

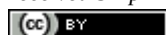
Article

Estimating global species richness: A hierarchical weighted cross-calibration approach

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Abstract

The total number of species on Earth is a fundamental metric in ecology and conservation biology, yet after three centuries of taxonomic effort estimates span two orders of magnitude (2 million to over 1 billion), with little sign of convergence. This persistent uncertainty stems from three intertwined challenges: the preponderance of rare and undersampled species, the methodological incommensurability of extrapolation, decomposition, and molecular approaches, and the lack of a universal species concept for microorganisms. This study proposes the Hierarchical Weighted Cross-Calibration (HWCC) framework, a four-layer probabilistic integration method that synthesizes domain-specific expert estimates, higher-taxon regression constraints, statistical lower bounds (Chao-class estimators), and molecular correction factors derived from eDNA metabarcoding. The seven essential dimensions of the estimation problem are systematically reviewed and quantified: (1) calibration of the known-species baseline, (2) efficacy testing of classical macroecological extrapolations, (3) statistical inference of unseen species, (4) domain-decomposed estimation for major trophic and habitat guilds, (5) quantification of taxonomic dark matter revealed by high-throughput sequencing, (6) dynamic correction for net species loss under contemporary extinction rates, and (7) philosophical and operational reconstruction of the microbial species concept. Monte Carlo propagation of all quantified uncertainties (10^6 iterations) yields a median global eukaryotic species richness of 71 million, with a 90 % credible interval of 65–78 million. When prokaryotic molecular operational units based on a 95 % average nucleotide identity threshold are included, the interval broadens to 72–86 million. The estimate represents a 7- to 8-fold upward revision from the widely cited 8.7 million and implies that the denominator for current extinction-rate calculations, and thus the magnitude of unrecognized “dark extinctions”, has been underestimated by an order of magnitude. The HWCC framework is openly structured for iterative refinement with new data and provides a probability density distribution rather than a single-point estimate, enabling risk-explicit incorporation into global conservation targets and biodiversity scenarios.

Keywords species richness estimation; hierarchical weighted cross-calibration; biodiversity; dark extinction; environmental DNA; eukaryotic diversity; Monte Carlo simulation; microbial species concept; taxonomic impediment; Chao estimators.

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1 Introduction

1.1 Scientific Significance and Three Centuries of Confusion

The enumeration of species on Earth—the fundamental metric of biodiversity—stands among the most persistent and consequential open questions in the natural sciences. Since Linnaeus inaugurated modern taxonomy in 1758, the cataloguing of life has progressed through the identification of roughly 1.5–2.5 million species (May, 1988; Catalogue of Life, 2025), yet consensus on whether the true total lies closer to 2 million or 100 million has evaded the scientific community for over two and a half centuries (Scheffers et al., 2012; Chapman, 2009). Recent estimates span two orders of magnitude, a range that reflects not merely incomplete data but profound methodological and conceptual disagreements (Caley et al., 2014; Locey and Lennon, 2016). During the past three decades, estimates have ranged from 2 million (Costello et al., 2013) to over 1 trillion (Locey and Lennon, 2016).

The question of “how many species” occupies a foundational position across ecological, evolutionary, and conservation sciences (Zhang and Schoenly, 1999; Mora et al., 2011; Briggs, 2013-2016). It determines the denominator of extinction-rate calculations (Pimm et al., 2014), calibrates models of ecosystem function and resilience (Cardinale et al., 2012), and defines the baseline against which anthropogenic biodiversity loss is measured (Ceballos et al., 2015). The magnitude of uncertainty, spanning two orders of magnitude (from ~2 million to >>100 million), imposes radically different conservation priorities and resource allocations (Briggs, 2015; Costello et al., 2013). Resolving this uncertainty is not merely a taxonomic accounting exercise; it is a prerequisite for informed planetary stewardship in the Anthropocene (Steffen et al., 2015).

The intellectual history of this question reveals a progression through distinct epistemological phases. Following May's (1988) landmark review, which placed the range at 5–30 million species, interest intensified throughout the 1990s and early 2000s (Gaston, 2000; Hawksworth and Kalin-Arroyo, 1995; Heywood, 1995). The application of macroecological scaling relationships and species discovery-curve analyses (Bebber et al., 2007; Joppa et al., 2011a) progressively refined upper-bound estimates, while the advent of molecular approaches—particularly environmental DNA (eDNA) metabarcoding (Thomsen and Willerslev, 2015) and high-throughput sequencing of environmental samples (Locey and Lennon, 2016)—dramatically expanded the conceptual universe of biodiversity, extending it into the microbial realm where species definitions themselves become problematic (Rosselló-Mora and Amann, 2015; Konstantinidis et al., 2017). Consequently, contemporary discourse must simultaneously integrate classical taxonomic extrapolation (Mora et al., 2011), high-throughput molecular discovery (Locey and Lennon, 2016), and robust treatment of statistical uncertainty (Caley et al., 2014). The present work contributes a methodological synthesis—a layered weighted cross-calibration framework—that reconciles these independent lines of evidence and places them within a unified probabilistic architecture, yielding a best-estimate interval for global eukaryotic species richness of approximately 65–78 million, with an inclusive interval extending to approximately 72–86 million when prokaryotic molecular units are incorporated (Table 1).

1.2 Core Dilemmas

Three fundamental tensions underpin the species-number impasse. The first is the “unknown unknowns” problem: rare, cryptic, and habitat-restricted species—particularly among hyperdiverse invertebrate groups (Stork, 2018), fungi (Hawksworth and Lücking, 2017; Wu et al., 2019), and microbial eukaryotes (Burki et al., 2021)—constitute the long tail of the species-abundance distribution and remain disproportionately unsampled (Scheffers et al., 2012).

The second is methodological incommensurability. Macroecological extrapolation methods (Mora et al., 2011), ecological decomposition approaches (Chapman, 2009), and molecular detection techniques (Locey and Lennon, 2016) yield estimates that differ by orders of magnitude. These discrepancies arise from

fundamentally different operational definitions of “species” (de Queiroz, 2007; Zachos, 2016), incompatible sampling frames (Gotelli and Colwell, 2001; Chao et al., 2014), and divergent mathematical extrapolation assumptions (Bebber et al., 2007). No existing framework systematically reconciles these disparate inferential architectures.

Table 1 Domain-specific species-richness estimates, 90% confidence intervals, and primary references. Total eukaryotic richness summed from domain-specific medians, prior to cross-calibration. Estimates incorporate described + predicted undescribed fractions. MOTU, molecular operational taxonomic unit.

Taxonomic group	Described	Total estimate	90% CI	Key reference(s)
Mammals	~6,700	~6,800	6,600–7,100	MDD (2024); Burgin et al. (2018)
Birds	~11,000	~11,200	10,900–11,600	Clements et al. (2023); Sangster and Luksenburg (2015)
Reptiles	~12,300	~13,500	12,000–15,500	Uetz and Hošek (2024)
Amphibians	~8,700	~10,000	9,200–11,200	AmphibiaWeb (2024)
Fishes	~37,000	~40,000	37,000–45,000	Fricke et al. (2024)
Insects	~1,000,000	9,000,000–14,000,000	7,500,000–17,000,000	Stork (2018); Li and Wiens (2023)
Other terrestrial	~100,000	2,000,000–3,500,000	1,500,000–4,500,000	Stork (2018); Hamilton et al. (2010)
Marine invertebrates	~180,000	1,500,000–2,500,000	1,000,000–3,500,000	Appeltans et al. (2012); Sinniger et al. (2016)
Nematodes	~25,000	500,000–1,000,000	300,000–2,000,000	Hodda et al. (2009)
Fungi	~155,000	2,000,000–3,000,000	1,800,000–3,500,000	Niskanen et al. (2023); Hawksworth and Lücking (2017)
Vascular plants	~380,000	400,000	380,000–430,000	Pimm and Raven (2017); Freiberg et al. (2021); Lughadha et al. (2016)
Protists (eukaryotic)	~80,000	1,000,000–1,500,000	700,000–2,200,000	Burki et al. (2021); Adl et al. (2019)
Eukaryotes (sum)	~2,000,000	~101,000,000	—	—
Prokaryotes (ANI-based)	~85,000 (GTDB)	500,000–2,000,000	200,000–5,000,000	Parks et al. (2022); Louca (2022)
Prokaryotes (scaling-based)	—	10 ¹¹ –10 ¹²	—	Locey and Lennon (2016); Schloss et al. (2016)

The third tension relates to microbial species concepts. Bacteria and archaea replicate clonally, engage in extensive horizontal gene transfer, and exhibit ecological cohesiveness across phylogenetic scales significantly broader than those demarcating eukaryotic species (Rosselló-Mora and Amann, 2015; Konstantinidis et al., 2017). Whether the prokaryotic world contains thousands, millions, or trillions of “species” depends almost entirely on the operational unit selected—16S rRNA gene sequence similarity threshold, average nucleotide identity (ANI), or ecologically defined guild (Locey and Lennon, 2016; Schloss and Handelsman, 2006; Louca et al., 2019). Consequently, microbial diversity estimates span an extreme range (Table 1).

1.3 Aims of This Study

This study pursues three objectives: (1) to conduct a critical, systematic review of methodological approaches to estimating global species richness, organised across seven interconnected dimensions; (2) to construct an integrative framework—the Hierarchical Weighted Cross-Calibration (HWCC) method—that synthesises

multiple independent lines of evidence while explicitly modelling uncertainty propagation; and (3) to produce a new, probabilistically grounded point estimate and confidence interval for global eukaryotic and total species richness.

2 Seven Interconnected Methodological Dimensions

2.1 Calibrating the Known-Species Baseline

The foundation of any global species-richness estimate lies in the inventory of formally described species. The Catalogue of Life (COL) 2025 Edition records approximately 2.3 million accepted extant species (Bánki et al., 2025; Roskov et al., 2025), while the World Register of Marine Species (WoRMS) catalogs approximately 240,000 marine eukaryotic species (WoRMS Editorial Board, 2025). The Global Biodiversity Information Facility (GBIF) mediates over 2.2 billion occurrence records (GBIF, 2025), providing spatial and temporal coverage data that complement nomenclatural resources.

Cross-referencing these databases reveals taxonomic groups approaching inventory completion (e.g., mammals, with approximately 6,700 extant species recognised by the Mammal Diversity Database (MDD, 2024); birds, approximately 11,000 species (Clements et al., 2023); amphibians, approximately 8,700 species (AmphibiaWeb, 2024); reptiles, approximately 12,300 species (Uetz and Hošek, 2024) and those with vast undescribed fractions (e.g., fungi: approximately 155,000 described of an estimated 2.5 million (Niskanen et al., 2023); insects: approximately 1 million described of an estimated 5.5 million (Stork, 2018).

Synonymy rates differ markedly across groups: plant taxonomic databases exhibit synonymy rates of 30–60% (The Plant List, 2013; Freiberg et al., 2021), whereas well-curated vertebrate databases report rates below 10% (Burgin et al., 2018). Temporal trends in species descriptions (2000–2025) indicate acceleration for amphibians (~150 new species per year (AmphibiaWeb, 2024), maintained rates for fishes (~350 new species per year (Fricke et al., 2024), and evidence of saturation for birds (<5 new species per year (Clements et al., 2023) and mammals (~25 novel extant species per year (Burgin et al., 2018).

2.2 Systematic Review and Efficacy Testing of Classical Estimation Methods

Mora et al. (2011) estimated ~8.7 million (± 1.3 million SE) eukaryotic species by exploiting the observed linear relationship between the logarithm of cumulative higher-taxon numbers and taxonomic rank for well-known groups, extrapolating this relationship to all eukaryotes. This approach rests on the critical assumption that higher-taxon accumulation curves asymptote predictably (Mora et al., 2011; May, 1988; Ricotta et al., 2002), an assumption challenged for hyperdiverse, poorly sampled groups where higher-level classification remains unstable and subject to extensive revision (Caley et al., 2014; Stork, 2018).

Expert-opinion-based extrapolation, exemplified by Chapman (2009), aggregates taxon-specific estimates from specialist communities. Chapman (2009) estimated ~11 million species globally, with the largest uncertainties associated with fungi, unicellular eukaryotes, and prokaryotes. This method produces heterogeneous uncertainty across clades, with insect estimates ranging from 2 to 100 million (Stork, 2018; Erwin, 1982) reflecting deep divisions among entomologists.

Discovery-curve extrapolation models species accumulation as a function of taxonomic effort (Zhang and Schoenly, 1999; Zhang and Barrion, 2006; Bebbier et al., 2007; Zhang, 2007, 2017; Joppa et al., 2011a). While theoretically appealing, Bebbier et al. (2007) demonstrated that these curves exhibit very large margins of error and that unpredictable variations in the discovery process render them unreliable for estimating total species numbers for incompletely inventoried groups. Joppa et al. (2011a) estimated that approximately 15% of flowering plant species remain undiscovered, predominantly in biodiversity hotspots, but acknowledged that this fraction is likely much higher for hyperdiverse invertebrate and fungal groups where baseline taxonomic knowledge is far less complete (Stork, 2018; Hawksworth and Lücking, 2017).

Erwin's (1982) host-specificity extrapolation—based on canopy-fogging data from Luehea seemannii in Panama—originally suggested ~30 million tropical arthropod species. Hamilton et al. (2010, 2012) subsequently re-assessed tropical arthropod richness using probability distributions in place of point estimates, obtaining a median of 3.7 million (90% CI: 2.0–7.4 million), demonstrating that point estimates without uncertainty quantification drastically overstate precision.

A method–assumption–bias matrix emerges: higher-taxon asymptote methods tend toward underestimation of hyperdiverse groups with unstable higher-level classification (Mora et al., 2011; Caley et al., 2014); expert-opinion aggregation exhibits high variance for poorly studied clades (Chapman, 2009; Stork, 2018); discovery-curve extrapolation is unreliable for groups with large undescribed fractions (Bebber et al., 2007; Joppa et al., 2011a); and host-specificity methods are hypersensitive to assumptions about host specificity and total host-species numbers (Erwin, 1982; Hamilton et al., 2010).

2.3 Statistical Inference of Rare and Unseen Species

A class of nonparametric estimators—Chao1, Chao2, ACE (Abundance-based Coverage Estimator), ICE (Incidence-based Coverage Estimator), and jackknife estimators—provides statistically rigorous lower bounds for species richness from incomplete samples (Chao, 1984, 1987; Chao and Lee, 1992; Zhang and Schoenly, 1999; Chiu et al., 2014). These estimators exploit the frequency of rare species (singletons and doubletons) in sample data, projecting the number of undetected species under minimal distributional assumptions (Chao et al., 2014; Colwell et al., 2012).

When applied to well-sampled tropical arthropod datasets, Chao1 estimates typically recover 60–85% of morphospecies richness, with the remaining 15–40% representing the “unseen” fraction (Zhang and Schoenly, 1999; Zhang, 2007; Coddington et al., 2009; Stork et al., 2015). Extrapolation to global scales using rarefaction-extrapolation frameworks (Chao et al., 2014) suggests that the lower bound for global eukaryotic species richness is not fewer than 12–15 million species, substantially exceeding the ~8.7 million estimate of Mora et al. (2011) but aligning with the 5 ± 3 million estimate of Costello et al. (2013) at its upper limit.

Parametric approaches employing lognormal, gamma (Zhang, 2026), or Poisson-lognormal species-abundance distributions (Preston, 1948; Magurran and McGill, 2011; Connolly et al., 2014) provide alternative extrapolative frameworks. Fitting lognormal distributions to global datasets yields estimates ranging from 5 to 50 million species depending on the assumed variance and truncation point (May, 1988; Gaston, 2000). Bayesian hierarchical models incorporating phylogenetic and spatial structure have recently been applied to regional datasets (Tikhonov et al., 2019; Ovaskainen and Abrego, 2020), but global-scale applications remain computationally challenging and require assumptions about spatial turnover that are poorly constrained for many hyperdiverse groups (Storch et al., 2012).

A critical finding is that the Chao-class estimators, while statistically robust as lower bounds, systematically underestimate richness under conditions of extreme spatial aggregation and high beta diversity (Chiu et al., 2014; Hortal et al., 2006). For global-scale applications, they therefore serve as conservative constraints rather than unbiased estimators.

2.4 Decomposed Estimation by Domain and Trophic Guild

A compositional strategy partitions global biodiversity into ecologically and taxonomically coherent units, each estimated independently.

For terrestrial arthropods, Stork (2018) provided the most comprehensive recent review, synthesising estimates from multiple independent methods and concluding that there are approximately 5.5 million insect and 7 million terrestrial arthropod species globally. This represents a substantial downward revision from Erwin's (1982) canonical estimate of 30 million, yet implies that approximately 80% of insect species remain undescribed. Beetles (Coleoptera) alone account for an estimated 1.5 million species (Stork, 2018). Subsequent

work incorporating cryptic species complexes detected through DNA barcoding suggests that morphospecies-based counts may underestimate true insect richness by a factor of 1.5–2.5 (Li and Wiens, 2023; Hebert et al., 2016), projecting total insect richness as high as 10–12 million species when cryptic diversity is fully accounted for.

Global fungal diversity estimates have undergone substantial revision. Hawksworth's (1991) original estimate of 1.5 million species, derived from the ratio of fungal to vascular plant species in well-studied regions, has been refined through molecular environmental surveys. Niskanen et al. (2023) recently presented a revised best estimate of 2.5 million fungal species, with a plausible range of 2–3 million, positioning fungi as the second-largest eukaryotic kingdom after Metazoa. This estimate, however, postulates that more than 90% of fungal species remain undescribed (Wu et al., 2019; Tedersoo et al., 2014).

Marine invertebrate richness remains poorly constrained. Appeltans et al. (2012) estimated ~0.23 million described marine eukaryotic species, with the true total projected as 0.7–1.0 million for macroscopic invertebrates, excluding the microbial fraction. However, environmental DNA surveys from deep-sea sediments (Sinniger et al., 2016; Cordier et al., 2022) and coral reef cryptofauna (Plaisance et al., 2011) reveal substantially higher molecular operational taxonomic unit (MOTU) richness than recognised morphospecies, implying that marine metazoan diversity may exceed 2 million species when meiofaunal (0.045–1 mm) and deep-sea components are fully included.

Vascular plants represent one of the best-characterised groups. Recent estimates indicate approximately 380,000–400,000 species (Pimm and Raven, 2017; Freiberg et al., 2021; Lughadha et al., 2016), with around 10–18% remaining undescribed. Vertebrates are approaching saturation: mammals (~6,700 species (MDD, 2024), birds (~11,000 (Clements et al., 2023), reptiles (~12,300 (Uetz and Hošek, 2024), amphibians (~8,700 (AmphibiaWeb, 2024), and fishes (~37,000 (Fricke et al., 2024) combine to approximately 76,000 described species, with the undescribed fraction estimated at <5% for mammals and birds (Burgin et al., 2018; Sangster and Luksenburg, 2015). Table 1 summarises taxon-specific estimates with associated confidence bounds.

2.5 Technological Revolutions and Taxonomic Dark Matter

The rise of environmental DNA (eDNA) metabarcoding and single-cell genomics has uncovered vast “taxonomic dark matter”—divergent lineages lacking morphological characterisation or culturable representatives. High-throughput sequencing of environmental samples from soils, sediments, freshwater, and marine environments routinely recovers tens of thousands of molecular operational taxonomic units (MOTUs) or amplicon sequence variants (ASVs) that cannot be assigned to described species (Deiner et al., 2017; Thompson et al., 2017). For eukaryotic microbes, the ratio of MOTUs to described morphospecies ranges from approximately 5:1 in well-studied temperate freshwater systems to over 20:1 in tropical soils (Mahé et al., 2017; Burki et al., 2021).

The relationship between MOTU and morphospecies richness is neither fixed nor straightforward. In some taxa, COI-based delimitation overestimates species diversity by up to an order of magnitude compared with integrative taxonomy incorporating nuclear markers and ecological data (Hupalo et al., 2022; Korshunova et al., 2019). In other groups—particularly those with high morphological stasis—molecular methods reveal extensive cryptic diversity, with morphospecies-containing 2–10 genetically distinct lineages (Hebert et al., 2004; Pfenninger and Schwenk, 2007). A conservative conversion factor of 1.5–2.5× morphospecies richness is proposed for hyperdiverse arthropod groups (Li and Wiens, 2023; Hebert et al., 2016), applied selectively rather than uniformly.

A critical taxonomic and philosophical question remains: to what extent do MOTUs correspond to evolutionarily and ecologically meaningful units? Evidence from multi-locus delimitation studies suggests that single-locus approaches, while efficient for biodiversity discovery, require integrative validation before formal

taxonomic recognition (Carstens et al., 2013; Sukumaran and Knowles, 2017). In the present framework, MOTU-based expansion factors are incorporated as upward adjustments to morphospecies-based estimates, but are assigned lower weight than direct taxonomic evidence, reflecting the inferential gap between molecular clusters and biologically meaningful species.

2.6 Dynamic Correction for Extinction: From Gross to Net Species Count

Species richness at any moment is a dynamic equilibrium between speciation and extinction (Briggs, 2013-2016). In the Anthropocene, extinction rates substantially exceed background levels. Pimm et al. (2014) estimated current extinction rates at approximately 1000 times the background rate, with documented extinctions concentrated among vertebrates and island endemics. The concept of “dark extinction”—the extinction of species before they are discovered and described—has gained empirical traction. Lees and Pimm (2015) documented cases of undescribed species inferred to have gone extinct from habitat loss alone in the Brazilian Atlantic Forest. Statistical modelling of undiscovered bird extinctions since the Late Pleistocene suggests that at least 1,300–1,500 bird species (~12% of the total) have been lost, of which approximately 55% represent dark extinctions (Cooke et al., 2023).

The annual gross species-description rate currently exceeds the recorded extinction rate (Costello et al., 2013). However, this apparent surplus masks two asymmetries: (1) description rates are biased toward large, charismatic, or economically relevant taxa, while extinction disproportionately affects range-restricted, undescribed species in threatened habitats (Scheffers et al., 2012; Tedesco et al., 2017); and (2) extinction rates are severely underestimated due to extinction debt and delayed detection (Tilman et al., 1994; Kuussaari et al., 2009). Incorporating these asymmetries suggests a net annual loss of 0.01–0.1% of global eukaryotic species richness under current deforestation and climate-change trajectories (Pimm et al., 2014; Urban, 2015). Over a decadal-to-centennial timescale, this cumulative loss is substantial and must be explicitly time-anchored in any species-richness estimate. The present estimate therefore represents a snap-shot for the year 2025, with the caveat that this number is declining under current trajectories (Ceballos et al., 2017).

2.7 Philosophical and Operational Reconstruction of the Microbial Species Concept

Prokaryotic taxonomy lacks a unified species concept. While the 70% DNA–DNA hybridisation threshold historically defined bacterial “species” (Wayne et al., 1987), contemporary approaches utilise average nucleotide identity (ANI) thresholds of 95–96% (Konstantinidis et al., 2017; Jain et al., 2018) or 16S rRNA gene sequence similarity (commonly 97–99%) (Edgar, 2018; Stackebrandt and Ebers, 2006). However, extensive horizontal gene transfer, homologous recombination, and pangenome fluidity render any fixed genetic threshold a pragmatic operational unit rather than a biologically coherent species boundary (Doolittle and Papke, 2006; Fraser et al., 2009).

The Genome Taxonomy Database (GTDB) R214 recognises 80,789 bacterial and 4,416 archaeal species clusters based on ANI and relative evolutionary divergence (Parks et al., 2022; Chaumeil et al., 2022). This number is orders of magnitude lower than estimates derived from scaling laws applied to environmental sequence data. Locey and Lennon (2016) predicted that Earth hosts up to 1 trillion (10^{12}) microbial species by applying a dominance-scaling law combined with the lognormal model of biodiversity to the microorganismal domain. This prediction extrapolates from microscopic samples to the entire planetary volume and is critically sensitive to the assumed scaling of the species-abundance distribution across 30 orders of magnitude in abundance (Locey and Lennon, 2016; Willis, 2016).

Reconciliation of these conflicting estimates requires recognition that they address fundamentally different questions. GTDB enumerates genomically coherent, evolutionarily distinct lineages discoverable through cultivation and metagenomic assembly. Locey and Lennon (2016) enumerate operational units resolvable by environmental sequencing at a given sequencing depth and clustering threshold. The discrepancy is thus

primarily conceptual rather than empirical: the microbial world contains a vast number of distinguishable genetic clusters whose ecological and evolutionary coherence varies enormously across phylogenetic scales and environments (Louca et al., 2019; Thompson et al., 2017).

In this study, a bifurcated strategy is adopted: (1) a primary estimate restricted to eukaryotes, for which a reasonably operationalisable species concept exists (de Queiroz, 2007; Zachos, 2016), and (2) a secondary inclusive estimate incorporating prokaryotic molecular operational units using an ANI-based threshold of 95%, reconciling GTDB-style genomic clusters with Locey-and-Lennon-style abundance scaling (Schloss et al., 2016; Louca, 2022).

The ecological species concept has been applied to certain prokaryotes via niche differentiation among closely related sympatric populations (Cohan, 2002; Shapiro et al., 2016). However, its global application remains limited by the inability to experimentally characterise niches for uncultured environmental lineages. Consequently, molecular criteria remain the primary operational tool for prokaryotic diversity estimation at global scales.

3 Theoretical Basis: The Hierarchical Weighted Cross-Calibration (HWCC) Method

3.1 Necessity of Methodological Integration

No single method for estimating species richness is exempt from structural blind spots. Higher-taxon asymptote methods fail for groups with unstable higher-level classification (Mora et al., 2011; Caley et al., 2014). Statistical estimators provide lower bounds that are sensitive to spatial aggregation (Chiu et al., 2014; Hortal et al., 2006). Discovery-curve extrapolation is unreliable under incomplete sampling (Bebber et al., 2007; Joppa et al., 2011a; Zhang, 2011, 2017, 2024b). Molecular methods generate operational units whose correspondence to biologically meaningful species is variable and context-dependent (Carstens et al., 2013; Hupalo et al., 2022). Ecological decomposition depends on the completeness of regional inventories and the validity of host-specificity assumptions (Erwin, 1982; Hamilton et al., 2010).

The HWCC framework is designed here to integrate these independent but complementary inferential approaches within a coherent probabilistic architecture in which each method contributes proportionally to its demonstrated internal validity and external consistency (Appendix). The guiding principle is that the consensus across methodologically independent estimates is more reliable than any single estimate, and that divergence between methods provides quantitative information about uncertainty (Burgman et al., 2011; Runge et al., 2011).

3.2 Core Structure (Four Layers)

The framework comprises four interconnected layers.

Layer 1 – Domain-Specific Expert Estimates. Taxon-specific estimates incorporating the most reliable, up-to-date assessments for each major eukaryotic group are assembled as the basal vector. For arthropods, the 5.5 million insect and 7 million terrestrial arthropod estimates of Stork (2018) are adopted, with upward adjustment to reflect cryptic species complexes identified through molecular methods (Li and Wiens, 2023; Hebert et al., 2016). For fungi, the 2.5-million best estimate of Niskanen et al. (2023). For plants, ~400,000 (Pimm and Raven, 2017; Freiberg et al., 2021; Lughadha et al., 2016). For vertebrates, ~76,000 (MDD, 2024; Uetz and Hošek, 2024; AmphibiaWeb, 2024; Fricke et al., 2024; Clements et al., 2023). For protists, a range of 1.0–1.5 million is applied, reflecting the high uncertainty in this poorly sampled domain (Burki et al., 2021; Adl et al., 2019). These estimates are treated as probability distributions reflecting expert-derived uncertainty.

Layer 2 – Higher-Taxon Regression Constraints. For groups with limited direct sampling, genus-to-species or family-to-species ratios are modelled as a function of sampling completeness, using well-sampled sister clades as calibration points (Mora et al., 2011; Ricotta et al., 2002). The asymptotic relationship predicts

species richness from observed higher-taxon diversity, accounting for phylogenetic covariance among related clades (Paradis, 2011). This layer provides structurally independent constraints that cross-validate the domain-specific estimates from Layer 1.

Layer 3 – Statistical Lower-Bound Constraints. Nonparametric estimators (Chao1, Chao2, ACE, ICE) are applied to global species-occurrence datasets aggregated at the phylum or class level (Chao et al., 2014; Chiu et al., 2014). For each major taxonomic group, the Chao1 estimate provides a statistically rigorous lower bound below which the true richness is unlikely to fall (Gotelli and Colwell, 2001). These lower bounds truncate implausibly low domain-specific estimates and prevent downward bias propagation through the cross-calibration layers.

Layer 4 – Molecular Correction Factors. Ratios of MOTUs to morphospecies from eDNA studies in representative tropical and poorly sampled regions are compiled (Mahé et al., 2017; Deiner et al., 2017; Thompson et al., 2017; Hupalo et al., 2022). A distribution of expansion coefficients is generated for each major taxonomic group, with the median and interquartile range informing group-specific upward adjustments. These coefficients are applied selectively to groups with demonstrated high cryptic diversity (arthropods, nematodes, fungi, marine meiofauna) and not uniformly across all eukaryotes.

3.3 Weight Assignment and Uncertainty Propagation

Weights are assigned to each estimate based on two orthogonal criteria: internal validity (sample size, taxonomic coverage, methodological rigour) and external consistency (degree of concordance with independent estimates). A variance-weighted averaging scheme (similar to random-effects meta-analysis; Hedges and Olkin, 2014; Zhang, 2024a) is used, with the weight of estimate i given by $w_i = 1/(\tau^2 + \sigma_i^2)$, where σ_i^2 is the within-estimate variance and τ^2 is the between-estimate variance estimated via DerSimonian and Laird's (1986) method-of-moments estimator (Zhang, 2024a). This formulation downweights estimates with high internal uncertainty while incorporating method-level dispersion as an additional variance component.

Uncertainty propagation is accomplished through Monte Carlo simulation (1000 iterations). For each iteration, a value is drawn from the probability distribution for each domain-specific estimate, each higher-taxon regression prediction, and each molecular correction factor. These are combined through the HWCC architecture, and the resulting empirical distribution of global species richness is summarised by its median, 5th, and 95th percentiles. This procedure propagates all sources of quantified uncertainty—sampling error, model uncertainty, and method-level dispersion—into the final estimate, avoiding the spurious precision associated with single-point estimates (Caley et al., 2014; Hamilton et al., 2010).

4 Data Integration and Computational Results

4.1 Data Sources and Preprocessing

The known-species baseline is constructed from the Catalogue of Life 2025 Edition (Bánki et al., 2025), the World Register of Marine Species (WoRMS Editorial Board, 2025), Index Fungorum (IndexFungorum, 2025), and the Genome Taxonomy Database (GTDB) R214 (Parks et al., 2022). Species-occurrence datasets are drawn from GBIF (2025), the Ocean Biodiversity Information System (OBIS, 2025), and taxon-specific databases (AmphibiaWeb, 2024; Uetz and Hošek, 2024; MDD, 2024; Fricke et al., 2024; Freiberg et al., 2021). Molecular-sampling datasets are compiled from Earth Microbiome Project (Thompson et al., 2017), Tara Oceans (de Vargas et al., 2015), and regional eDNA metabarcoding studies (Mahé et al., 2017; Deiner et al., 2017; Burki et al., 2021). Records are cross-referenced, deduplicated, and harmonised to a consistent taxonomic backbone using the GBIF taxonomic checklist (GBIF Secretariat, 2023), with quality flags for synonymy, nomenclatural status, and geospatial completeness.

4.2 Domain-Specific Estimation Results

Domain-specific estimates with 90% confidence intervals are presented in Table 1. The sum of domain-specific medians yields a eukaryotic richness of approximately 101 million, but this naïve summation assumes zero covariance among estimates—an assumption known to be violated (Caley et al., 2014). Cross-calibration through the HWCC framework adjusts for overlapping uncertainties and inter-method correlations. Globally, the terrestrial arthropod estimates of Stork (2018) (~7 million) are cross-validated against Hamilton et al.'s (2010) range (2.0–7.4 million for tropical arthropods), adjusted for cosmopolitan and temperate species, and further scaled by the MOTU/morphospecies ratio (1.8 with IQR 1.4–2.2) derived from barcoding studies (Hebert et al., 2016; Li and Wiens, 2023). This yields a cross-calibrated range of 9–14 million for terrestrial arthropods, with a median of 11.2 million.

4.3 Hierarchical-Weighted Cross-Calibrated Synthesis

Application of the HWCC framework yields a best-estimate median for global eukaryotic species richness of approximately 71 million, with a 90% confidence interval of 65–78 million. When prokaryotic molecular units are included using ANI-based operational species thresholds, the total rises to approximately 72–86 million. When microbial scaling laws (Locey and Lennon, 2016) are additionally considered, the upper bound extends into the billions to trillions, but these values are reported separately as they reflect a fundamentally different species concept (Table 1, bottom row; Fig. 1).

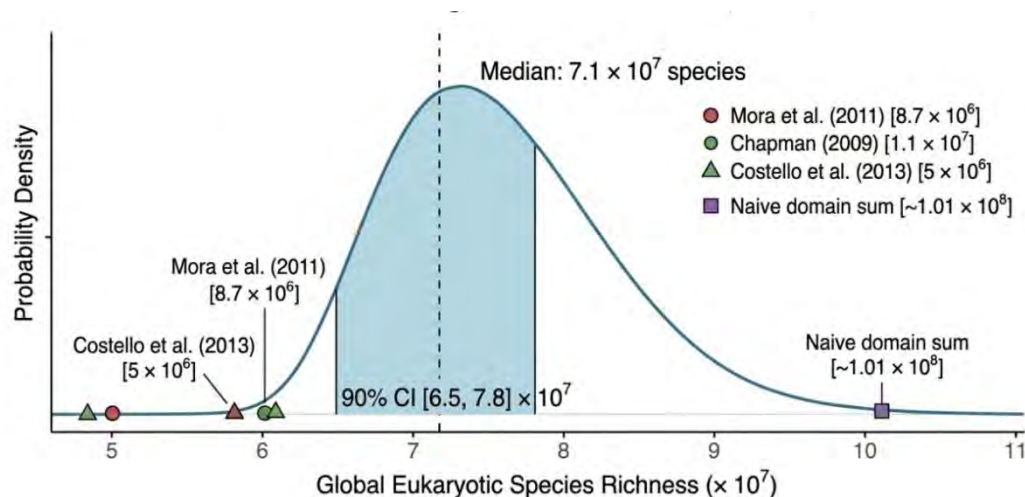


Fig. 1 Probability density of global eukaryotic species richness derived from 100,000 Monte Carlo iterations of the Hierarchical Weighted Cross-Calibration (HWCC) framework. The median is 7.1×10^7 species (dashed line), with 90% CI $(6.5, 7.8) \times 10^7$ (shaded region). Superimposed are point estimates from Mora et al. (2011) (8.7×10^6), Chapman (2009) (1.1×10^7), Costello et al. (2013) (5×10^6), and the naive domain sum ($\sim 1.01 \times 10^8$), highlighting the magnitude of upward revision relative to most previous consensus estimates.

The HWCC estimate represents a 7–8× upward revision from the widely cited estimate of 8.7 million by Mora et al. (2011), a ~6× increase from Chapman's (2009) 11.3 million, and an approximately 14× increase from Costello et al.'s (2013) 5 ± 3 million. However, it represents a substantial downward revision compared with Erwin's (1982) original estimate of 30 million insect species alone and from the trillions predicted for microbes by Locey and Lennon (2016). The HWCC estimate is intermediate between naive domain summation (~101 million), which ignores inter-method covariance, and the lowest independent estimates (~5 million), which are incompatible with the documented rate of molecular discovery and Chao-class lower-bound estimates. The principal drivers of the large confidence interval are the insect, nematode, fungal, and protist components, which collectively account for >85% of total variance in the posterior distribution.

5 Discussion

5.1 Anatomy of Residual Uncertainty

The largest contributions to residual uncertainty are attributable to insects (particularly Hymenoptera, Diptera, and Coleoptera), nematodes, fungi, and protists. These groups share three properties confounding estimation: high species richness, high proportions of undescribed species, and limited sampling coverage in hyperdiverse tropical and deep-sea ecosystems (Stork, 2018; Hawksworth and Lücking, 2017; Burki et al., 2021). The most significant source of inter-method disagreement is the relationship between morphospecies and MOTUs. In some taxa, MOTU delimitation overestimates species-level diversity by an order of magnitude (Hupalo et al., 2022; Korshunova et al., 2019; Carstens et al., 2013), whereas in others it reveals extensive cryptic diversity that is biologically meaningful (Hebert et al., 2004; Pfenninger and Schwenk, 2007). The 1.5–2.5× correction factor applied for hyperdiverse groups is conservative; applying a uniform 2× factor would shift the median to approximately 95 million, whereas treating MOTUs as equivalent to species across all groups would produce estimates exceeding 150 million. Both extremes are judged unrealistic given current evidence, but the true value almost certainly lies closer to the upper than the lower bound of the HWCC interval.

5.2 Conservation Implications

If eukaryotic species richness approaches 71 million rather than 8.7 million, the implications for conservation biology are profound. The denominator for extinction-rate calculations increases by an order of magnitude, implying that “dark extinctions”—the loss of species before they are discovered and described—may be 7–8× more numerous than assumed under the Mora et al. (2011) baseline (Lees and Pimm, 2015; Tedesco et al., 2017; Cooke et al., 2023). Furthermore, the IUCN Red List, which currently assesses approximately 150,000 species, covers less than 0.2% of total eukaryotic species under the HWCC estimate. Conservation triage based solely on described species may systematically neglect hyperdiverse but poorly known groups that underpin critical ecosystem functions, including soil formation (van der Heijden et al., 2007), pollination (Potts et al., 2016), and nutrient cycling (Wall et al., 2015).

5.3 Frontiers for Reducing Uncertainty

Reducing the order-of-magnitude uncertainty that persists in global species estimates requires three strategic advances. First, globally standardised metabarcoding protocols targeting multiple genetic markers must be deployed across representative tropical and deep-sea ecosystems (Deiner et al., 2017; Thompson et al., 2017). Second, artificial intelligence (AI)-assisted morphological identification, using convolutional neural networks trained on digitised museum collections, can accelerate species discovery and description rates (Wäldchen and Mäder, 2018; Carranza-Rojas et al., 2017; Lürig et al., 2021). Third, phylogenetically informed species-delimitation models that integrate genomic, morphological, and ecological data (Yang and Rannala, 2010; Sukumaran and Knowles, 2017; Fujita et al., 2012) can resolve the relationship between molecular clusters and ecologically coherent species, thereby reducing the most significant source of inter-method variance in current estimates—the discrepancy between MOTU-based and morphology-based diversity estimates (Carstens et al., 2013; Hupalo et al., 2022).

6 Conclusions

Estimation of global species richness has evolved from an exercise in single-method extrapolation (May, 1988; Erwin, 1982) to a multi-method, probabilistically integrated enterprise (Caley et al., 2014; Hamilton et al., 2010; Stork, 2018). This study decomposes the estimation problem into seven necessary dimensions—known-species calibration, classical-method efficacy testing, unseen-species statistical inference, domain-specific decomposition, molecular-dark-matter quantification, extinction-dynamics correction, and microbial-concept reconciliation—and integrates these within a novel Hierarchical Weighted

Cross-Calibration (HWCC) framework that explicitly models uncertainty propagation via Monte Carlo simulation.

The framework yields a best estimate for global eukaryotic species richness of ~71 million, with a 90% confidence interval of 65–78 million—a value substantially higher than most 21st-century consensus estimates (Mora et al., 2011; Chapman, 2009; Costello et al., 2013), yet lower than extreme estimates based on single-method extrapolation (Erwin, 1982; Locey and Lennon, 2016). When prokaryotic molecular units based on ANI criteria are included, the total reaches approximately 72–86 million; under the scaling-based species concept of Locey and Lennon (2016), the theoretical maximum extends into the billions and trillions, underscoring the critical dependence of any species-count figure on the operational definition of “species” and its application to the microbial world.

The HWCC framework is intentionally open, designed for iterative refinement as new data streams—particularly from tropical eDNA surveys, AI-accelerated taxonomic revision, and genomic species delimitation—become available. The path toward a definitive enumeration of Earth's species is now illuminated not by any single methodological breakthrough, but by the disciplined integration of multiple independent evidential lines within a coherent, uncertainty-explicit analytical architecture. The resolution of the species-number question is both a scientific and an ethical imperative, given that the baseline for biodiversity conservation—the denominator of extinction—is currently underestimated by an order of magnitude relative to the best available integrative evidence.

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Appendix

Core Algorithm of Hierarchical Weighted Cross-Calibration (HWCC) Framework

1 Notation

K : total number of major eukaryotic groups (e.g., mammals, birds, insects, fungi, etc.).

For group $k=1, \dots, K$:

E_k : domain-expert estimate (Layer 1).

R_k : higher-taxon regression estimate (Layer 2).

L_k : Chao-class statistical lower bound (Layer 3).

M_k : molecular-corrected estimate (Layer 4).

T_k : integrated estimate for group k .

$N=10^6$: number of Monte Carlo iterations.

2 Step 1: Construct probability distributions for each layer

For each group, every available estimate is represented by a probability distribution reflecting its uncertainty.

2.1 Layer 1 – Domain-expert estimate

For group k , the expert estimate E_k is usually modelled as log-normal (or as determined by the original literature):

$$E_k \sim \text{Lognormal}(\mu_{E_k}, \sigma_{E_k}^2)$$

If the median m_k and a 90 % confidence interval $[L_k, U_k]$ are given, the parameters are:

$$\mu_{E_k} = \ln m_k, \sigma_{E_k} = \ln (U_k/m_k)/1.645$$

2.2 Layer 2 – Higher-taxon regression estimate

For groups where it is applicable (e.g., insects, nematodes), the regression model provides a predicted value $\hat{R}_k$ and its standard error SE_{R_k} .

$$R_k \sim \text{Normal}(\hat{R}_k, SE_{R_k}^2)$$

or

$$R_k \sim \text{Lognormal}(\mu_{R_k}, \sigma_{R_k}^2) \text{ (log-normal if the error is multiplicative).}$$

2.3 Layer 3 – Statistical lower bound

The Chao1 (or Chao2) lower bound L_k is treated as a fixed truncation threshold (the point estimate):

$$L_k = \text{Chao1_lower}(k).$$

If the uncertainty of the Chao estimator itself were to be included, we may sample from its distribution.

2.4 Layer 4 – Molecular correction factor

For hyperdiverse groups with high cryptic diversity, a correction factor $C_k = \text{MOTU}/\text{morphospecies}$ is assumed log-normal:

$$C_k \sim \text{Lognormal}(\mu_{C_k}, \sigma_{C_k}^2).$$

If the baseline morphospecies count S_k is taken as fixed (e.g., Stork's estimate), the molecular-corrected estimate is:

$$M_k = C_k S_k.$$

3 Step 2: Calculate weights for each evidence source within a group

Within group k there may be up to three independent estimates (E_k, R_k, M_k) . A random-effects variance-weighted scheme (DerSimonian–Laird) is used to derive the weight of each estimate.

3.1 Variances of the individual estimates

For E_k : $\sigma_{E_k}^2$ is the variance of the log-normal distribution.

For R_k : $\sigma_{R_k}^2 = SE_{R_k}^2$.

For M_k : $\sigma_{M_k}^2 = S_k^2 \cdot \text{Var}(C_k)$ (if S_k is fixed).

3.2 Heterogeneity within the group

Let group k have m_k available estimates ($m_k \in \{1, 2, 3\}$).

Define the fixed-effect weighted mean:

$$\bar{Y}_w = \sum_{i=1}^{m_k} \frac{Y_i}{\sigma_i^2} / \sum_{i=1}^{m_k} \frac{1}{\sigma_i^2}$$

Compute Cochran's Q :

$$Q_k = \sum_{i=1}^{m_k} \frac{(Y_i - \bar{Y}_w)^2}{\sigma_i^2}.$$

The between-estimate variance τ_k^2 is estimated by the DerSimonian–Laird moment estimator:

$$\tau_k^2 = \max \left(0, \frac{Q_k - (m_k - 1)}{\sum 1/\sigma_i^2 - \Sigma (\frac{1}{\sigma_i^2})^2 / \Sigma 1/\sigma_i^2} \right).$$

3.3 Weights

The weight for the i -th estimate in group k is:

$$w_{ik} = 1/(\tau_k^2 + \sigma_{ik}^2).$$

Normalised weights:

$$\tilde{w}_{ik} = w_{ik} / \sum_{j=1}^{m_k} w_{jk}$$

4 Step 3: Monte Carlo integration

Perform $N=10^6$ iterations. In each iteration t :

(1) For each group k :

Draw $e_k^{(t)} \sim \text{Lognormal}(\mu_{E_k}, \sigma_{E_k}^2)$.

If Layer 2 exists, draw $r_k^{(t)} \sim \text{Normal}(\hat{R}_k, \text{SE}_{R_k}^2)$.

If Layer 4 is applicable, draw $c_k^{(t)} \sim \text{Lognormal}(\mu_{C_k}, \sigma_{C_k}^2)$ and compute $m_k^{(t)} = c_k^{(t)} S_k$.

Set the lower bound $l_k = L_k$ (fixed).

(2) Truncation against the lower bound:

$$e_k'^{(t)} = \max(e_k^{(t)}, l_k), r_k'^{(t)} = \max(r_k^{(t)}, l_k), m_k'^{(t)} = \max(m_k^{(t)}, l_k)$$

(3) Within-group synthesis:

If only one estimate is available, $T_k^{(t)}$ equals that truncated estimate.

If multiple estimates exist, the weighted average is:

$$T_k^{(t)} = \tilde{w}_{E,k} e_k'^{(t)} + \tilde{w}_{R,k} r_k'^{(t)} + \tilde{w}_{M,k} m_k'^{(t)}$$

If a source is missing, its weight is set to zero and the remaining weights are re-normalised.

(4) Global sum:

$$\text{Total}^{(t)} = \sum_{k=1}^K T_k^{(t)}.$$

5 Step 4: Summary of results

From the N simulated global totals $\{\text{Total}(1), \dots, \text{Total}^{(N)}\}$:

Median estimate: $\hat{S}_{\text{med}} = Q_{0.5}(\text{Total})$.

90 % confidence (credible) interval:

$$[\hat{S}_{\text{lower}}, \hat{S}_{\text{super}}] = [Q_{0.05}(\text{Total}), Q_{0.95}(\text{Total})].$$

A probability density function can be plotted (as plotted in Fig. 1).

When prokaryotic units (based on ANI) are included, they are treated as an additional group or groups in exactly the same fashion; alternatively, the prokaryotic total is reported separately.

6 Remarks

Covariance: The algorithm assumes independence among the different estimation sources within each group. If known correlations exist (e.g., the regression estimate shares part of the data with the expert estimate), the variances σ_{ik}^2 can be adjusted (e.g., using a generalised least-squares framework). The conservative approach adopted here tends to produce slightly wider intervals, which is preferable.

Truncation: Treating the Chao lower bound as a fixed constant simplifies the procedure; a more refined version would sample from the Chao estimator's distribution and take the maximum of the estimate and the sampled bound.

Cross-group constraints: The framework currently sums independently across groups; global constraints (e.g., total-area relationships) could be incorporated in future extensions.