sea.<sup>34</sup>

However, no studies have tracked the trends in human-related stressors in the high seas or identified the risks concerning the intensity and trends of human-related stressors. International shipping and fishing are important high-sea human activities. Climate

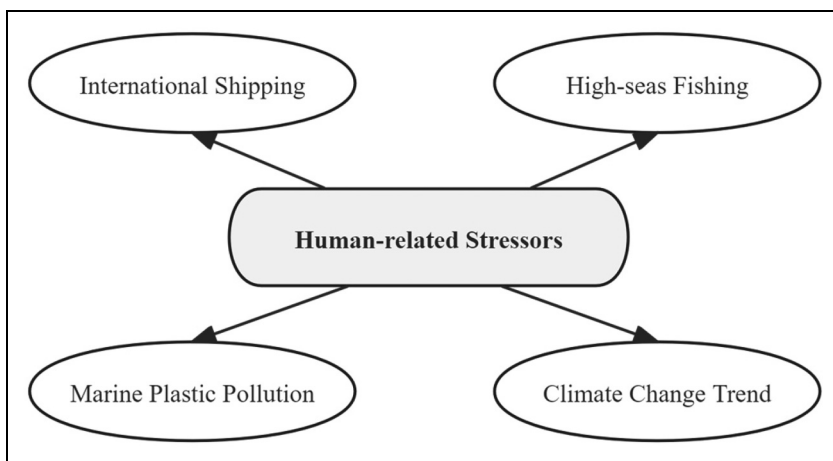
change<sup>35</sup> and marine plastic pollution<sup>36,37</sup> are also stressors for high-seas biodiversity. The entanglement and ingestion of plastic harms marine species<sup>11,38,39</sup>; the amount of marine plastic floating is increasing annually.<sup>40,41</sup> Anthropogenic climate change also poses a major threat,<sup>42</sup> resulting in declines of marine biomass and fishery production.<sup>43</sup> To address gaps in knowledge, we quantified and mapped the distribution and trends of human-related stressors in the high seas based on four factors: shipping, fishing, marine debris, and climate change trends. The results provide a reference for future governance of the high seas.

## Methods

### *Data sources and interpretation of the indicators*

*International shipping and high-sea fishing.* In this study, we used the automatic identification system (AIS) data to quantify the stress caused by international shipping and fishing in the high seas. Originally, AIS data were used to enhance shipping collision avoidance and improve shipping radar functions, traffic management and report. Moreover, they also provide additional information, and improve communication between ships.<sup>44,45</sup> As required by the International Maritime Organization, cargo ships of over 300 tonnes are required to be equipped with an AIS system, providing the basis for shipping activity analysis using AIS data.<sup>46</sup> Currently, because of its strong availability and extensive information coverage, AIS data are widely used as a data source for research on ship activities,<sup>46,47</sup> including the activities of high-sea fishing vessels,<sup>48</sup> and for mapping fishing grounds<sup>49–51</sup> (Figure 1).

AIS data of ships from 2014 to 2022 were used to map the distribution of international trade transport and high-sea shipping. For this study, three categories of vessels, namely, bulk carriers, container ships, and oil tankers, which play the main role in international maritime trade,<sup>6</sup> were selected to represent the activities of global maritime trade. The



**Figure 1.** Factors of human-related stressors in the high seas.

AIS data were obtained from the public service website Shipxy (<http://www.shipxy.com>), which contains both shore-based and space-based data and provides a comprehensive picture of vessel activity in the high seas.<sup>52</sup> The data includes ship type, vessel speed, position, course, Maritime Mobile Service Identity, time, number of vessels, and shipping deadweight. For the distribution of high-sea fisheries, vessel speed is generally used to distinguish between fishing and non-fishing behaviors.<sup>49,53</sup> Global Fishing Watch has also begun to use AIS data to analyze fishing activities.<sup>54</sup> Following the methodology used in Global Fishing Watch by Alzorritz et al.,<sup>55</sup> this study used four knots as a speed criterion to distinguish between the fishing and cruising hours of vessels. We screened AIS data for fishing vessels below four knots as a basis for calculating. AIS data were cleaned and processed according to the method described by Fiorini et al.<sup>56</sup>

**Marine plastic pollution.** Plastic pollution in the high seas is mainly accumulates in garbage patches within ocean gyres,<sup>57</sup> and is carried and accumulated by ocean currents.<sup>58</sup> The work use the estimation by Kaandorp et al. that the amount of plastic waste marine environments is increasing at an annual a rate of 4%.<sup>36</sup> Plastic waste in the high seas is influenced by ocean circulation, and waste discharged into the ocean is concentrated in several areas forming garbage patches that are relatively fixed.<sup>59</sup> Considering that the annual increase in plastic discharge into the sea is approximately 4%, in this work, we estimated that the distribution of global marine plastic pollution increased equally from 2016 to 2022 at a growth rate of 4% per year,<sup>36</sup> and used the density grids of marine plastic distribution in 2014 from the study of Clark et al.<sup>11</sup> as a baseline.

**Climate change.** Sea surface temperature (SST) is an essential climatic variable that describes the influence on climate dynamics.<sup>54</sup> Fluctuations in SST are an important factor influencing the abundance of marine species in the high seas.<sup>60,61</sup> In this study, we used annual SST data from the European Space Agency Climate Change Initiative<sup>62</sup> to calculate the SST trends and map the extent of ocean stressors caused by climate change,<sup>63</sup> as a faster rate of climate change has been shown to have more negative impacts on marine biodiversity and ecosystem services.<sup>64</sup> SST data from 2008 to 2022 were used to calculate SST trends every 7 years<sup>62,65</sup> (e.g. the trend in SST for 2014 was calculated from the annual SST from 2008 to 2014).

## Data analysis

**Intensity analysis of human activities and climate change.** First, all vector datasets were projected before converting the data into raster format. The study used the “Extract in Mask” function to convert all data into a uniform range layer (resolution  $0.2^\circ \times 0.2^\circ$ ) with high seas layers as a base to facilitate further processing and calculation of the data. High-sea layers were downloaded from “Marine Regions.”<sup>66</sup> The layers of shipping and fishing activities were created by using the “Density Function” in ArcGIS (v10.5)<sup>67</sup> with a resolution of  $0.2^\circ \times 0.2^\circ$  based on a nearest neighbor routine. We normalized each of the human activity threat data layers concerning the maximum score of each stressor such their maximum value was one (after this normalization, stressors that existed everywhere did not have a minimum value of zero). The grid layers of shipping, fishing, SST trends,

and marine plastic distribution were projected using the Mollweide projection with WGS84 datum. For each map, we summed the intensity of the different factors for each year in all cells and normalized the results of mapping to facilitate comparison of human-related stressors.

**Spatiotemporal trend analysis.** The Mann–Kendall (MK) monotonic trend analysis method was used in this study to quantitatively calculate time-series changes in SST and intensity of human activities in the high seas from 2014 to 2022. The MK test is widely used to access trends in time series data.<sup>68–70</sup> It has the advantage of handling missing values and partial outliers.<sup>71–73</sup> The MK test (S) showed a linear or nonlinear trend in the data series, which was calculated based on the test statistic S as follows. S indicates the sum of the differences of the values  $x$ , and  $sgn(x)$  is a signum function that indicate the value  $-1$ ,  $+1$  and  $0$ , according to the positive and negative of  $x$ .

$$S = \sum_{i=1}^{n-1} \sum_{j=i+1}^n sgn(x(i) - x(j))$$

while

$$\begin{cases} \text{if } x > 0, & sgn(x) = 1 \\ & sgn(0) = 0 \\ \text{if } x < 0, & sgn(x) = -1 \end{cases}$$

The positive or negative values of S correspondingly reflects to upward or downward trends, respectively.<sup>73–75</sup> To obtain a more accurate result, the variance of S  $VAR(S)$  was calculated as

$$VAR(S) = \frac{n(n-1)(2n+5)}{18}$$

where  $n$  is the number of data points. Then the Kendall standardized test statistic  $Z$  was then calculated by the following equation:

$$Z = \begin{cases} \frac{S-1}{\sqrt{VAR(S)}}, & S > 0 \\ 0, & S = 0 \\ \frac{S+1}{\sqrt{VAR(S)}}, & S < 0 \end{cases}$$

where  $Z > 0$  and  $Z < 0$  refers to increasing and decreasing trends, respectively.<sup>70</sup> The results of this analysis indicated the status of human activity risks in the context of increasing or decreasing trends. All data handling was carried out using R “trend” package.<sup>76</sup> We used 2022 as the base year for the trend calculations in this study.

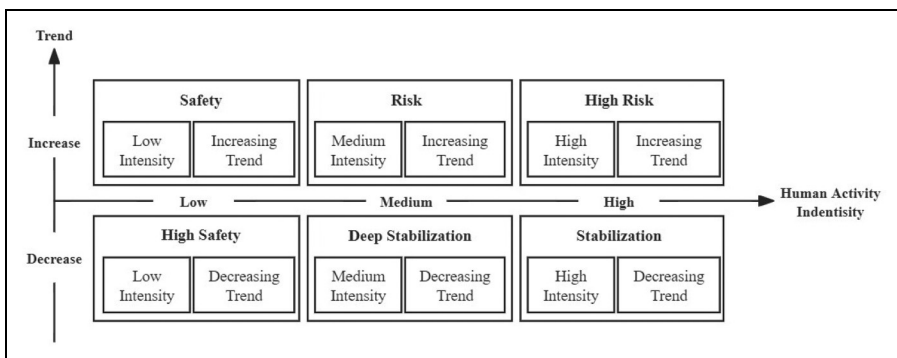
**Risk analysis.** The intensity of human activities and climate change has created a risk to the protection and conservation of marine biodiversity, and temporal changes in human activities and climate change stressors may increase or decrease the risk to marine

species.<sup>28</sup> In line with the intensity and variability of the risk of human activities in the high seas, we proposed a spatiotemporal approach to characterize the risk of human activities.

This study categorized the risk of marine anthropogenic activities with reference to the classification of global PM<sub>2.5</sub> pollution as proposed by Lim et al.<sup>77</sup> The six classifications are “High Risk,” “Risk,” “Stabilization,” “Deep Stabilization,” “Safety,” and “High Safety.” Firstly, the intensity of human activity was divided evenly into three categories, namely high, medium, and low. Trends in the intensity of human activity were classified as either increasing or decreasing. Using data from 2022 as a baseline, the two categorized layers were overlaid to obtain six categories of risk-rated areas.

High Risk regions are marine areas with a combination of high intensity and increasing trends in human activity and climate change, whereas, Risk regions refer to sea areas with medium intensity and increasing trends in human activity. The criteria used to classify of the other four types of marine areas were similar (Figure 2). Marine areas in Stabilization refer to regions with a combination of medium intensity and a decreasing trend in human activities, whereas, Deep Stabilization refers to regions with high intensity and a decreasing trend in human activities. Safety regions refer to marine areas with low intensity but increasing trends in human activities and climate change, whereas, High Safety regions are sea areas with both low intensity and decreasing trends in human activities and climate change. We used the R packages “trend” and “raster” for calculations and ArcGIS (v10.5) to visualize the results.

**Data visualization.** After classifying the risks in the high seas, spatial analyses were used to disaggregate the statistics for the data at different risk levels. The distribution data of Important Marine Mammal Areas (IMMAs)<sup>78</sup> and Important Bird and Biodiversity Areas (IBAs)<sup>79</sup> were converted to a raster format and overlaid with the risk classification result. All results were visualized using ArcGIS (v10.5). The shapefiles of land boundaries<sup>80</sup> and Exclusive Economic Zone<sup>81</sup> were from the open source website.



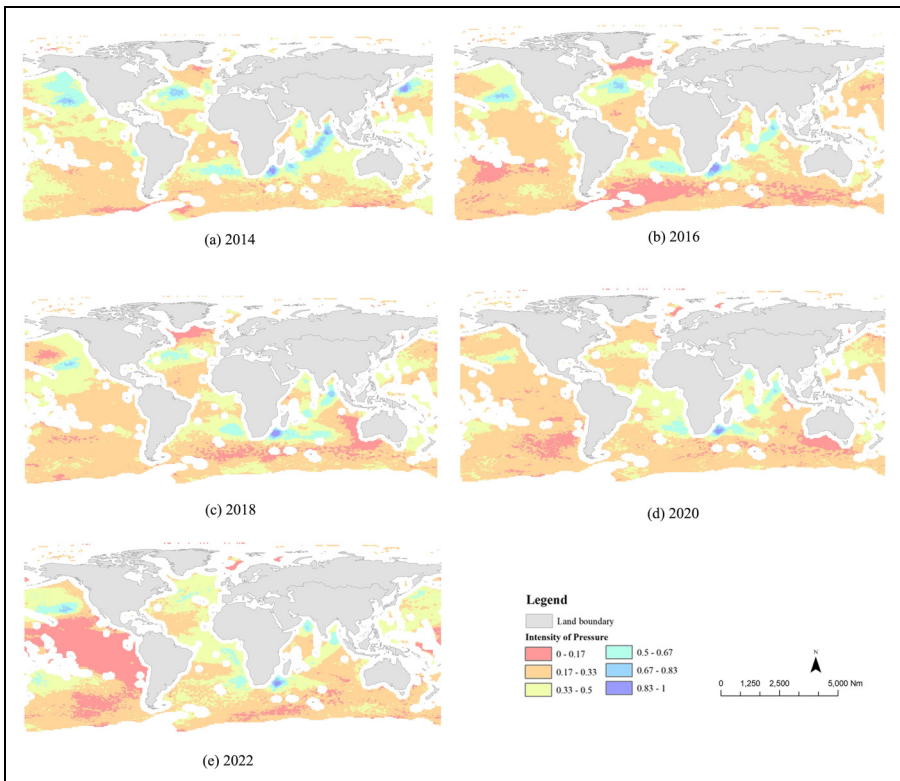
**Figure 2.** Classifications of the human-related stressors risks on the high seas from 2014 to 2022.

## Results

### *Changing intensities of human activities on the high sea*

Regions with high-intensity of human activities vary from year to year based on intensity assessments. The global distribution of high human-related pressures over the last decade indicates a high risks in the high seas. Sea areas with higher intensities of human-related pressures are generally spatially distributed along the main shipping routes of the Indian Ocean, Cape of Good Hope in southern Africa, Northeast Pacific, and Northwest Atlantic, with intensities above 0.83. Regions such as the Southeast Pacific and Southeast Indian Ocean have relatively low intensities of human-related environmental pressures (less than 0.17 from 2014 to 2022) (Figure 3).

A monotonic MK trend analysis of human-related stressors intensity from 2014 to 2022 was used to identify areas with increasing or decreasing concentrations. There was a clear increase in the western and central Pacific, Philippine Sea, Bay of Bengal, and Arabian Sea, as well as in the Northeast and mid-Atlantic regions of the high seas, with MK values >0.5 (more than 50%, extreme increase). Human activities and climate change in the high-sea regions, such as the eastern Pacific and Northwest



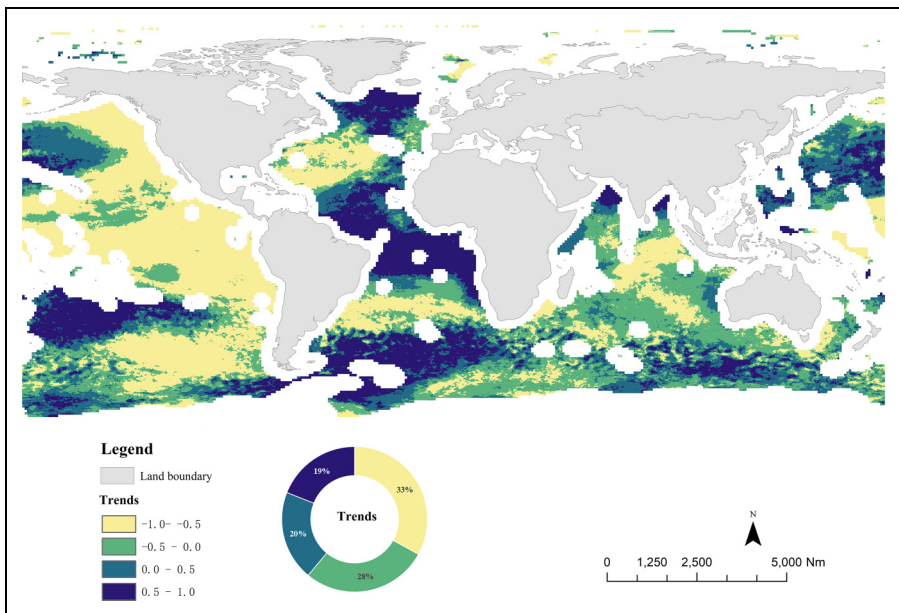
**Figure 3.** Distribution of the normalized intensity of human activities and climate change in the high seas from 2014 to 2022 (per 2-year period).

Atlantic, generally show a decreasing trend, with MK values  $< -0.5$  (more than 50%, extreme decrease). Overall, 39% of the total high sea area showed an upward trend, while 61% showed a downward trend (Figure 4).

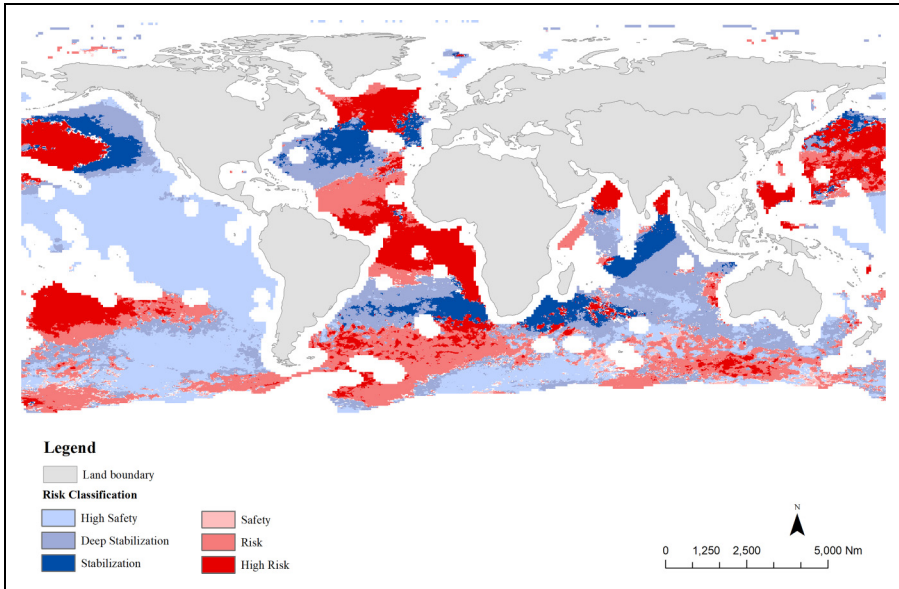
### *Spatiotemporal classification and risk assessment of human-related pressures on the high seas*

According to the calculation, the high-risk areas which represent areas of high and increasing human-related pressures are mainly located in the North Pacific, the Philippine Sea and the central and southern of the Pacific Ocean, the Arabian Sea and the Bay of Bengal of the Indian Ocean, as well as the North Atlantic and South-East Atlantic Oceans. The Indian Ocean shipping lane area and the Northwest Atlantic are in Stabilization Status, indicating areas of high and decreasing human-related pressures. The south Atlantic and the north-central Atlantic are at Risk status, with medium and increasing human-related pressures, and the other areas are in a state of Deep Stabilization and High Safety Status. The marine areas within Deep Stabilization were widespread in the Indian Ocean, indicating that stable changes are underway (Figure 5).

The risk assessment results varied from one sea area to another. Overall, the largest areas were in the High Safety category, with roughly equal proportions of areas in the High Risk, Risk, and Deep Stabilization categories, and the smallest areas in the Safety category. The largest proportion of High Risk and Risk areas is in the Atlantic



**Figure 4.** Mann-Kendal trends of normalized intensity of human-related stressors on the high seas from 2014 to 2022.



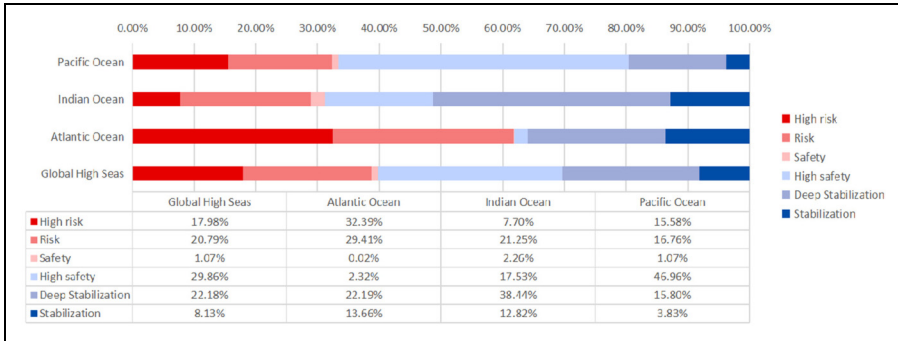
**Figure 5.** Risk classification distributions of human-related stressors on the high seas in 2022.

Ocean, while the Indian and Pacific Oceans have larger proportion of High Safety, Deep Stabilization, and Stabilization statuses, with the Indian Ocean having the largest proportion of areas in Deep Stabilization. Overall, the Atlantic region shows the highest risk to marine biodiversity from human activities, followed by the Indian Ocean, the Pacific had the lowest relative risk to marine biodiversity from human activities (Figure 6). These spatiotemporal results will facilitate the identification of marine areas at High Risk status for marine biodiversity conservation, taking into account the increasing or decreasing trends, as well as the intensity of human activities and climate change.

### *Drivers influencing global risk classifications on the high seas*

Trends in the exposure to various human-related stressors vary from one marine area to another. The factors causing this variation include the development of international regulations and influence of the global economy.

The distribution of global high-sea trade routes is essentially relatively stable in space and time, and the high intensity of international shipping activities is concentrated in shipping route areas. Owing to the impact of COVID-19 postpandemic trends from 2019, the war in Ukraine, and tightening global supply chains over the past 2 years, global seaborne trade volume has decreased marginally by 0.4% in 2022. The overall trend of shipping activity in key route areas, such as the northern Indian Ocean has declined to some extent. Consequently, the routes in these regions are in a state of high stability risk with high intensity and an overall downward trend. This may change in the future as the global economy recovers and develops, particularly, with the development of foreign trade in China, an important global shipping country.<sup>6</sup>



**Figure 6.** Proportions of different risk categories in the Pacific, Indian, and Atlantic oceans.

As far as high-sea fisheries are concerned, the changes in their activity vary from region to region. The intensity of high-sea fishery activities in the western and central Pacific and the eastern and central Atlantic has increased annually. In particular, catches of tribe Thunnini and skipjack tuna (*Katsuwonus pelamis*), has increased significantly in the western and central Pacific. China is the world’s largest marine fishery, but catches have declined annually from 2015 to 2022, with the policy of reducing fishing to continue in the 14th Five-Year Plan (2021–2025),<sup>82</sup> the intensity of fishing activities near China is expected to continue to decline in the coming years. This is related to the risk categorized in sea areas such as the western and central Pacific and the eastern central Atlantic.

Marine plastic debris pollution is an important human-related issue that threatens the safety of marine mammals and other marine species, particularly through the risk of entanglement and ingestion.<sup>12</sup> The North Pacific, North Atlantic, South Pacific, South Atlantic, and Southern Indian oceans are five oceanic garbage patches,<sup>59</sup> over time, the amount of plastic accumulating in these waters has increased, posing a risk to marine mammals, marine reptiles, and other marine species. Recently, with the development of the International Legally Binding Instrument to end plastic pollution, global regulations have been proposed to address marine plastic pollution.<sup>83</sup> If the rate of increase in plastic litter in the oceans can be slowed, and at the same time, salvage and other methods implemented in areas where plastic litter has accumulated (e.g. the “Ocean Clean System”), it may be possible to reduce the marine risk caused by plastic pollution in the future.

The threat posed by climate change to the oceans is global. The threat posed by climate change has increased, particularly in the northern and southern mid-latitude waters of the Atlantic Ocean. Therefore, there is a need to reduce the impacts of climate change, including through carbon reduction on ships. In addition, climate change mitigation cannot be focused solely on human activities in the marine sector, this requires global coordination across sectors and industries.

## Discussion

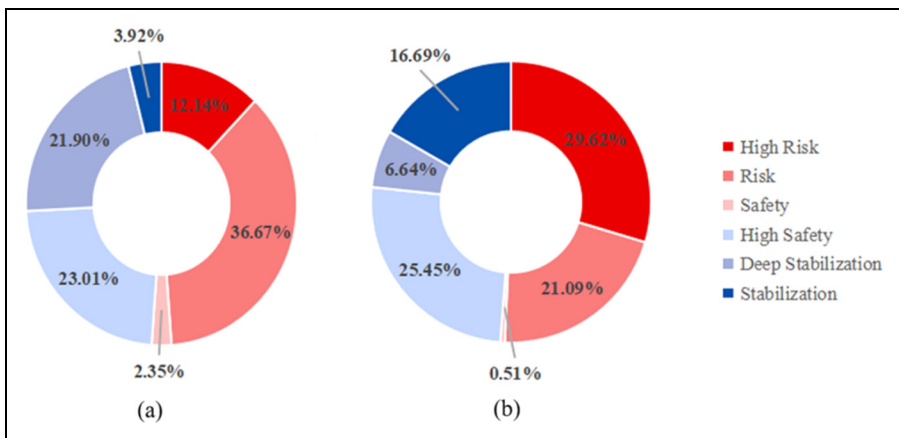
### *Protection of marine species based on risk classification results*

Mapping and quantifying the stressors of human activity and climate change are fundamental to understanding the risks of marine protection and conservation in the high seas.

This study considered the application of human activity stressors in the design and management of MPAs. To identify the hotspots that need to be protected by marine mammal and seabird specifications, IMMAs have been described by the International Union for Conservation of Nature, and the marine IBAs have been identified by BirdLife International. IMMAs and marine IBAs beyond national jurisdictions can be considered priority areas for marine conservation in the high seas. Descriptions of IMMAs and marine IBAs are based on the distribution data of marine mammals and seabird species, and they are generally selected according to criteria and thresholds such as species or population vulnerability, species distribution and abundance, activities of key life cycle stages and special attributes, and expertise from expert seminars,<sup>84,85</sup> without considering human activities such as global shipping, high-sea fishing, and marine plastic pollution.

Comparative analysis revealed a significant difference in the potential social and economic risks of human activities between the IMMA and marine IBA regions. The proportions of high-risk areas in IMMAs and marine IBAs were 12.14% and 29.62%, respectively, with 36.67% and 21.09% at the Risk level and, 21.9% and 6.64% at the Deep Stabilization status, respectively; over 20% of areas were at the High safety level. As shown in Figure 7, there are significant differences in the impacts of risks from human-related pressures in marine areas that also require protection.

At present, the establishment and management of MPAs are associated with high monitoring and administrative costs,<sup>86</sup> as such, more cost-efficient<sup>87</sup> and effective MPA management will be needed in the future.<sup>88</sup> Blind and overly large-scale MPAs with high ambition are difficult to monitor effectively,<sup>89</sup> leading to the emergence of “paper parks.”<sup>90</sup> It is recommended that different types of areas be given priority in the designation of future protected areas in the high seas. For areas with a high risk to marine biodiversity in the high seas, MPAs should be designated as early as possible, and resources should be concentrated to monitor and implement stringent management measures. Areas that are in a Safe condition and where the risk of human activities is relatively low can be exempted from designation as priority protection areas to



**Figure 7.** Proportions of existing important marine mammal areas (a) and marine important bird and biodiversity areas (b) in different risk categories.

reduce the administrative costs of MPAs and improve the actual protective effect in high-sea marine ecosystems. The same applies to marine spatial planning (MSP) in the high seas, which is performed to determine the balance between human activities and environmental protection.<sup>91</sup> Economic and social aspects must be considered in MSP.<sup>92</sup> Consequently, studies on human-related stressors in the high seas are also useful for the development of MSP.

### **Limitations**

This study has some limitations and uncertainties that should be addressed in future studies. The data used in this study were all from open sources. Although the latest available data were used, data such as the spatial distribution of plastics are relatively outdated. While this study estimates trends in marine plastic pollution, it is understood that this may not fully reflect the true spatial distribution of marine plastic litter. In addition, although the AIS data have recently been widely used for research, they were not originally developed for research use, and there are gaps and uncertainties in some cases.<sup>45,93</sup> For example, some small vessels that do not carry an AIS system,<sup>94</sup> some AIS coverage is lacking,<sup>95</sup> and some vessels lag in AIS reception.<sup>93</sup> The known and common gaps in AIS data may lead to an underestimation of overall activity.<sup>95</sup>

### **Conclusions**

Marine socioeconomic factors are among the criteria that determine the designation of MPAs and the development of MSP in the high seas. In this study, human-related pressures on high-sea ecosystems from shipping, high-sea fishing, marine plastic pollution, and climate change, as well as the intensity and trends of these pressures, were used to delineate risk categories in different high-sea areas. The intensity of activity in the North Pacific, Arabian Sea, and Southern Central Atlantic are high and show an increasing trend. We propose using the intensity of human activities and climate change in different areas as a reference for prioritizing the establishment of future high-sea MPAs. This will make the designation of MPAs more effective and lead to the development of more targeted management measures. This study proposes an analytical framework for assessing the risk of human activities in the high seas that incorporates the intensity and trends of human activities.

### **Author contributions**

Chang Zhao put forward the overall research ideas and framework, responsible for data analysis, the drafting and revision of the paper; Miaozhuang Zheng is responsible for data curation and Visualization; Yuejing Ge is responsible for the revision of the paper.

### **Consent**

There are no human participants in this article and informed consent is not required.

### **Declaration of conflicting interests**

The authors declared no potential conflicts of interest with respect to the research, authorship, and/or publication of this article.


## Ethical considerations

Not applicable.

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## References

1. Banet C, ed. Deep-sea ecosystems: biodiversity and anthropogenic impacts. In: Brill | Nijhoff (ed) *The law of the seabed*, 2020, pp.36–60. DOI: 10.1163/9789004391567\_004.
2. Harrison AL, Costa DP, Winship AJ, et al. The political biogeography of migratory marine predators. *Nat Ecol Evol* 2018; 2: 1571–1578.
3. Jarvis RM and Young T. Pressing questions for science, policy, and governance in the high seas. *Environ Sci Policy* 2023; 139: 177–184.
4. Dunn DC, Jablonicky C, Crespo GO, et al. Empowering high seas governance with satellite vessel tracking data. *Fish Fish* 2018; 19: 729–739.
5. Fourchault L, Dahdouh-Guebas F, Dunn DC, et al. Generating affordable protection of high seas biodiversity through cross-sectoral spatial planning. *One Earth* 2024; 7: 253–264.
6. IMO. Review of maritime transport 2023, [https://unctad.org/system/files/official-document/rmt2023\\_en.pdf](https://unctad.org/system/files/official-document/rmt2023_en.pdf) (2023).
7. World Bank Group. Blue economy, <https://www.worldbank.org/en/topic/oceans-fisheries-and-coastal-economies> (2024).
8. Sala E, Mayorga J, Costello C, et al. The economics of fishing the high seas. *Sci Adv* 2018; 4: eaat2504.
9. Merrie A, Dunn DC, Metian M, et al. An ocean of surprises – trends in human use, unexpected dynamics and governance challenges in areas beyond national jurisdiction. *Glob Environ Change* 2014; 27: 19–31.
10. Womersley FC, Humphries NE, Queiroz N, et al. Global collision-risk hotspots of marine traffic and the world’s largest fish, the whale shark. *Proc Natl Acad Sci U S A* 2022; 119: e2117440119.
11. Clark BL, Carneiro APB, Pearmain EJ, et al. Global assessment of marine plastic exposure risk for oceanic birds. *Nat Commun* 2023; 14: 3665.
12. Zantis LJ, Carroll EL, Nelms SE, et al. Marine mammals and microplastics: a systematic review and call for standardisation. *Environ Pollut* 2021; 269: 116142.
13. Sala E, Mayorga J, Bradley D, et al. Protecting the global ocean for biodiversity, food and climate. *Nature* 2021; 592: 397–402.
14. Ban NC, Bax NJ, Gjerd KM, et al. Systematic conservation planning: a better recipe for managing the high seas for biodiversity conservation and sustainable use. *Conserv Lett* 2014; 7: 41–54.
15. Gullett W. The contribution of the law of the sea convention to marine environmental protection: 40 years of shimmering seas or high hopes dashed against juridical rocks? *Mar Policy* 2023; 49: 105498.
16. United Nations. Intergovernmental conference on an international legally binding instrument under the United Nations Convention on the Law of the Sea on the conservation and

- sustainable use of marine biological diversity of areas beyond national jurisdiction further resumed fifth session, <https://documents.un.org/doc/undoc/ld/n23/177/28/pdf/n2317728.pdf?token=LdluQ1GjhpZMd4LJgr&fe=true> (2023).
17. CBD. Kunming-Montreal global biodiversity framework, p. 15/4, <https://www.cbd.int/doc/decisions/cop-15/cop-15-dec-04-en.pdf> (2022).
  18. Chowdhury S, Zalucki MP, Hanson JO, et al. Three-quarters of insect species are insufficiently represented by protected areas. *One Earth* 2023; 6: 139–146.
  19. Holness SD, Harris LR, Chalmers R, et al. Using systematic conservation planning to align priority areas for biodiversity and nature-based activities in marine spatial planning: a real-world application in contested marine space. *Biol Conserv* 2022; 271: 109574.
  20. Correa Ayram CA, Etter A, Díaz-Timoté J, et al. Spatiotemporal evaluation of the human footprint in Colombia: four decades of anthropic impact in highly biodiverse ecosystems. *Ecol Indic* 2020; 117: 106630.
  21. Riggio J, Baillie JEM, Brumby S, et al. Global human influence maps reveal clear opportunities in conserving earth’s remaining intact terrestrial ecosystems. *Glob Change Biol* 2020; 26: 4344–4356.
  22. Sanderson EW, Jaiteh M, Levy MA, et al. The human footprint and the last of the wild. *BioScience* 2002; 52: 891.
  23. Venter O, Sanderson EW, Magrath A, et al. Sixteen years of change in the global terrestrial human footprint and implications for biodiversity conservation. *Nat Commun* 2016; 7: 12558.
  24. Watson JEM and Venter O. Mapping the continuum of humanity’s footprint on land. *One Earth* 2019; 1: 175–180.
  25. Williams BA, Venter O, Allan JR, et al. Change in terrestrial human footprint drives continued loss of intact ecosystems. *One Earth* 2020; 3: 371–382.
  26. Allan H, Levin N and Kark S. Quantifying and mapping the human footprint across earth’s coastal areas. *Ocean Coast Manag* 2023; 236: 106476.
  27. Elliott M, Borja A and Cormier R. Activity-footprints, pressures-footprints and effects-footprints – walking the pathway to determining and managing human impacts in the sea. *Mar Pollut Bull* 2020; 155: 111201.
  28. Halpern BS, Frazier M, Potapenko J, et al. Spatial and temporal changes in cumulative human impacts on the world’s ocean. *Nat Commun* 2015; 6: 7615.
  29. O’Hara C, Frazier M, Valle M, et al. Cumulative human impacts on global marine fauna highlight risk to fragile functional diversity of marine ecosystems. *Preprints* 2023. DOI: 10.22541/au.168328097.74947503/v1.
  30. Halpern BS, Walbridge S, Selkoe KA, et al. A global map of human impact on marine ecosystems. *Science* 2008; 319: 948–952.
  31. Micheli F, Halpern BS, Walbridge S, et al. Cumulative human impacts on Mediterranean and Black Sea marine ecosystems: assessing current pressures and opportunities. *PLoS One* 2013; 8: e79889.
  32. Ban N and Alder J. How wild is the ocean? Assessing the intensity of anthropogenic marine activities in British Columbia, Canada. *Aquat Conserv Mar Freshw Ecosyst* 2008; 18: 55–85.
  33. Jóhannesson SE, Heinonen JT and Davíðsdóttir B. Increasing the accuracy of marine footprint calculations. *Ecol Indic* 2019; 99: 153–158.
  34. Stelzenmüller V, Rehren J, Örey S, et al. Framing future trajectories of human activities in the German North Sea to inform cumulative effects assessments and marine spatial planning. *J Environ Manage* 2024; 349: 119507.
  35. Dias MP, Martin R, Pearmain EJ, et al. Threats to seabirds: a global assessment. *Biol Conserv* 2019; 237: 525–537.
  36. Kaandorp MLA, Lobelle D, Kehl C, et al. Global mass of buoyant marine plastics dominated by large long-lived debris. *Nat Geosci* 2023; 16: 689–694.

37. Kühn S and Van Franeker JA. Quantitative overview of marine debris ingested by marine megafauna. *Mar Pollut Bull* 2020; 151: 110858.
38. Provencher JF, Bond AL, Avery-Gomm S, et al. Quantifying ingested debris in marine megafauna: a review and recommendations for standardization. *Anal Methods* 2017; 9: 1454–1469.
39. Ryan PG, Moore CJ, van Franeker JA, et al. Monitoring the abundance of plastic debris in the marine environment. *Philos Trans R Soc, B* 2009; 364: 1999–2012.
40. Aliani S, Basurko OC, Arias M, et al. Editorial: marine litter windrows. *Front Mar Sci* 2022; 8: 827907.
41. Amador MFG and Sansón LZ. Surface dispersion of coastal discharges in North America towards the Great Pacific Garbage Patch. *Mar Pollut Bull* 2023; 191: 114961.
42. Bass A, Wernberg T, Thomsen M, et al. Another decade of marine climate change experiments: trends, progress and knowledge gaps. *Front Mar Sci* 2021; 8: 714462.
43. Yin J, Zhang C, Xue Y, et al. Potential impacts of ocean warming on the trophic control of a threatened marine ecosystem. *Prog Oceanogr* 2023; 211: 102976.
44. Weng J, Shi K, Gan X, et al. Ship emission estimation with high spatial-temporal resolution in the Yangtze River estuary using AIS data. *J Clean Prod* 2020; 248: 119297.
45. Yang Y, Liu Y, Li G, et al. Harnessing the power of machine learning for AIS data-driven maritime research: a comprehensive review. *Transp Res E* 2024; 183: 103426.
46. Yang J, Bian X, Qi Y, et al. A spatial-temporal data mining method for the extraction of vessel traffic patterns using AIS data. *Ocean Eng* 2024; 293: 116454.
47. Lin Q, Yin B, Zhang X, et al. Evaluation of ship collision risk in ships' routing waters: a Gini coefficient approach using AIS data. *Phys Stat Mech Appl* 2023; 624: 128936.
48. Sun Y, Lian F and Yang Z. Analysis of the activities of high sea fishing vessels from China, Japan, and Korea via AIS data mining. *Ocean Coast Manag* 2023; 242: 106690.
49. Chen R, Wu X, Liu B, et al. Mapping coastal fishing grounds and assessing the effectiveness of fishery regulation measures with AIS data: a case study of the sea area around the Bohai Strait, China. *Ocean Coast Manag* 2022; 223: 106136.
50. Le Guyader D, Ray C, Gourmelon F, et al. Defining high-resolution dredge fishing grounds with Automatic Identification System (AIS) data. *Aquat Living Resour* 2017; 30: 39.
51. Natale F, Gibin M, Alessandrini A, et al. Mapping fishing effort through AIS data. *PLoS One* 2015; 10: e0130746.
52. Shipxy, <https://www.shipxy.com/> (2024).
53. Vespe M, Gibin M, Alessandrini A, et al. Mapping EU fishing activities using ship tracking data. *J Maps* 2016; 12: 520–525.
54. Taconet M, Kroodsmas D and Fernandes JA. Global atlas of AIS-based fishing activity challenges and opportunities, <https://www.fao.org/3/ca7012en/CA7012EN.pdf> (2019).
55. Alzorri N, Arregi L, Herrmann B, et al. Questioning the effectiveness of technical measures implemented by the Basque bottom otter trawl fleet: implications under the EU landing obligation. *Fish Res* 2016; 175: 116–126.
56. Fiorini M, Capata A and Bloisi DD. AIS data visualization for maritime spatial planning (MSP). *Intl J E-Navig Marit Econ* 2016; 5: 45–60.
57. van Sebille E, Wilcox C, Lebreton L, et al. A global inventory of small floating plastic debris. *Environ Res Lett* 2015; 10: 124006.
58. Lebreton LCM, Greer SD and Borrero JC. Numerical modelling of floating debris in the world's oceans. *Mar Pollut Bull* 2012; 64: 653–661.
59. Lebreton L. The status and fate of oceanic garbage patches. *Nat Rev Earth Environ* 2022; 3: 730–732.
60. García-Monteiro S, Sobrino JA, Julien Y, et al. Surface temperature trends in the Mediterranean Sea from MODIS data during years 2003–2019. *Reg Stud Mar Sci* 2022; 49: 102086.

61. Kessler A, Goris N and Lauvset SK. Observation-based sea surface temperature trends in Atlantic large marine ecosystems. *Prog Oceanogr* 2022; 208: 102902.
62. Merchant CJ, Embury O, Bulgin CE, et al. Satellite-based time-series of sea-surface temperature since 1981 for climate applications. *Sci Data* 2019; 6: 223.
63. Bulgin CE, Merchant CJ and Ferreira D. Tendencies, variability and persistence of sea surface temperature anomalies. *Sci Rep* 2020; 10: 7986.
64. Venegas RM, Acevedo J and Treml EA. Three decades of ocean warming impacts on marine ecosystems: a review and perspective. *Deep Sea Res II* 2023; 212: 105318.
65. Merchant C and Embury O. Adjusting for desert-dust-related biases in a climate data record of sea surface temperature. *Remote Sens* 2020; 12: 2554.
66. Flanders Marine Institute. Maritime boundaries geodatabase, version 11. MarineRegion, <https://www.marineregions.org/>. DOI: 10.14284/382 (2019).
67. ESRI ArcGIS Desktop: Release 10. Redlands, CA: Environmental Systems Research Institute, 2016.
68. Bayazit Y, Bakış R and Koç C. Mapping of stream flow trends in Porsuk Basin using GIS environment. *J Geosci Environ Prot* 2019; 07: 58–66.
69. Bera S. Trend analysis of rainfall in Ganga Basin, India during 1901–2000. *Am J Clim Change* 2017; 06: 116–131.
70. Xue Y, Song D, Chen J, et al. Integrated rockburst hazard estimation methodology based on spatially smoothed seismicity model and Mann-Kendall trend test. *Int J Rock Mech Min Sci* 2023; 163: 105329.
71. Kundzewicz ZW, Graczyk D, Maurer T, et al. Trend detection in river flow series: 1. Annual maximum flow/Détection de tendance dans des séries de débit fluvial: 1. Débit maximum annuel. Trend detection in river flow series. *Hydrol Sci J* 2005; 50: 5.
72. Mann HB. Nonparametric tests against trend. *Econometrica* 1945; 13: 245.
73. Sang YF, Wang Z and Liu C. Comparison of the MK test and EMD method for trend identification in hydrological time series. *J Hydrol* 2014; 510: 293–298.
74. Ngwira GM, Bolaane B and Parida BP. Investigating the trend of road traffic fatalities in Malawi using Mann-Kendall statistic. *Heliyon* 2023; 9: e13700.
75. Tuan NH and Canh TT. Analysis of trends in drought with the non-parametric approach in Vietnam: a case study in Ninh Thuan province. *Am J Clim Change* 2021; 10: 51–84.
76. R: a language and environment for statistical computing, <https://lib.stat.cmu.edu/R/CRAN/doc/manuals/r-devel/fullrefman.pdf> (2024).
77. Lim CH, Ryu J, Choi Y, et al. Understanding global PM<sub>2.5</sub> concentrations and their drivers in recent decades (1998–2016). *Environ Int* 2020; 144: 106011.
78. IUCN. *Made available under agreement on terms of use by the IUCN Joint SSC/WCPA Marine Mammal Protected Areas Task Force and made.* MMPATF. Global dataset of important marine mammal areas (IUCN-IMMA), <https://www.marinemammalhabitat.org/imma-atlas> (2024).
79. BirdLife International. Important Bird and biodiversity Area (IBA) digital boundaries: September 2020 version. 2020.
80. Runfola D, Anderson A, Baier H, et al. Geoboundaries: a global database of political administrative boundaries. *PLoS One* 2020; 15: e0231866.
81. Flanders Marine Institute. Maritime boundaries geodatabase: maritime boundaries and exclusive economic zones (200NM). 2023. DOI: 10.14284/632.
82. FAO. Towards blue transformation. In: *State world fish aquacult*, Vol. 2022. 2022, p.44/266. DOI: 10.4060/cc0461en.
83. Kantai T. *IISD*. The road to a treaty on (marine) plastic pollution, <https://www.iisd.org/articles/policy-analysis/treaty-marine-plastic-pollution> (2022).

84. Corrigan CM, Ardron JA, Comeros-Raynal MT, et al. Developing important marine mammal area criteria: learning from ecologically or biologically significant areas and key biodiversity areas. *Aquat Conserv* 2014; 24: 166–183.
85. Donald PF, Fishpool LDC, Ajagbe A, et al. Important Bird and biodiversity Areas (IBAs): the development and characteristics of a global inventory of key sites for biodiversity. *Bird Conserv Int* 2019; 29: 177–198.
86. Cardoso-Andrade M, Queiroga H, Rangel M, et al. Setting performance indicators for coastal marine protected areas: an expert-based methodology. *Front Mar Sci* 2022; 9: 848039.
87. Oinonen S, Hyytiäinen K, Ahlvik L, et al. Cost-effective marine protection - a pragmatic approach. *PLoS One* 2016; 11: e0147085.
88. Van Diggelen AD, Worden SE, Frimodig AJ, et al. California’s lessons learned and recommendations for effective marine protected area network management. *Mar Policy* 2022; 137: 104928.
89. Leenhardt P, Cazalet B, Salvat B, et al. The rise of large-scale marine protected areas: conservation or geopolitics? *Ocean Coast Manag* 2013; 85: 112–118.
90. Relano V and Pauly D. The ‘Paper Park Index’: evaluating marine protected area effectiveness through a global study of stakeholder perceptions. *Mar Policy* 2023; 151: 105571.
91. Ehler C. Conclusions: benefits, lessons learned, and future challenges of marine spatial planning. *Mar Policy* 2008; 32: 840–843.
92. Douvere F and Ehler CN. New perspectives on sea use management: initial findings from European experience with marine spatial planning. *J Environ Manage* 2009; 90: 77–88.
93. Robards M, Silber G, Adams J, et al. Conservation science and policy applications of the marine vessel Automatic Identification System (AIS)—a review. *Bull Mar Sci* 2016; 92: 75–103.
94. Merchant ND, Pirotta E, Barton TR, et al. Monitoring ship noise to assess the impact of coastal developments on marine mammals. *Mar Pollut Bull* 2014; 78: 85–95.
95. Tasseti AN, Ferrà C and Fabi G. Rating the effectiveness of fishery-regulated areas with AIS data. *Ocean Coast Manag* 2019; 175: 90–97.